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The relative importance of phytoplankton aggregates and zooplankton fecal pellets to carbon export: insights from free-drifting sediment trap deployments in naturally iron-fertilised waters near the Kerguelen Plateau

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Abstract. The first KErguelen Ocean and Plateau compared Study (KEOPS1), conducted in the naturally iron-fertilised Kerguelen bloom, demonstrated that fecal material was the main pathway for exporting carbon to the deep ocean during summer (January–February 2005), suggesting a limited role of direct export via phytodetrital aggregates. The KEOPS2 project reinvestigated this issue during the spring bloom initiation (October–November 2011), when zooplankton communities may exert limited grazing pressure, and further explored the link between carbon flux, export efficiency and dominant sinking particles depending upon surface plankton community structure. Sinking particles were collected in polyacrylamide gel-filled and standard free-drifting sediment traps (PPS3/3), deployed at six stations between 100 and 400 m, to examine flux composition, particle origin and their size distributions. Results revealed an important contribution of phytodetrital aggregates (49 ± 10 and 45 ± 22 % of the total number and volume of particles respectively, all stations and depths averaged). This high contribution dropped when converted to carbon content (30 ± 16 % of total carbon, all stations and depths averaged), with cylindrical fecal pellets then representing the dominant fraction (56 ± 19 %).

At 100 and 200 m depth, iron- and biomass-enriched sites exhibited the highest carbon fluxes (maxima of 180 and 84 ± 27 mg C m−2 d−1, based on gel and PPS3/3 trap collection respectively), especially where large fecal pellets dominated.
over phytodetrital aggregates. Below these depths, carbon fluxes decreased (48 ± 21 % decrease on average between 200 and 400 m), and mixed aggregates composed of phytodetritus and fecal matter dominated, suggesting an important role played by physical aggregation in deep carbon export.

Export efficiencies determined from gels, PPS3/3 traps and $^{234}\text{Th}$ disequilibria (200 m carbon flux/net primary productivity) were negatively correlated to net primary productivity with observed decreases from ~0.2 at low-iron sites to ~0.02 at high-iron sites. Varying phytoplankton communities and grazing pressure appear to explain this negative relationship. Our work emphasises the need to consider detailed plankton communities to accurately identify the controls on carbon export efficiency, which appear to include small spatio-temporal variations in ecosystem structure.

1 Introduction

Physical and biological processes occurring in the surface ocean generate a vast diversity of particles. These particles represent potential vehicles to export organic carbon to the deep ocean, where a small fraction can eventually be sequestered in the sediments. This process, known as the “biological carbon pump” (BCP), influences the level of atmospheric carbon dioxide and thus the global climate system (Volk and Hoffert, 1985; Lam et al., 2011).

Primary production in the euphotic layer builds a stock of phytoplankton cells. If their concentration and stickiness are high enough (Jackson, 1990), these cells can collide, attach and form large phytodetrital aggregates (Burd and Jackson, 2009; McCave, 1984), with those reaching sizes greater than 0.5 mm known as “marine snow” (Allardedge and Silver, 1988). Alternatively, phytoplankton cells can be tightly packed into dense fecal pellets through zooplankton grazing (Silver and Gowing, 1991). Because of their large size and high density respectively, phytodetrital aggregates and fecal pellets are major constituents of the downward flux, and several studies have found either fecal pellets (Fowler and Knauer, 1986; Pilskahn and Honjo, 1987; Bishop et al., 1977; Wassmann et al., 2000; Ebersbach and Trull, 2008; Cavagna et al., 2013) or large organic aggregates (Turner, 2002; Alldredge and Gotschalk, 1989; De La Rocha and Passow, 2007; Jackson, 1990; Burd and Jackson, 2009) to be the dominant vectors of carbon to depth.

Because grazing causes losses of organic carbon by respiration (Michaels and Silver, 1988; Alldredge and Jackson, 1995), direct export via the sinking of phytodetrital aggregates represents the most efficient operating mode of the BCP. However, the ecosystem structure and environmental conditions under which primary production can be exported directly via phytodetrital aggregates are still unclear, and their determination would considerably improve the predictions of the efficiency of the BCP in varying conditions.

The volume fraction of phytodetrital aggregates vs. fecal pellets in the total flux and their volume-to-carbon-content ratio select the dominant carbon export mode; these relative contributions depend on numerous parameters, including primary productivity, biomass, interactions between primary producers and heterotrophic communities (Michaels and Silver, 1988), physical fragmentation, microbial decomposition, coprophagy and the velocity at which particles settle (Turner, 2002). The Southern Ocean contains the largest high-nutrient, low-chlorophyll (HNLC) area of the world ocean and is an essential player in global biogeochemistry (Sigman and Boyle, 2000). In these waters, abundant macronutrients (sili-cic acid, nitrate and phosphate) can fuel primary production given available light and sufficient iron, a limiting micronutrient (de Baar et al., 1995; Martin, 1990). The Kerguelen Plateau offers the opportunity to study the functioning of the BCP in a naturally iron-fertilised region (Blain et al., 2007). The first Kerguelen Ocean and Plateau compared Study (KEOPS1), demonstrated that most of the sinking flux collected in polycrylamide gel sediment traps was derived from copepod fecal detritus (intact or degrading pellets and fecal material reaggregated with phytodetritus, hereafter called “fecal aggregates”), and reported limited evidence for phytodetrital aggregates formed by direct flocculation of phytoplankton cells (Ebersbach and Trull, 2008). Number and volume fluxes were dominated by aggregates but represented a small fraction of the total carbon flux, owing to their low volume-to-carbon-content ratio. Several natural and artificial iron-fertilisation experiments conducted at the same time of the year but in different locations in the Southern Ocean (e.g. SAZ-Sense study and SOFeX) displayed similar export modes relying mainly on fecal matter (Bowie et al., 2011; Ebersbach et al., 2011; Coale et al., 2004; Lam and Bishop, 2007). In contrast, other artificial and natural iron experiments (SOERE, CROZEX and EiFeX) have demonstrated a direct export via the sinking of phytodetrital aggregates or single phytoplankton cells (Boyd et al., 2000; Waite and Nodder, 2001; Pollard et al., 2007; Salter et al., 2007; Smetacek et al., 2012). These variations among studies may reflect the time-varying aspects of export. In his review of Southern Ocean ecosystem carbon contribution to carbon export, Quéguiner (2013) suggests that from the onset of a bloom to its decline and subsequent export event, phytoplankton, and to a lesser extent zooplankton communities, is subject to several rapid successions. The complexity of the processes is also reflected by the past 30 years of empirical and modelling studies attempting to relate deep carbon export variations to surface productivity (Eppley and Peterson, 1979; Suess, 1980; Wassmann, 1990; Guidi et al., 2009). In general, the ratio between export and production in the surface ocean is low (<5–10 %; Buesseler, 1998), but decoupling associated with high-export events (e.g. high-latitude blooms), or even negative relationships, has been noted (Maiti et al., 2013; Buesseler, 1998;
Ebersbach et al., 2011; Lam and Bishop, 2007). This highlights the complexity of food web structure and its multiple controls on carbon export (Wassmann, 1998; Michaels and Silver, 1988).

In the present study we test the hypothesis that direct export via phytodetrital aggregates occurs during the early stage of the Kerguelen naturally iron-fertilised bloom, when zooplankton communities present in the water column are not fully developed. We further explore the relative export abilities of each carbon export mode (i.e. phytodetrital aggregates vs. fecal pellets) by looking at their variation with depth and over time and their links to spatio-temporal variations in plankton communities.

We collected sinking particles in free-drifting polyacrylamide gel and standard sediment traps. Gel traps allowed for the collection of intact natural particles as they sank in the water column (Ebersbach and Trull, 2008; Janasch et al., 1980; McDonnell and Buesseler, 2010), and thus gave a direct “picture” of the sinking flux at the depth of trap deployment. Image analysis of particles embedded in gels provided particle statistics (e.g. number and volume fraction of each category of particle), and conversion from area to volume and from volume to carbon content, using empirical relationships, allowed for estimation of the carbon flux and the relative importance of each category of particle. In parallel, standard sediment traps serving as a reference permitted direct quantitative estimates based on bulk chemical analyses of the material collected and from $^{234}$Th depletion method (Planchnon et al., 2014). Then, to test our main hypothesis, the relative contribution of each category of particles was linked to the amount of carbon effectively exported in order to determine which one led the carbon export.

2 Material and methods

2.1 The KEOPS2 study

The second KErguelen Ocean and Plateau compared Study (KEOPS2) was conducted onboard the RV Marion Dufresne over and downstream of the Kerguelen Plateau, from 8 October to 30 November 2011. Sinking particle flux and composition were assessed by the use of free-drifting sediment traps deployed at six stations, inside and outside the naturally iron-fertilised area, in waters with varying biomass and surface chlorophyll $a$ (Chl $a$) levels (Figs. 1 and 2). For more information on the complex spatio-temporal evolution of the phytoplankton bloom over the full 2011–2012 annual cycle, we refer the reader to an animation of NASA MODIS Aqua chlorophyll images, provided as a supplementary material in Trull et al. (2014). Combination of sediment trap collection with volume-to-carbon conversion factors allowed to determine preferential modes of carbon export (Ebersbach and Trull, 2008; Ebersbach et al., 2011).

2.2 Water column properties and biomass at each station

In addition to trap-derived measurements, POC concentrations were estimated in the water column using a WET Labs C-Star (6000 m) transmissiometer (660 nm wavelength and 25 cm path length) linked to a conductivity–temperature–depth (CTD) system (Sea-Bird SBE-911+CTD). Xmiss transmissiometer data (%) were converted to POC concentrations ($\mu$mol L$^{-1}$) following a calibration based on in situ POC measurements from Niskin bottles. A Seapoint Chelsea Aquatracka III (6000 m) chlorophyll fluorometer linked to the CTD was used to determine fluorescence profiles. Fluorescence was converted to chlorophyll $a$ (Chl $a$; $\mu$g L$^{-1}$) by comparison with total Chl $a$ in situ measurements from Niskin bottles (Lasbleiz et al., 2014).

Figure 2 shows water column properties and biomass at each station. The HNLC reference station R-2 located outside the fertilised area was characterised by a relatively deep mixed layer (96 m), low net primary productivity (NPP) (euphotic zone 1% PAR-integrated NPP = 135 ± 6 mg C m$^{-2}$ d$^{-1}$; Cavagna et al., 2014), low surface chlorophyll (chlorophyll $a$ mixed layer average = 0.6 $\mu$g Chl $a$ L$^{-1}$), and biomass (mixed-layer-integrated POC = 4.7 g C m$^{-2}$). Stations E-1, E-3 and E-5 were located in an eddy-like, bathymetrically trapped recirculation feature in deep waters east of the Kerguelen Islands (stationary meander of the polar front), with a mixed layer depth varying from 33 (E-3) to 70 m (E-1). These stations had moderate NPP (523 ± 55, 686 ± 97 and 943 ± 113 mg C m$^{-2}$ d$^{-1}$ respectively), Chl $a$ (0.8, 0.7 and 1.1 $\mu$g Chl $a$ L$^{-1}$ respectively), and biomass (5.3, 3 and 4.8 g C m$^{-2}$ respectively). They were used as a time series assuming a pseudo-Lagrangian evolution (d’Ovidio et al., 2014). F-L was the only station located north of the polar front and exhibited the shallowest mixed layer (31 m). A3-2 was the second visit to the on-plateau bloom reference station of KEOPS1 and had the deepest mixed layer (149 m). F-L and A3-2 displayed the highest NPP (3.4 ± 0.1 and 1.9 ± 0.2 g C m$^{-2}$ d$^{-1}$ respectively), chlorophyll $a$ (3 and 1.8 $\mu$g Chl $a$ L$^{-1}$ respectively) and biomass (6.2 and 20.4 g C m$^{-2}$).

2.3 Sediment trap preparation, deployments and recovery

Two different types of trap were deployed during KEOPS2. Bulk fluxes of particulate organic carbon (POC), total particulate nitrogen (TPN), biogenic silica (BSi), particulate inorganic carbon (PIC), particulate iron (PFe; data shown in Bowie et al., 2014) and thorium $^{234}$ (Th) were estimated using PPS3/3 traps (Technicap, La Turbie, France). A PPS3/3 trap consists of a single cylindrical trap with an internal conical funnel at its base with a collection area of 0.125 m$^{2}$ that transfers samples into a carousel of 12 cups. During KEOPS2, these traps were deployed for a maximum period
of 6 days. Cups were filled with brine with a salinity of $\sim$52 psu, made by freezing filtered (0.2 µm pore size) surface seawater. Some cups were also amended with mercuric chloride (1 g L$^{-1}$) as a biocide (as detailed in Table 4). No poison was added to the cups used for trace metal studies (Bowie et al., 2014).

To examine sinking flux characteristics (particle type, number and size), intact particles were also collected in cylindrical polyacrylamide gel-filled sediment traps with a collection area of 0.011 m$^2$. These deployments lasted less than 2 days so as not to overload the gels (Table 1). Polyacrylamide gels were prepared following the method developed by Lundsgaard (1995), modified as described in Ebersbach and Trull (2008).

Due to different required deployment durations (shorter for gel traps to avoid overloading; see above), each category of trap was deployed on separate arrays, except at A3-2 (combined deployment, Table 1). All separated deployments of gel and PPS3/3 traps overlapped in time and location (except at station E-3, where they were successive), to optimise the collection of similar particle fields. The arrays had broadly the same design consisting of a surface float sustaining a mooring line where the traps were fixed at different depths. PPS3/3 traps were fixed at 210 m, and one to four gel traps, depending on the station, were fixed at 110, 210, 330 and 430 m. Wave-induced motions were dampened by an elastic link to keep the traps at a constant depth (Trull et al., 2008). Pressure sensors mounted on the deepest gel trap and PPS3/3 trap on most of the arrays confirmed very small vertical motions during the deployments, with depth standard deviations ranging from 0.6 m at E-1 to 2.4 m at E-5 (Table 1). The average trap drift speed of 8.5 ± 5 cm s$^{-1}$ was in the range of horizontal velocities determined by drogued drifter trajectories (Zhou et al., 2014). Inclinometers recorded small tilts of the mooring lines (from 0.3 ± 1$^\circ$ at E-3 to a maximum of 4 ± 1.7$^\circ$ at E-5), guaranteeing minimum perturbation of particle collection due to hydrodynamic conditions. No particular difficulties were encountered.

Figure 1. MODIS-Aqua satellite (CLS-CNES) images of surface chlorophyll a concentration (Chl a) at different bloom stages from 28 October to 20 November 2011. Images show free-drifting sediment trap deployment locations in contrasting biomass levels. On each map, red labels represent the station(s) sampled at the date of the map ±3 days.
Figure 2. Water column properties and biomass at each site. Chl $a$: chlorophyll $a$ ($\mu$g L$^{-1}$); $\sigma_\theta$: potential density anomaly (kg m$^{-3}$); POC: particulate organic carbon ($\mu$mol L$^{-1}$). Grey lines indicate CTD profiles and black lines represent their average values. $E_\text{z} 1\%$ PAR: base of the euphotic zone assumed at 1% of the photosynthetic available radiation (PAR).

during trap recoveries, ensuring unperturbed gel structure. The seawater overlying the gels was removed directly after recovery to prevent particles collected in the trap cylinder during the recovery from entering the gels. Unfortunately, the PPS3/3 trap array deployed at R-2 was lost.

2.4 Chemical analysis

Protocols used for particulate organic carbon (POC), total particulate nitrogen (TPN), particulate inorganic carbon (PIC) and biogenic silica (BSi) analyses are described in Trull et al. (2008). $^{234}$Th flux analysis is detailed in Planchon et al. (2014).

2.5 Image analysis

Within a few hours after recovery, each gel was photographed onboard against a laser-etched glass grid of 36 cells (each 14 mm $\times$ 12.5 mm) at a magnification of $\times 6.5$ using a light field transmitted illumination and a Zeiss Stemi 2000-CS stereomicroscope coupled to a Leica DFC-280 1.5 MP digital camera and Leica Firecam software on an Apple iMac G4 computer. Observations at higher magnification (from $\times 10$ to $\times 50$) confirmed particle identifications when needed.

Pictures of incomplete grid cells, with inequally distributed particles or large zooplankton, were removed from the analysis to avoid bias. Ten grid cells per gel (total of 180 pictures) were selected randomly. The average sum of the surface analysed per gel was $15.7 \pm 0.7$ cm$^2$, corresponding to $14.3 \pm 0.7\%$ of the trap collection area.

Particles collected in gels (Fig. 3) were phytodetrital aggregates (PA), cylindrical fecal pellets (CFP), oval fecal pellets, fecal aggregates (FA) and diatoms in the form of chains (e.g. the pennate *Fragilariopsis* spp.) or single cells (e.g. the centric *Thalassiosira* spp.). A few mesozooplankton specimens were collected (less than 10 per gel), and were mostly
represented by copepods (adult and copepodite stages) and appendicularians. Foraminifera and radiolarians were also occasionally observed. Phytodetrital aggregates were loose and green, while fecal aggregates contained dense, brown material. Most cylindrical fecal pellets had sharp edges and relatively constant diameters, but some were tapered along their length and had blurred edges composed of unpacked fecal material or attached phytodetritus (Fig. 3, panel b).

A preliminary image analysis was conducted to select the best analysis method in terms of particle identification. Particles were classified into three main categories based on their significant contribution to the flux: phytodetrital aggregates, cylindrical fecal pellets and fecal aggregates. A fourth category, oval fecal pellets, was rare (less than one pellet per image in total), and its contribution to the flux was assumed negligible. Pictures were converted to binary images, with threshold levels adjusted manually on each picture to ensure a minimum alteration of particle areas. The average alteration of particle area estimated on a subsample was an increase of 21.6 ± 7% (n = 169) for particles with irregular shapes (e.g. aggregates sensu lato including phytodetrital and fecal aggregates), and an increase of 11.6 ± 7% (n = 44) for cylindrical fecal pellets. Cylindrical fecal pellet and aggregate areas were systematically corrected for this overestimation.

Pictures were analysed with the US National Institutes of Health’s free software ImageJ. Typical shapes of each category of particle were determined manually on a subsample of particles. MATLAB routines using specific sets of shape descriptors were then applied to all images to identify and separate each category of particle. Because fecal and phytodetrital aggregates had similar complex shapes, automated

Table 1. Deployment schedules for free-drifting sediment trap arrays.

<table>
<thead>
<tr>
<th>Area</th>
<th>Site ID</th>
<th>Array</th>
<th>Trap depths (m)</th>
<th>± SD (m)</th>
<th>Event</th>
<th>Time (UTC)*</th>
<th>Latitude</th>
<th>Longitude</th>
<th>Duration (days)</th>
<th>Drift (km)</th>
<th>Tilt ± SD (°)</th>
</tr>
</thead>
<tbody>
<tr>
<td>HNLC reference</td>
<td>E-1</td>
<td>Gel traps</td>
<td>110, 210, 330, 430</td>
<td>1.0</td>
<td>Deploy</td>
<td>16 Oct 2011, 00:50</td>
<td>50°38.47' S</td>
<td>66°36.47' E</td>
<td>0.92</td>
<td>3.8</td>
<td>–</td>
</tr>
<tr>
<td>Off-plateau meander</td>
<td>E-2</td>
<td>Gel traps</td>
<td>110, 210, 330, 430</td>
<td>0.9</td>
<td>Deploy</td>
<td>16 Oct 2011, 00:50</td>
<td>50°38.47' S</td>
<td>66°36.47' E</td>
<td>0.92</td>
<td>3.8</td>
<td>–</td>
</tr>
<tr>
<td>On-plateau reference</td>
<td>E-3</td>
<td>Gel traps</td>
<td>110, 210, 330, 430</td>
<td>0.9</td>
<td>Deploy</td>
<td>16 Oct 2011, 00:50</td>
<td>50°38.47' S</td>
<td>66°36.47' E</td>
<td>0.92</td>
<td>3.8</td>
<td>–</td>
</tr>
<tr>
<td>North polar front</td>
<td>E-4</td>
<td>Gel traps</td>
<td>110, 210, 330, 430</td>
<td>0.9</td>
<td>Deploy</td>
<td>16 Oct 2011, 00:50</td>
<td>50°38.47' S</td>
<td>66°36.47' E</td>
<td>0.92</td>
<td>3.8</td>
<td>–</td>
</tr>
</tbody>
</table>

* Times and locations are at the start of the deck operation.

Figure 3. High-resolution pictures of particles embedded in polyacrylamide gels showing the main categories of particles collected. Panel (a) – a: Phytodetrital aggregate; b: oval fecal pellet; b: radiolarian; d: foraminifera; panel (b) large cylindrical fecal pellet; panel (e) – e: small and large centric diatom single cells; f: chains of pennate diatoms of the genera Fragilaria spp.; g: chain of small centric diatom cells; panel (d) fecal aggregate. Note the difference in compactness and optical density between phytodetrital and fecal aggregates.
routines could not separate these particles efficiently. All fecal material was thus isolated manually from all other particles based on the assumption that fecal matter is brown and denser than biologically unprocessed phytoplankton (Ebersbach et al., 2011). From the resulting set of pictures, fecal aggregates were easily separated from cylindrical fecal pellets due to their very contrasted shapes. Tests on the efficiency of our automated selection, conducted on a large sample, showed that 93.4% \((n = 397)\) of cylindrical fecal pellets and 67.2% \((n = 171)\) of fecal aggregates were correctly identified by the set of shape descriptors chosen.

All particle characteristics investigated in this study and their units are reported in Table 2. An area cut-off applied at 0.004 mm\(^2\) \((0.07 \text{ mm equivalent spherical diameter})\) removed all “fake particles” deriving from small gel imperfections and glass grid or microscope lens cleanliness. This cut-off removed 38% of the total number of particles (mostly spurious particles and small single cells) but represented a loss of only 5.2% of the total area of particles in the images, introducing a negligible bias.

Aggregate area was converted to equivalent spherical diameter (ESD) assuming spherical shape, and the volume was calculated from the ESD. Because cylindrical fecal pellets were not always straight, their volume could not be accurately measured directly from their length and was calculated from their perimeter and area (independent of pellet curvature), assuming a cylinder. The radius \(r\) of the cylinder section was determined by finding the minimum root of the polynomial

\[
4r^2 - Pr + A = 0,
\]

where \(P\) is the perimeter and \(A\) is the projected area of the aggregate. The length \(L\) was calculated from the projected area and radius using the formula

\[
L = A/2r.
\]

The volume was then calculated from the radius and length.

The conversion from volume to carbon content was done by using different ratios and relationships depending on the particle considered. Figure 4 shows the relationship between carbon content and particle volume for different algorithms from the literature and those selected in this study. Based on values published by González and Smetacek (1994), the volume of cylindrical fecal pellets was converted to their organic carbon content using a ratio of 0.036 mg C mm\(^{-3}\) \((Fig. 4, line 2)\), as an average value for copepod \((Fig. 4, line 1)\), and euphausiid fecal pellets \((Fig. 4, line 3)\). For fecal aggregates, we used the power relationship between POC content and aggregate volume \(V\), \(\text{POC} (\mu g) = 1.05V (\text{mm}^3)^{0.51}\), based on the fractal decrease in carbon content with size and determined empirically by Allard (1998) for fecal marine snow \((Fig. 4, line 4)\). The volume of phytodetrital aggregates was converted to carbon content using also a power relationship determined by Allard (1998) for diatom marine snow, \(\text{POC} (\mu g) = 0.97V (\text{mm}^3)^{0.56}\) \((Fig. 4, line 5)\), assuming aggregates composed of phytoplankton not biologically processed. In contrast to Ebersbach and Trull (2008; Fig. 4, line 6), very small particles (large single cells and aggregates composed of few cells) were included in the category of phytodetrital aggregates and their volume-to-carbon conversion was done using the same relationship \((Fig. 4, line 5)\).

Particle number and volume fluxes are presented in Sect. 3 as a function of size spectra. All particles were binned in 10 size classes spaced logarithmically to give the best representation of the whole size range \((Jackson et al., 1997, 2005)\). To avoid bias, bins containing five or fewer particles were not included in the flux spectrum analyses, as recommended by Jackson et al. (2005).

3 Results

3.1 Particles collected in polyacrylamide gel-filled sediment traps

3.1.1 Particle number, projected area and volume fluxes

Despite variations in deployment duration among sites exceeding 80% \((between 0.9 and 5.3 \text{ days, Table 1})\), an observation of raw images \((Fig. 5)\) gives a broad preliminary
Table 2. Particle characteristics and bins for phytodetrital aggregates, cylindrical fecal pellets and fecal aggregates.

<table>
<thead>
<tr>
<th>Characteristics name</th>
<th>Unit</th>
<th>Definition</th>
</tr>
</thead>
<tbody>
<tr>
<td>Projected area</td>
<td>cm²</td>
<td>Pixel area of the particle image</td>
</tr>
<tr>
<td>Volume</td>
<td>cm³</td>
<td>Volume calculated from area</td>
</tr>
<tr>
<td>Equivalent spherical diameter (ESD)</td>
<td>cm</td>
<td>Diameter of a sphere with the same image area</td>
</tr>
<tr>
<td>Perimeter</td>
<td>cm</td>
<td>Sum of pixel lengths at particle edge</td>
</tr>
<tr>
<td>Length</td>
<td>cm</td>
<td>Major axis of ellipse fit to particle</td>
</tr>
<tr>
<td>Numerical flux</td>
<td>m⁻² d⁻¹</td>
<td>Number flux of sinking particles</td>
</tr>
<tr>
<td>Volume flux</td>
<td>cm³ m⁻² d⁻¹</td>
<td>Volume flux of sinking particles</td>
</tr>
<tr>
<td>Carbon flux</td>
<td>mg C m⁻² d⁻¹</td>
<td>Organic carbon flux in sinking particles</td>
</tr>
<tr>
<td>Number flux spectrum</td>
<td>cm⁻¹ m² d⁻¹</td>
<td>Number flux per unit ESD size interval</td>
</tr>
<tr>
<td>Volume flux spectrum</td>
<td>cm³ m⁻² d⁻¹ cm⁻¹</td>
<td>Volume flux per unit ESD size interval</td>
</tr>
<tr>
<td>Number, volume and carbon flux fractional contributions</td>
<td>none</td>
<td>Number, volume and carbon flux of particle types as a fraction of total</td>
</tr>
</tbody>
</table>

Bins (cm) | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
<table>
<thead>
<tr>
<th></th>
<th></th>
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<th></th>
<th></th>
<th></th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td>Lower limits (ESD)</td>
<td>0.0071</td>
<td>0.0102</td>
<td>0.0145</td>
<td>0.0207</td>
<td>0.0296</td>
<td>0.0422</td>
<td>0.0603</td>
<td>0.0860</td>
<td>0.1228</td>
<td>0.1752</td>
</tr>
<tr>
<td>Upper limits (ESD)</td>
<td>0.0102</td>
<td>0.0145</td>
<td>0.0207</td>
<td>0.0296</td>
<td>0.0422</td>
<td>0.0603</td>
<td>0.0860</td>
<td>0.1228</td>
<td>0.1752</td>
<td>–</td>
</tr>
<tr>
<td>Centre (ESD)</td>
<td>0.0087</td>
<td>0.0124</td>
<td>0.0176</td>
<td>0.0252</td>
<td>0.0359</td>
<td>0.0513</td>
<td>0.0732</td>
<td>0.1044</td>
<td>0.1490</td>
<td>–</td>
</tr>
</tbody>
</table>

Figure 5. Images of sinking particles embedded in polyacrylamide gels, collected at each site at 210 m. Comparison of images suggests differences in terms of particles abundance and nature at each site.

indication on flux differences in terms of particle abundance (e.g. low fluxes at R-2 and F-L, and higher at E stations and A3-2). The lowest particle numbers, projected particle area and volume fluxes were collected at R-2 and F-L (Table 3 and Fig. 6), with particle volume fluxes of 2.5 ± 1 and 3 ± 0.7 cm³ m⁻² d⁻¹ respectively (all depths averaged). In contrast, high fluxes were collected at E stations with an average volume flux of 7.5 ± 3 cm³ m⁻² d⁻¹ (all E stations and depths averaged). Station A3-2 also presented a relatively high flux of 6.1 cm³ m⁻² d⁻¹.

Phytodetrital aggregates dominated in number at most stations and depths (49 ± 10 % of the total number of particles for all stations and depths averaged). Particles not selected automatically as phytodetrital aggregates, cylindrical fecal pellets or fecal aggregates (“others” in Table 3) represented the second-largest numerical fraction (38 ± 8 %) but less than 9 % of the total projected particle area, and thus were assumed negligible in volume fluxes. Phytodetrital aggregates also dominated the volume fluxes (45.3 ± 22 %, all stations and depths averaged), with a maximum of 70 % at A3-2. However, volumes of cylindrical fecal pellets collected at E-5 (44 ± 33 %, all depths averaged) and volumes of fecal aggregates collected at F-L (57 ± 18 %, all depths averaged) represented the highest fractions at these stations.

Projected area fluxes at all stations and depths (Fig. 6) showed a clear attenuation of the total flux between 210 and 430 m (loss of 38 ± 21 % on average), with a maximum attenuation of 74 % at E-5 (Fig. 6a). A decrease in the flux of cylindrical fecal pellets with depth was combined with an increase in the flux of aggregates (mainly phytodetrital), except
at R-2, where a general flux attenuation was observed (all particle categories), and only a small increase in phytodetrital aggregates at 430 m.

Fluxes at E stations at 110 and 210 m decreased with time between E-1 and E-3, followed by a strong increase in cylindrical fecal pellet flux at E-5 (Fig. 6c).

### 3.1.2 Number and volume flux spectra

Smallest particles were the most numerous at every site and depth (Fig. 7). Particle numbers decreased by more than 3 orders of magnitude for a 1 order of magnitude increase in depth (Fig. 7). Particle numbers decreased by more than 3 orders of magnitude for a 1 order of magnitude increase in depth (Fig. 7). Smallest particles were the most numerous at every site and depth (Fig. 7).

At all sites, most of the volume flux of phytodetrital aggregates was carried by middle-sized particles (ESD of 0.01–0.1 cm), followed by the same spectra. Most cylindrical fecal pellets and fecal aggregates contributed significantly to smaller particles (except at R-2, where the largest cylindrical fecal pellets and fecal aggregates contributed significantly to the volume flux).

The most notable change in the number flux spectra with depth was observed for middle-sized cylindrical fecal pellets at E stations, for which a decrease in number was generally combined with an increase in size. E-1 presents the best illustration, with most of the cylindrical fecal pellets with a size around 0.01 cm at 110 m increasing to 0.06 cm at 210 m.

### 3.1.3 POC flux from image analysis

The lowest carbon fluxes were estimated at R-2 and F-L (Table 3), with values of 27 ± 12 and 36 ± 10 mg C m⁻² d⁻¹ respectively (all depths averaged). The highest carbon fluxes were observed at E stations (107 ± 33 mg C m⁻² d⁻¹, all E stations and depths averaged), with a maximum value of 180 mg C m⁻² d⁻¹ at E-5, 110 m. A3-2 presented a moderate carbon flux of 66 mg C m⁻² d⁻¹ at 210 m.

Cylindrical fecal pellets carried most of the carbon flux at all stations and depths, with an average fractional contribution of 56 ± 19 % (Table 3). This was particularly true at E stations, where fecal pellets drove on average 63 ± 17 % of the carbon flux (maximum of 88 % at E-5, 110 m), and at F-L (62 ± 20 %, all depths averaged). However, at several stations, a transition was observed at 430 m, where phytodetrital ag-

<table>
<thead>
<tr>
<th>Site ID</th>
<th>Depth (m)</th>
<th>Numerical flux (#10⁶ m⁻² d⁻¹)</th>
<th>Fractional contributions</th>
<th>Volume flux (cm³ m⁻² d⁻¹)</th>
<th>Fractional contributions</th>
<th>POC flux (mg C m⁻² d⁻¹)</th>
<th>Fractional contributions</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td></td>
<td>PA</td>
<td>CFP</td>
<td>FA</td>
<td>O</td>
<td>PA</td>
</tr>
<tr>
<td>R-2</td>
<td>210</td>
<td>0.46</td>
<td>0.03</td>
<td>0.05</td>
<td>0.46</td>
<td>2.8</td>
<td>0.39</td>
</tr>
<tr>
<td>E-1</td>
<td>110</td>
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<td>0.16</td>
<td>0.04</td>
<td>0.28</td>
<td>5.2</td>
<td>0.34</td>
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<tr>
<td>E-3</td>
<td>110</td>
<td>0.51</td>
<td>0.07</td>
<td>0.06</td>
<td>0.36</td>
<td>8</td>
<td>0.3</td>
</tr>
<tr>
<td>E-5</td>
<td>210</td>
<td>0.56</td>
<td>0.07</td>
<td>0.31</td>
<td>13</td>
<td>13</td>
<td>0.56</td>
</tr>
<tr>
<td>F-L</td>
<td>110</td>
<td>0.45</td>
<td>0.04</td>
<td>0.09</td>
<td>0.41</td>
<td>3.5</td>
<td>0.8</td>
</tr>
<tr>
<td>A3-2</td>
<td>210</td>
<td>0.46</td>
<td>0.04</td>
<td>0.09</td>
<td>0.41</td>
<td>3.3</td>
<td>0.23</td>
</tr>
</tbody>
</table>

Table 3. Total numerical, volume and particulate organic carbon (POC) fluxes and fractional contributions of each category of particle. Maximum and minimum fluxes are indicated in bold. PA: phytodetrital aggregates; CFP: cylindrical fecal pellets; FA: fecal aggregates; O: others.
3.2 Biogeochemical fluxes collected in PPS3/3 traps

Bulk fluxes from PPS3/3 traps are reported in Table 4. The highest mass, POC, 234\(^{\text{Th}}\) and TPN fluxes were collected at E stations. POC fluxes decreased over time from 84 ± 27 at E-1, to 58 ± 18 at E-3, to 24 ± 12 at E-5 mg C m\(^{-2}\) d\(^{-1}\). A3-2 presented a POC flux of 27 mg C m\(^{-2}\) d\(^{-1}\). An average 234\(^{\text{Th}}\) activity of 988 ± 127 dpm m\(^{-2}\) d\(^{-1}\) was recorded at E stations, with a maximum of 1129 ± 177 dpm m\(^{-2}\) d\(^{-1}\) at E-3. 234\(^{\text{Th}}\) fluxes are detailed in Planchon et al. (2014). Over all sites, BSi fluxes were very high (7 ± 2 to 21 ± 10 mmol BSi m\(^{-2}\) d\(^{-1}\)), suggesting the large contribution of diatoms to the phytoplankton community. Conversely, very low particulate inorganic carbon (PIC) fluxes (1–4 orders of magnitude lower than POC fluxes) suggested the limited role of calcium carbonate (CaCO\(_3\)) in biogenic mineral fluxes. POC:TPN ratios were close to the canonical Redfield ratio of 6.6 for phytoplankton at all stations except E-5 (7.5), which also displayed the lowest POC:BSi ratio (0.1). At E stations, POC:234\(^{\text{Th}}\) and POC: mass ratios decreased over time (POC:234\(^{\text{Th}}\) ratios from 8 at E-1 to 2.1 \(\mu\)mol dpm\(^{-1}\) at E-5; POC: mass ratio from 0.05 at E-1 to 0.03 g g\(^{-1}\) at E-5), suggesting an attenuation of export fluxes combined with a degradation of sinking particles. A3-2 displayed POC:234\(^{\text{Th}}\) and POC: mass ratios of 4.4 \(\mu\)mol dpm\(^{-1}\) and 0.06 g g\(^{-1}\) respectively. In general, no consistent differences in fluxes could be resolved between poisoned and unpoisoned cups.

3.3 POC flux comparisons and export efficiencies

POC fluxes determined from gel images (using particle volume-to-carbon-content conversion factors) were in the same range of values as those determined from particle collection in PPS3/3, with maximum differences at a same station never exceeding 1 order of magnitude (Tables 3 and 4). POC fluxes from PPS3/3 were systematically lower than those derived from image analysis (on average 57 ± 22 \% less).

E-ratios, calculated as the ratio of POC fluxes from gel image analysis to 1 \% PAR-integrated net primary productivity (Cavagna et al., 2014; Table 5) indicated a high export efficiency at R-2 and E-1 (0.2±0.08 and 0.23±0.07 respectively, all depths averaged), intermediate at E-3 and E-5 (0.1 ± 0.02 and 0.13 ± 0.09 respectively, all depths averaged), and very low at F-L (0.01 ± 0.0, similar value at all depths) and A3-2 (0.03). E-ratios derived from POC fluxes estimated from PPS3/3 traps showed lower values but followed the same trend: E-1 > E-3 > E-5 > A3-2. Export efficiencies derived from 234\(^{\text{Th}}\) disequilibria, ThEC (Planchon et al., 2014), are shown in Table 5 for comparison, and are discussed in the next section.

According to calculations based on gel trap POC flux and transmissometer POC concentration estimates (Fig. 2), E stations exported the largest percentage of their mixed-layer integrated POC (ΣPOC\(_{\text{ML}}\)) per day (2.4 ± 1 \%, all E stations and depths averaged) with the maximum observed at E-5 (2.7 ± 1.8 \%, all depths averaged) and values of 2.3 ± 0.7 and 2.3 ± 0.5 \% at E-1 and E-3 respectively (all depths averaged). R-2 and F-L exported respectively 0.58 ± 0.2 and 0.59 ± 0.15 \% of their ΣPOC\(_{\text{ML}}\) per day (all depths averaged), and A3-2 exported 0.32 \% of its ΣPOC\(_{\text{ML}}\) per day (210 m). A similar trend was observed using POC fluxes from PPS3/3 traps (E stations > A3-2).
4 Discussion

4.1 Comparison of POC flux estimations

Two different approaches were used to estimate POC fluxes. PPS3/3 trap collection providing a direct determination of the flux served as a reference method. POC fluxes estimated from image analysis of particles embedded in polyacrylamide gels were in the same range as those derived from PPS3/3 but were systematically higher (see Sect. 3). This difference is most likely due to the uncertainty in the volume-to-carbon conversion factors (Fig. 4) used to estimate POC fluxes from particle image analysis. A comparison with the direct estimation of bulk fluxes collected in PPS3/3 suggests that our volume-to-carbon-content conversion factors tended to slightly overestimate the carbon carried by sinking particles (Tables 3 and 4), especially at E-5, where it was up to 7-fold higher. At this station the large contribution of cylindrical fecal pellets to the volume flux (Table 3; 72% at 110 m and 51% at 210 m) suggests that the volume-to-carbon conversion factor used for these particles may be responsible for the mismatch observed. The value of 0.036 mg C mm$^{-3}$ used as an average for copepod and euphausiid fecal pellets may not reflect the actual carbon con-
Table 4. Particle fluxes at 210 m depth from free-drifting deployments of the 12-cup-carousel cylindrical PPS3/3 trap. The trap collection area was 0.125 m$^2$. Particles were washed through a 350 µm Nitex screen to remove zooplankton and were collected on a 1 mm silver filter. Mean values and their standard deviations are indicated in bold.

<table>
<thead>
<tr>
<th>Cup ID</th>
<th>Opening time (UTC)$^a$</th>
<th>Poison</th>
<th>Mass flux</th>
<th>POC flux</th>
<th>TPN flux</th>
<th>BS flux</th>
<th>PIC flux</th>
<th>$^{234}$Th flux</th>
<th>POC : TPN</th>
<th>POC : BS</th>
<th>POC : PIC</th>
<th>$^{234}$Th : massID</th>
<th>$^{234}$Th : g C</th>
<th>$^{234}$Th : mmol</th>
<th>$^{234}$Th : dpm</th>
<th>$^{234}$Th : µmol dpm</th>
<th>$^{234}$Th : g g$^{-1}$</th>
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<td>84.31</td>
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<tr>
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<td>84.13</td>
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<td>752</td>
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<td></td>
</tr>
</tbody>
</table>

$^a$ Local time was UTC + 5 h.

$^b$ For all stations, mean values are the total collection divided by the total time over the entire deployment.

$^c$ For all stations, flux standard deviations are weighted by cup duration times. Component ratio standard deviations are unweighted.

n.r.: not rotated.

NA: not available.

Blk: blank.
Table 5. Export efficiency at each site estimated from several methods. Maximum and minimum export efficiencies are indicated in bold.

<table>
<thead>
<tr>
<th>Site</th>
<th>Depth (m)</th>
<th>E-ratios</th>
<th>% ( \sum \text{POC}_{\text{ML}} \text{ export 1 d} )</th>
<th>Gels</th>
<th>PPS3/3</th>
<th>Gels</th>
<th>PPS3/3</th>
</tr>
</thead>
<tbody>
<tr>
<td>R-2</td>
<td>100 ± 10</td>
<td>0.32</td>
<td>–</td>
<td>0.34</td>
<td>0.92</td>
<td>–</td>
<td>–</td>
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<tr>
<td></td>
<td>200 ± 10</td>
<td>0.22</td>
<td>–</td>
<td>0.16</td>
<td>0.64</td>
<td>–</td>
<td>–</td>
</tr>
<tr>
<td></td>
<td>330</td>
<td>0.15</td>
<td>–</td>
<td>0.43</td>
<td>–</td>
<td>–</td>
<td>–</td>
</tr>
<tr>
<td></td>
<td>430 ± 10</td>
<td>0.12</td>
<td>–</td>
<td>0.34</td>
<td>–</td>
<td>–</td>
<td>–</td>
</tr>
<tr>
<td>E-1</td>
<td>100 ± 10</td>
<td>0.21</td>
<td>0.27</td>
<td>2.12</td>
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<td>–</td>
<td>–</td>
</tr>
<tr>
<td></td>
<td>200 ± 10</td>
<td>0.34</td>
<td>0.16 ± 0.05</td>
<td>0.18</td>
<td>3.34</td>
<td>1.59 ± 0.52</td>
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</tr>
<tr>
<td></td>
<td>330</td>
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<td>–</td>
<td>2.05</td>
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<tr>
<td></td>
<td>430</td>
<td>0.18</td>
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<td>1.82</td>
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</tr>
<tr>
<td>E-3</td>
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<tr>
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<td></td>
<td>430</td>
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<td>1.86</td>
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<tr>
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<td>200 ± 10</td>
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</tr>
<tr>
<td>F-L</td>
<td>100 ± 10</td>
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<td>0.01</td>
<td>0.73</td>
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<tr>
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<td>200 ± 10</td>
<td>0.01</td>
<td>–</td>
<td>0.01</td>
<td>0.62</td>
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<tr>
<td></td>
<td>430</td>
<td>0.01</td>
<td>–</td>
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<td>A3-2</td>
<td>100 ± 10</td>
<td>0.001</td>
<td>0.001</td>
<td>0.02</td>
<td>0.32</td>
<td>0.13 ± 0.04</td>
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</tr>
<tr>
<td></td>
<td>200 ± 10</td>
<td>0.03</td>
<td>0.001 ± 0.004</td>
<td>0.02</td>
<td>0.32</td>
<td>0.13 ± 0.04</td>
<td>–</td>
</tr>
</tbody>
</table>

E-ratios: POC flux (gels, PPS3/3)/NPP (E-3 100 ± 100, 200 ± 100); data from Cavagna et al., 2014). ThE = POC flux (234Th; data from Planchon et al., 2014)/NPP. % \( \sum \text{POC}_{\text{ML}} \text{ export 1 d} \): percentage of mixed-layer-integrated POC exported in 1 day.

4.2 Evolution of the flux at depth

POC fluxes presented in Fig. 8 were estimated through two different approaches: gel trap image analysis (at 110, 210, 330 and 430 m) and total \(^{234}\text{Th}\) activity measured at 11–14 depths at all stations (Planchon et al., 2014) and calculated at 100, 150 and 200 m. Fluxes estimated from PPS3/3 trap collection at only one depth (210 m) are not presented here. Figure 8 shows the evolution of POC fluxes with depth and its comparison with the empirical flux attenuation known as the “Martin curve” (Martin et al., 1987), estimating the flux at depth from values at 100 m ranging from 20 to 500 mg C m\(^{-2}\) d\(^{-1}\). Agreements between POC flux determination methods and this empirical relationship were the best for R-2 and F-L, showing a continuous attenuation of the flux with depth, but always at a lower rate than predicted by the Martin curve.

For all other stations POC fluxes above 210 m presented complex patterns suggesting distinct POC export episodes to be more likely than a continuous downward flux. Between 210 and 430 m the attenuation of POC fluxes estimated from the gel traps tends to be more consistent with the Martin curve, except for E-5, which displayed a strong decrease (as already noted in the Sect. 3). A fecal pellet loss at depth was particularly strong at E-5, due to the large role played by these particles at this site, but was observed at all stations (Fig. 6).

Our data revealed two major trends of particle flux evolution with depth: (i) the fecal pellet flux decreased, and (ii) phytodetrital and fecal aggregate fluxes remained constant or even increased. Establishing a link between these two processes is tempting. It suggests the importance of physical reaggregation in sustaining the carbon flux at depth from fecal pellets that have undergone bacterial degradation or zooplankton coprophagy (Suzuki et al., 2003; Lampitt et al., 1990; Iversen and Poulsen, 2007). A recent study from Giering et al. (2014) suggests that half of fast-sinking particles in the twilight zone of the eastern Atlantic Ocean (between 50 and 1000 m) are fragmented and ingested by zooplankton, and that more than 30% may be released as suspended and slowly sinking organic matter. Even if the gel trap technique does not offer enough information on
aggregation processes and particle sources to permit any clear conclusion, the hypothesis of a reaggregation of unpacked fecal pellets into “secondary” phytodetrital aggregates still deserves careful consideration.

Since the rate of physical aggregation is largely controlled by particle concentration (Jackson, 1990), a reaggregation at depth implies that sufficient material has been released by fecal pellet disaggregation. If single cells represented most of the material released during fecal pellet disaggregation, their concentration should have increased with depth in the case of no secondary aggregation or remained constant as a balance between aggregate formation and loss by sinking (notion of critical concentration; Jackson, 1990, 2005). The number flux spectra (Fig. 7) suggest that the smallest particles had a constant concentration until 210 m at almost every site. Station E-3 shows an increase in the number of small particles between 110 and 210 m and then a decrease at 430 m, which could indicate reaggregation processes occurring at depth. This decrease at 430 m is also observable at E-5 and F-L. However, data evaluation in this way implies a steady-state assumption which considers that traps measured the occurrence of a unique sinking event; the flux collected at depth being a direct temporal evolution of the same shallower flux. This appears unlikely considering the episodic nature of export and its dependence on highly dynamical ecosystem interactions responsible for high flux variability at short spatio-temporal scales as evidenced by the PPS3/3 individual cup variations (Table 4). In addition, if assuming phytodetrital aggregates at E-3, sinking at an average velocity of 150 m d⁻¹ (based on results from Laurenceau-Cornec et al., 2015), a particle field would need approximately 1.5 days to sink from 210 to 430 m, neglecting any advection. Considering this calculation and the short trap deployment at E-3 (1.02 days), a non-steady-state assumption appears more reasonable, and the increase in phytodetrital and fecal aggregates observed at depth could reflect an earlier production event.
4.3 Temporal POC flux variations during KEOPS2 and comparison with KEOPS1

From E-1 to E-5, the POC flux varied with the depth and estimation method. Collection of POC flux in PPS3/3 trap at 210 m revealed a monotonic decrease in the flux with time (Table 4). Temporal evolution of the flux between E-1, E-3 and E-5, at 100±10 and 200±10 m, using gel trap and 234Th methods (Planchnon et al., 2014), shows a almost constant flux (undistinguishable differences within the uncertainties). At 430 m, gel traps measured flux evolutions comparable to those identified in the PPS3/3 at 210 m, i.e. a continuous decrease in the flux with time. With the results from 110 and 210 m at E-5 excluded (likely linked to an episodic flux of euphausiid fecal pellets at these depths; see text above), the gel traps also show a decrease in the total flux over time, consistent with PPS3/3 trap method. The unusual increase at E-5, against the steady background of the other E stations, highlights the importance of zooplankton in modifying the particle flux.

At the KEOPS1 (January–February 2005) bloom reference station A3, POC flux values estimated at 200 m from gel trap image analysis and PPS3/3 traps were 62 and 13–20 mg C m⁻² d⁻¹ respectively (Ebersbach and Trull, 2008), i.e. in the same range as during KEOPS2 at the same station and using the same methods (gels: 66 mg C m⁻² d⁻¹; PPS3/3: 27 mg C m⁻² d⁻¹). During KEOPS1, the 234Th-based method assuming a non-steady-state system (NSS) yielded 200 m POC fluxes of 294 mg C m⁻² d⁻¹ at A3 (flux averaged over 21 days) and 124 mg C m⁻² d⁻¹ at the KEOPS1 HNLC reference station C11 (flux averaged over 10 days; Savoye et al., 2008). These values are well above the KEOPS2 values of 46 and 22 mg C m⁻² d⁻¹ determined at 200 m at A3-2 and R-2 respectively using the same method (average over 28 days, except for R-2 assumed in steady state; Planchnon et al., 2014). The 234Th-based method assuming NSS integrated the POC flux over a period longer than 20 days, contrasting with the 1 day to 1 week period provided by gel and PPS3/3 trap estimations.

Seasonal trends are more reliable if calculated over a longer period, and the 234Th-based method then gives the best insight into the temporal evolution of the POC flux from the onset of the bloom to its decline. 234Th results suggest that the POC flux was approximately 5- to 6-fold higher at the decline of the bloom (January–February) than during its onset (October–November), agreeing with the common view that most of the export flux occurs in late bloom stage (Wassmann, 1998). During KEOPS1, at A3 and C11, the NPP integrated within the euphotic zone was 1030±43 and 224±30 mg C m⁻² d⁻¹ respectively (based on 13C incorporation; Mosseri et al., 2008; Lefêvre et al., 2008). In comparison, values of 1903±186 and 135±6 mg C m⁻² d⁻¹ were determined at A3-2 and R-2 during KEOPS2 (euphotic zone, Ez 1 % PAR-integrated NPP based on 13C incorporation; Cavagna et al., 2014). Carbon export efficiencies estimated at 200 m, based on 234Th-derived POC export flux (reported as ThEC), were 30 % at A3 and 49 % at C11 during KEOPS1 (calculations using data from Savoye et al., 2008; Mosseri et al., 2008). In contrast, ThEC values of 2 % (NSS model) and 16 % (SS model) were calculated at 200 m at A3-2 and R-2 respectively during KEOPS2 (Planchnon et al., 2014). These results show that (i) primary productivity at the on-plateau site was approximately 2-fold higher in spring than during summer and (ii) carbon export fluxes were approximately 5-fold lower during early than late bloom stage, leading to (iii) carbon export efficiencies up to 10-fold lower during the early bloom stage (spring) than during late bloom stage (summer).

4.4 Toward an explanation of the negative relationship between primary productivity and carbon export efficiency

We examined two different export efficiency indicators (Table 5): (i) e-ratios calculated as the ratio between POC fluxes estimated from gel images or PPS3/3 traps, and net primary productivity integrated over the euphotic zone (Ez 1 % PAR; Cavagna et al., 2014), and (ii) ThEC calculated as the ratio between POC flux estimated from 234Th method and net primary productivity. KEOPS2 results suggest a negative relationship between primary productivity and carbon export efficiency, the most productive sites being those where carbon is exported the least efficiently. Figure 9a shows the relationship between primary productivity and export efficiency (with POC fluxes estimated at 200±10 m from gels, PPS3/3 traps and 234Th water column disequilibria) for KEOPS2 sites. For comparison purposes, KEOPS1 data are also indicated (Savoye et al., 2008). The empirical relationship proposed recently by Maiti et al. (2013), based on surface tethered cylindrical sediment traps and 234Th data from up to 130 stations in the Southern Ocean, is also reported. While this negative relationship has now been observed in several field studies in the Southern Ocean (Savoye et al., 2008; Morris et al., 2007; Jacquet et al., 2011), and elsewhere (e.g. González et al., 2009), the reasons for its existence remain unclear. Maiti et al. (2013) mentioned differences in trophic structure, grazing intensity, recycling efficiency, high bacterial activity, or increase in DOC export as possible explanations for high-productivity, low-export-efficiency regimes. Phytoplankton physiological state has also been suggested as a possible control of carbon export mode and efficiency (González et al., 2009), although this could not be verified here due to a generally good phytoplankton physiological state confirmed via microscopy over the course of the KEOPS2 study (M. Lasbleiz and K. Leblanc, personal communication, 2014). In addition, due to their degradation-resistant and heavily silicified valves (Hargraves and French, 1983; Kuwata and Takahashi, 1990), the abundance of diatom resting spores in the sinking flux, as observed during KEOPS1 (Armand et al., 2008), could also be a major factor to consider when evaluating
carbon export efficiency as suggested by Salter et al. (2012) and Rynearson et al. (2013).

It was beyond the scope of this study to explore each of these potential controls of carbon export efficiency. However, in the light of KEOPS1 and KEOPS2 results, phytoplankton and zooplankton community structure and their trophic relationships through grazing, seem to have played an important role in carbon export mode and efficiency via controls on sinking particle composition.

Table 6 presents a summary of site characteristics based on net primary productivity, surface plankton communities determined from Niskin bottle sampling and net haulings (most abundant species and biomass), carbon export features at ±10 m (mode and efficiency) and iron fertilisation status (e.g. HNLC or iron-fertilised). Due to their importance in export fluxes demonstrated by high BSI fluxes (see Sect. 3), only diatoms were examined in the phytoplankton community. Data are presented for all stations, but only stations R-2 and A3-2 will be discussed in detail here because of their reference status (i.e. HNLC and on-plateau bloom).

At the HNLC reference station R-2, characterised by the lowest net primary productivity (Cavagna et al., 2014), the diatom community was dominated by the heavily silified Fragilariopsis spp. (34.6 %: fraction of total live diatom cell counts; see Table 6 for sampling depths; M. Labrèze, personal communication, 2014) and Thalassionema nitzschioides (25.6 %), as well as by a limited mesozooplankton biomass represented mainly by middle-sized copepods (Carlotti et al., 2015). The export was mostly mediated through physical aggregation as suggested by the dominance of phytoplankton and fecal aggregates. The highest e-ratio estimated during KEOPS2 was observed at R-2. In contrast, the iron-fertilised on-plateau bloom station A3-2 displayed a high net primary productivity (Cavagna et al., 2014), and a diatom community largely dominated by the lightly silicified Chaetoceros subgenus Hyalochaete (87 %). The mesozooplankton biomass was up to 3-fold higher at A3-2 than at R-2. Small- and middle-sized copepods dominated, along with euphausiid eggs and appendicularians (Carlotti et al., 2015). Particles exported were mostly phytodetrital aggregates. One of the lowest e-ratios was recorded at A3-2. At E stations, used as a time series, the net primary productivity was moderate (Cavagna et al., 2014), and a shift from a high e-ratio at E-1 to a low e-ratio at E-5 was associated with plankton community shifts. This is indicated, for instance, by the remarkable increase in Chaetoceros subgenus Hyalochaete biomass from 10 and 2.3 % at E-1 and E-3 respectively to 22.5 % at E-5.

At stations R-2 and A3-2, although presenting very contrasted export efficiencies, physical aggregation seemed to dominate over biological aggregation, as suggested by the rarity of fecal pellets. If explained from this perspective, the inverse relationship between net primary productivity and export efficiency somewhat needs to be linked to the different nature of the aggregates produced at each station and their ability to export carbon to depth (e.g. slow- or fast-sinking). In parallel with the present study, roller tank experiments have been conducted to explore the influence of different phytoplankton communities on the sinking velocity of large phytodetrital aggregates and their aggregation processes (Laurenceau-Cornec et al., 2015). These experiments consisted in the physical aggregation of natural assemblages sampled with Niskin bottles at high and low biomass sites during KEOPS2. Results suggest that the proportions of different phytoplankton types forming the phytodetrital aggregates could influence their sinking velocity (and potentially their efficiency at exporting carbon) via a control on their structure and excess density. A strong relationship (r² = 0.98) was found between the proportion of small spine-forming diatom cells included in marine snow aggregates (e.g. Chaetoceros subgenus Hyalochaete) and their sinking velocity, suggesting an important role for phytoplankton morphology with regard to export efficiency. However, no evidence has been found that natural phytoplankton communities present at each site, as determined from Niskin bottle sampling (Table 6), reflect the composition of their aggregates, something required to approximate their sinking velocity from roller tank experiment results. Experimental and field studies noted that the proportions of diatoms in aggregates are not necessarily the same as their proportions in the surrounding water (Riebesell et al., 1991; Crocker and Passow, 1995; Waite and Nodder, 2001). Without direct estimation of the sinking velocity of natural aggregates formed in the water column at each station, no conclusion is possible and further investigations will be needed.

The potential control of export efficiency through zooplankton grazing is another hypothesis that we explored here. In the case of high grazing pressure, carbon export is driven mostly via fecal pellets, but these, even if sinking fast, potentially experience coprophagy or coporhexy (Suzuki et al., 2003; Lampitt et al., 1990; Iversen and Poulsen, 2007), and disaggregation processes facilitating bacterial remineralisation (Giering et al., 2014). In Fig. 9b, mesozooplankton biomass data from KEOPS2 (250 m to surface Bongo net haulings; Carlotti et al., 2015) is shown as a simple index of zooplankton abundance against export efficiency. Considering all POC flux estimation methods, a correlation has been found (n = 15, r² = 0.72, p < 0.0005), suggesting that zooplankton may exert an important control on export efficiency. In this perspective, however, E-3 presented an unexpectedly high export efficiency considering its high zooplankton biomass, suggesting that factors predominantly affecting carbon export efficiency can vary locally and over time.

In the low-productivity systems (e.g. R-2), a direct export can be efficient if processed via fast-sinking aggregates composed of heavy-silicified diatoms that are also assumed to be grazing-resistant. In contrast, in the sites of high productivity (e.g. A3-2 and F-L), the export flux can be strongly attenuated if a large fraction of the organic carbon flows
Table 6. Summary of site characteristics based on their primary productivity, plankton communities, iron-fertilisation and carbon export efficiency.

<table>
<thead>
<tr>
<th>Site ID</th>
<th>E-MLD (m)</th>
<th>NPP</th>
<th>Diatom community</th>
<th>Mesozoo. community</th>
<th>Export mode</th>
<th>E-ratio</th>
<th>Fe source – biomass/export</th>
</tr>
</thead>
<tbody>
<tr>
<td>E-1</td>
<td>64–70</td>
<td>Moderate</td>
<td>Thalassionema nitzschioides Centrics (&lt; 25 µm)</td>
<td>Ctenocalanus citer (16) + cop. (11) Oithona frigida (12)</td>
<td>PA, FA</td>
<td>Highest</td>
<td>HNLC – low biomass, high export efficiency</td>
</tr>
<tr>
<td>E-3</td>
<td>68–33</td>
<td>Moderate</td>
<td>Thalassionema nitzschioides Centrics (&lt; 25 µm)</td>
<td>Ctenocalanus citer (27) + cop. (34) Oithona frigida (4)</td>
<td>CFP, FA</td>
<td>Moderate</td>
<td>Fe-fertilised, mesoscale recirculation feature – moderate biomass export efficiency</td>
</tr>
<tr>
<td>A3-2</td>
<td>38–149</td>
<td>High</td>
<td>Chaetoceros Hyalochaete spp.</td>
<td>Euphausiids eggs (19) Paraeuchaeta sp. C1–C3 (12) Appendicularians (12) Oithona similis (8) Ctenocalanus citer (7) + cop. (8)</td>
<td>PA</td>
<td>Low</td>
<td>Fe-fertilised, on-plateau bloom site – high biomass, low export efficiency</td>
</tr>
</tbody>
</table>

E: depth of the base of the euphotic zone assumed at 1 % of the photosynthetic available radiation (PAR).

M: mixed-layer depth calculated using the density difference criterion of 0.02 ρ (potential density at MLD = potential density at 10 m + 0.02 kg m⁻³) (Park et al., 1998).

N: euphotic zone (% PAR) integrated net primary productivity (Cavagna et al., 2014).

PA: phytodetrital aggregates.

CFP: cylindrical fecal pellets.

PA: fecal aggregates.

*a* Most abundant diatom genus or species presented in decreasing order of abundance. Data from M. Lasbleiz (personal communication, 2014), from Niskin bottle samples. Sampling depths: R-2: 116 m, E-1: 80 m, E-3: 137 m, E-5: 110 m; F-L: 52 m, A3-2: 151 m. See text for fractions of total live diatom cell counts (%).

*b* Mesozooplankton community and fraction of the total mesozooplankton biomass, presented by decreasing order of abundance. Data from Carlotti et al. (2015), from day net haulings only (<250 m to the surface). Cop.: composite stage; C1–C3: development stages 1–3.

towards paths promoting its retention in the surface layer (i.e. grazing, microbial remineralisation and biomass accumulation). At A3, Christaki et al. (2014) proposed a carbon budget integrated over the mixed layer showing the carbon flows through microbial and higher trophic levels for early and late bloom stages. This budget indicates that, during KEOPS2, 2400 mg C m⁻² d⁻¹ was still available for phytoplankton biomass accumulation and/or export after subtracting the different loss terms, such as bacterial, microplankton and mesozooplankton respiration, as well as viral lysis of bacterial cells, from the gross community production (GCP). Using our carbon flux value at 200 m and phytodetrital aggregate contributions to this export, the relative fractions of the available carbon actually used for biomass accumulation and/or export can be estimated here. At A3-2, the carbon flux at 200 m was 66 mg C m⁻² d⁻¹ (gel trap results), with 41 % contributed by phytodetrital aggregates (Table 3). This leads to 27 mg C m⁻² d⁻¹ exported (1.1 % of the remaining available carbon) and 2373 mg C m⁻² d⁻¹ used for biomass accumulation (98.9 %). The same calculations can be made for the late-bloom situation using the values of 384 mg C m⁻² d⁻¹ for the carbon still available for biomass accumulation and/or export (Christaki et al., 2014), the KEOPS1 200 m POC flux at A3 (62 mg C m⁻² d⁻¹; Ebersbach and Trull, 2008), and a 36 % aggregate contribution (including both phytodetrital and mixed aggregates; Ebersbach and Trull, 2008). Results lead to 22 mg C m⁻² d⁻¹ exported (5.7 %) and 362 mg C m⁻² d⁻¹ used for biomass accumulation (94.3 %). These estimations show that the fraction of the carbon available that is exported is subject to large variations during the season (increased by a factor of ~5), while the fraction allocated to biomass accumulation varied comparatively much less (decreased by a factor of ~1.05). This suggests that A3 progressed over the whole season from a retention- to an export-dominated food web system (Wassmann, 1998), possibly related to successions of plankton communities prone to large variations in their export ability, as suggested in this study.

This general picture can be compared with the conceptual scheme of the development of planktonic communities in the Southern Ocean recently proposed by Queguiner (2013). Direct export can occur efficiently when the phytoplankton community is dominated by the large heavily silicified species (e.g. *Fragilariopsis* spp.), which are highly grazing-resistant and form fast-sinking aggregates. This type of slow-growing species develops through the whole season and forms a “persistent” background encountered at almost all sites. In bloom conditions (during the growth season), smaller fast-growing, lightly silicified species are added to the community, leading to an increased primary productivity. Because these small species are possibly less efficient at exporting carbon (e.g. rapidly grazed and/or sinking slowly).
the increase in primary productivity is not accompanied by an increase in carbon export – though 2- to 5-fold higher in sites under Fe-fertilisation influence than in the HNLC site – required to obtain a high export efficiency.

5 Conclusions

Our study conducted in early spring, during bloom initiation, demonstrated the following points:

i. Phytodetrital aggregates represented the main numerical and volume fractions of the flux, especially at depth, and could have played a major role in sustaining export fluxes where fecal pellet flux attenuation occurred. This contrasts with summertime (KEOPS1), when fecal material largely dominated the flux, while phytodetrital aggregates brought only a minor contribution to the flux (Ebersbach and Trull, 2008). However, when converted to carbon content, and where their degradation was limited, cylindrical fecal pellets still represented the dominant fraction of the flux.

ii. Primary productivity was negatively correlated to export efficiency, the sites of highest productivity being the least efficient to export carbon. This supports the emergent vision of high-productivity, low-export regimes already noted in the Southern Ocean (Lam and Bishop, 2007). The decrease in productivity from bloom initiation (KEOPS2) to its decline (KEOPS1), related to a shift from autotroph- to heterotroph-dominated regimes (i.e. production exported via phytodetrital vs. fecal material), could explain why major export tends to occur at the end of the season essentially via the sinking of fecal matter.

iii. Plankton community structure influenced by productivity regimes could have controlled export efficiency via variations in phytoplankton species and zooplankton grazing pressure (Table 6).

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