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To cite this version:

Kevin Carolan, Andres Garchitorena, Gabriel E. García-Peña, Aaron Morris, Jordi Landier, et al.. Topography and Land Cover of Watersheds Predicts the Distribution of the Environmental Pathogen Mycobacterium ulcerans in Aquatic Insects. PLoS Neglected Tropical Diseases, Public Library of Science, 2014, 8 (11), pp.e3298. 10.1371/journal.pntd.0003298. hal-01342938

HAL Id: hal-01342938
https://hal.sorbonne-universite.fr/hal-01342938

Submitted on 7 Jul 2016

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Topography and Land Cover of Watersheds Predicts the Distribution of the Environmental Pathogen *Mycobacterium ulcerans* in Aquatic Insects

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Abstract

**Background:** An understanding of the factors driving the distribution of pathogens is useful in preventing disease. Often we achieve this understanding at a local microhabitat scale; however the larger scale processes are often neglected. This can result in misleading inferences about the distribution of the pathogen, inhibiting our ability to manage the disease. One such disease is Buruli ulcer, an emerging neglected tropical disease afflicting many thousands in Africa, caused by the environmental pathogen *Mycobacterium ulcerans*. Herein, we aim to describe the larger scale landscape process describing the distribution of *M. ulcerans*.

**Methodology:** Following extensive sampling of the community of aquatic macroinvertebrates in Cameroon, we select the 5 dominant insect Orders, and conduct an ecological niche model to describe how the distribution of *M. ulcerans* positive insects changes according to land cover and topography. We then explore the generalizability of the results by testing them against an independent dataset collected in a second endemic region, French Guiana.

**Principal Findings:** We find that the distribution of the bacterium in Cameroon is accurately described by the land cover and topography of the watershed, that there are notable seasonal differences in distribution, and that the Cameroon model does not predict the distribution of *M. ulcerans* in French Guiana.

**Conclusions/Significance:** Future studies of *M. ulcerans* would benefit from consideration of local structure of the local stream network in future sampling, and further work is needed on the reasons for notable differences in the distribution of this species from one region to another. This work represents a first step in the identification of large-scale environmental drivers of this species, for the purposes of disease risk mapping.


Editor: Joseph M. Vinetz, University of California San Diego School of Medicine, United States of America

Received August 1, 2014; Accepted September 25, 2014; Published November 6, 2014

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Data Availability: The authors confirm that all data underlying the findings are fully available without restriction. All relevant data are within the paper and its Supporting Information files.

Funding: This work was supported by a grant from the French National Research agency (ANR 11-CEPL-00704 EXTRA-MU) with additional funding from the Young International Research Team of AIRD/IRD (JEAI ATOMyc) and an “Investissement d’Avenir” grant managed by Agence Nationale de la Recherche (CEBA, ref. ANR-10-LABX-25-01) through its integrative research programme BIOHOPSYS on Biodiversity and infectious diseases. KG is funded by a PhD studentship from ANR EXTRA-MU and LabEx CEBA (grant ANR-10-LABX-25-01). AG from a PhD studentship from the EHESP, and AM from a PhD studentship from Bournemouth University. GEGP received a post-doctoral fellowship from Fondation pour la Recherche sur la Biodiversité (FRB) and its Centre de Synthèse et d’Analyse sur la Biodiversité (CESAB, research programme BIODIS). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

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Introduction

Knowledge of the spatial distribution of an environmentally persistent pathogen is often key in creation of environmental hazard maps for disease control. Yet, despite the importance of this spatial information, only 4% of such pathogens have been mapped [1]. The reason for this gap in our knowledge is practical. It is often difficult to produce large maps of the distribution of these microbial pathogens as they are difficult to detect in nature. It is often difficult to produce large maps of the distribution of these microbial pathogens as they are difficult to detect in nature. A solution to this is to describe the distribution of the pathogens suitable habitat. For example, an environmentally persistent pathogenic bacterium may have a certain pH range within which it can survive, a specific range of microaerobic oxygen concentrations [2], and survive preferentially on certain algae [3]. In cases where we have a suitable range of pH, a suitable range of oxygen, and suitable algae, we expect to find the bacterium. Herein, this suitable range of microhabitat is termed the ecological niche of the species. Every species in nature, including vectors such as mosquitoes, and pathogens such as *Plasmodium* protozoans, has a unique ecological niche [4,5].

Knowledge of the distribution of suitable habitats would allow us to predict the expected distribution of the pathogen. This approach has been successfully applied to the vectors of diseases such as malaria, plague and dengue [6,7,8], but it is rarely applied to environmentally persistent pathogenic microbes. The range of suitable habitat is, practically, much easier to describe for insect vectors than for microbes. For example, the suitable habitat of mosquitos is driven by factors such as rainfall, which is much easier to describe on a large scale. To describe pH in the environment we must visit each site and use a probe at each location. This quickly becomes expensive and time consuming when we consider multiple variables, or if we wish to describe the distribution of a pathogen over large extents.

We hypothesised that these microhabitat variables could be indirectly inferred from large scale macroecological patterns. The distribution of swamp and forested environment, the shape and structure of the landscape, should predict the distribution of these microhabitats. For example, while the suitable habitat of a bacterium may be driven by the suitable combination of pH, oxygen, and algae, and other factors, the distribution of these conditions is in turn driven by the landscape. For example, the pH and oxygen content of water in swamps is lower, on average, than of water in savannahs. We can use the landscape, which is more easily described, as a proxy to describe the spatial distribution of this suitable microhabitat. Though this approach is limited in lacking a physiological understanding of direct influences on the pathogen, it has the great benefit of inferring the potential distribution of the pathogen, opening new opportunities to disease control.

We undertook ecological niche modelling of *Mycobacterium ulcerans*, an environmentally acquired pathogenic bacterium, and causative agent of Buruli ulcer. The ecological niche refers to this range of conditions within which a species can survive and maintain a population. We infer that, if a species has a large population, it presumably is able to maintain that population, and is in a suitable environment. By understanding the environmental parameters that describe population size, we can predict the distribution of the pathogen. Maps of the distribution of pathogens are often a key step in control of disease, producing environmental hazard maps.

The pathogen of our study, *Mycobacterium ulcerans*, infects up to 10,000 people per year in more than 30 countries around the world [9,10]. Infection leads to the Buruli ulcer, an emerging neglected tropical disease [10] which results in a necrotizing infection of the skin and can lead to crippling deformity [9]. The transmission route of *M. ulcerans* remains unknown, and though several competing hypotheses exist [11,12] our work herein does not address transmission, but focuses on the distribution of the pathogen.

Identification of the landscape variants that indicate suitable habitat for this particular pathogen has proven remarkably difficult, despite decades of research (see [13] for a review). Previous research on *M. ulcerans* has found several apparently contradictory facts about the bacterium, making it difficult to establish a generalised picture of its ecology. In 2007 the genome of *M. ulcerans* was sequenced, and analysis revealed extensive evidence for reductive evolution, with massive gene loss. *M. ulcerans* evolved from *M. marinum*, and appears to have undergone a bottleneck event in the process, losing many of the genes *M. marinum* uses to sustain itself in free living environments, apparently now favouring protected environments with low sunlight [14]. This is suggestive of a highly specialised ecological niche, implying that the bacterium cannot survive in a large range of environmental conditions. Detection of the bacterium in the environment is normally via PCR; *M. ulcerans* is very slow growing and extremely difficult to culture from the wild [15], and most attempts at culture result in *M. ulcerans* being overgrown by other bacteria which are ubiquitous in the environment.

However, the implication that the microbe is a specialist has been (apparently) contradicted by recent detection of the bacterium in the environment. *M. ulcerans* DNA has been detected in a bewildering variety of environmental samples, including aquatic insects, biofilms, crustaceans, detritus, fish, frogs, possums and various small mammals, soil, snails, water and worms [3,5,15,16,17,18,19,20,21,22,23,24,25,26,27,28,29]. This large range of suitable conditions is odd, in light of the bacterium’s apparent status as a specialist with a small niche.

The many different species that *M. ulcerans* infects in the local community may become infected due to differences in their feeding habits, position in the trophic web, or relative abundance [13,30,31]. Herein, we use samples of the five dominant Orders of the aquatic insect community, which have been tested for *M. ulcerans* positivity rates, and correlate changes in *M. ulcerans* positivity in these 5 Orders to changes in the environmental conditions of land cover and topography. These 5 Orders may not
be the primary habitat of *M. ulcerans* in the wild, as the full biotic extent of *M. ulcerans* distribution is still unknown, but they are commonly found to be persistently infected and appear to be important hosts [32]. Previous work has found that *M. ulcerans* abundance does respond to water body type, being more commonly detected in swamps (still lentic systems) than rivers (flowing lotic systems) in Ghana [33,34]. The pathogen is associated with lowland, flat, swampy areas in contact with stagnant water [35], is known to have complex seasonal dynamics [32], and appears to be present at low levels throughout the entire local biotic community along the year [29]. The distribution of the disease may also inform us on the distribution of the pathogen; the distribution of Buruli ulcer is known to be more spatially restricted than the distribution of *M. ulcerans* [36], and is known to respond to low elevation, forested land cover, and previous rainfall [37,38], which would suggest that perhaps these factors are also important in the distribution of *M. ulcerans*. Taken together, these facts suggested that changes in the biotic distribution of the pathogen could be mapped using landscape variables. Often, sampling of river systems results in the unexpected presence of *M. ulcerans*; if factors at the larger watershed scale add substantial information on the distribution of *M. ulcerans* a description of the upstream region of the river may help to explain this unexpected presence. We describe the condition of the landscape using land cover, such as forest and savannah, and topography, such as elevation and slope. These landscape scale factors are expected to indirectly influence *M. ulcerans* abundance via their influence on the microhabitat the bacterium inhabits, for example affecting the pH, dissolved oxygen content, and composition of the aquatic insect community, which are known to influence *M. ulcerans* distribution [12,29].

To address our questions we describe landscape variables correlated to the presence of the bacterium in aquatic macroinvertebrates in Cameroon, Central Africa. We then test our model against data collected in French Guiana to explore the generalizability of our findings. This will contribute to an understanding of the spatial distribution of this environmental pathogen, and further our ability to control Buruli ulcer disease.

Materials and Methods

A model was constructed on the dataset from Akonolinga, Cameroon, and predicted into French Guiana, South America. This enabled us to describe the niche of *M. ulcerans*, and examine how well these models transferred to other areas.

Study sites, sampling methodology and response variable

The Cameroon dataset is a subset of that published in [29], which comprises 16 sites in Akonolinga, sampled every month for 12 months (Figure 1). Identical methods were carried out by the same investigators for all sites throughout the study. In brief, at each site, 4 locations were chosen in areas of slow water flow and among the dominant aquatic vegetation and at each location, 5 sweeps with a dip net within a surface of 1 m² were done to sample the aquatic community. Aquatic organisms were classified down to the Family level whenever possible and stored separately in 70% ethanol. Individuals belonging to the same taxonomic group were pooled together for detection of *M. ulcerans* DNA by quantitative PCR. Among these, the 5 most abundant Orders (Diptera, Hemiptera, Coleoptera, Odonata and Ephemeroptera) were consistently analysed for all sites and months. Pooled individuals were all ground together and homogenized and DNA from tissue homogenates was purified using QIAquick 96 PCR Purification Kit (QIAGEN). Finally, amplification and detection of MU DNA were performed through quantitative PCR by targeting the ketoreductase B domain (KR) of the mycolactone polyketide synthase and IS2404 sequence from MU genome. This resulted in 5 analyzed samples (each Order) per month, per site, which we use to infer *M. ulcerans* presence or absence. Summary statistics are described in Table 1. Sampling effort varied from month to month, as is discussed in [29], however we have used a subset of that data in order to gain the most consistent representation of the biotic community possible.

A data set following the same methodology was independently collected in French Guiana, South America [28]. DNA extraction was carried out with the same two primer pairs and methodology as above. In French Guiana eighteen sites were sampled twice during the wet season, which lasts from December to July. The entire biotic community was sampled, and for consistency the same 5 taxonomic Orders as in Akonolinga (Table 2) were compared.

Seasonal effects on *M. ulcerans* distribution

*M. ulcerans* has previously been found to respond to variables that are influenced by rainfall [35,38]. To explore differences in the seasonal distribution of the bacterium, the wet season months and the dry season months were analysed separately. In Cameroon wet season months are April, May, June, August, September and October. The dry season is January, February, March, July, November and December. For each site, the proportion of positive samples at a site in a season was determined by summing the number of positive samples in that season, then dividing by the total number of samples sampled in that season (which is 5 multiplied by the number of sampled months). This resulted in two response variables, Ywet and Ydry, which we use to describe the proportion of *M. ulcerans* positive samples in the 5 dominant insect Orders in the wet and dry seasons respectively. This resulted in a general, standardised view of the mycobacterium distribution in both the dry and wet seasons. The habitat suitability is determined by the proportion of samples of the biotic community that are *M. ulcerans* positive.

Land cover and topography

Land cover in Akonolinga was described using several multispectral satellite images; SPOT 2.5 meter resolution images (references: 50833380811220923092V0 and 50833371012210937422V0), and a Landsat image (reference L72186056_05620021107). The study area was categorised into the following classes; Agriculture, Forest, Flood plain, Road, Savannah, Swamp and Urban (Table S1). Classification was conducted in the Object Orientated Image Analysis software eCognition [39]. The resulting maps were validated and corrected where needed following onsite visits in November 2012. Topography was described using the Shuttle Radar Topography Mission (SRTM) digital elevation model [40], which has a spatial resolution of 90 meters. All topographical variables were derived using the Spatial Analyst extension of the software ArcMap 10.1 [41]. For each site we described the mean, standard deviation, minimum, maximum and variety of elevation, in meters above sea level, using SRTM (Table S1). From the SRTM we calculated the mean, standard deviation, minimum, maximum and variety of the topological slope, in degrees. Flow accumulation is the accumulated number of upstream cells flowing into a point, and ecologically represents the topographical potential for water to accumulate. We derived the mean, standard deviation, maximum and variety of the flow accumulation. We also calculated mean, standard deviation, maximum depth, variety, and proportion of buffer surface area covered by basins. Basins are depressions in the
landscape where water is expected to accumulate and, potentially, stagnate, and were detected using the Fill function in Spatial Analyst extension in Arc Map. Stream order indicates the distance from the source of the river, and is a simple index of the type of stream (1st order being small streams, larger orders being big rivers). Proportion of 1st to 8th order streams, defined by Strahler method [42], was recorded in each buffer. Finally, wetness index is the topographic potential for water to accumulate. It was derived from the flow accumulation and the slope, according to the Equation 1, where WI is the wetness index [43], FA is flow accumulation and S is the topographic slope in degrees. We derived the mean, standard deviation, maximum, and variety of wetness index values, and the proportion of buffer surface area covered by wetness index values which are positive (relatively wet areas) and negative (relatively dry areas).

\[ WI = \ln \left( \frac{FA}{\tan(S)} \right) \]  

(1)

Importance of local effects compared to regional effects in M. ulcerans distribution

The topography and land cover of the sample sites were described within two different buffers (Figure 2). These buffers corresponded to local and regional conditions. The first buffer was a 5 km radius circle around the sample site, which was chosen to represent the local conditions. 5 km is, approximately, the flight range of the 5 insect orders sampled [44,45,46,47]. The insects should be able to move throughout this region, be exposed to M. ulcerans, before being captured at the sample site. We describe the land cover and topography within this 5 km buffer and correlate the condition of this region to the proportion of M. ulcerans positive pools in each season.

The second buffer was defined using the watershed of the sample site (Figure 2). The watershed is the upstream catchment area. In principle, all water within this region, and any detritus floating in the water, will eventually flow through the sample site. Watersheds can vary greatly in size, easily being several kilometres long, and detritus from very distant locations can flow quite large distances. M. ulcerans is known to attach to such detritus [24]. This watershed buffer is created using the Watershed tool in ArcMap10.1, Spatial Analyst extension [42].

Principal component analysis

The 42 variables estimated to describe the landscape were reduced to permit modelling. Principal component analysis (PCA) was performed on the landscape variables centred at the mean \((\ln(x) - \ln(x_{mean}))\) to summarize the data in the watershed and the 5 km buffer. PCAs were performed with the PCA function in the FactoMineR library in R [48]. This generated two PCAs; a PCA of the 42 environmental variables in the watershed buffer, PCAws, and a PCA of the 42 environmental variables in the 5 km buffer, PCA5 km. In each PCA we examined the orthogonal axes that explained 95% of the variance in the 42 topography and land cover variables.

Firstly, 9 principal components explained 95% of the variance in the watershed of the sample site (PCAws). The magnitude and direction of each correlation is given in the supplementary materials (Tables S1 and S2). We describe PCAws,1 as “large watersheds that drain flood plains”, given its strongly positive correlations to watershed surface area and floodplains; PCAws,2 as “large watersheds that drain highland agriculture”; PCAws,3 as “large watersheds that drain lowland agriculture”; PCAws,4 as “small watersheds that drain swamp and forest at flat intermediate elevations”; PCAws,5 as “small watersheds that drain highland urban and savannah”; PCAws,6 as “small watersheds that drain highland urban and forest”; PCAws,7 as “large watersheds that...
### Table 1. *M. ulcerans* distribution at sample sites in Akonolinga, Cameroon.

<table>
<thead>
<tr>
<th>Site Code</th>
<th>Latitude</th>
<th>Longitude</th>
<th>Type of water body</th>
<th>Relative abundance of the 5 Orders (%)</th>
<th>PCR positive samples of the 5 Orders,% (positive/samples)</th>
<th>Relative abundance of the 5 Orders (%)</th>
<th>PCR positive samples of the 5 Orders,% (positive/samples)</th>
</tr>
</thead>
<tbody>
<tr>
<td>A1</td>
<td>N3 46.806</td>
<td>E12 16.133</td>
<td>Swamp</td>
<td>93.69</td>
<td>6.98% (3/43)</td>
<td>97.39</td>
<td>20% (4/20)</td>
</tr>
<tr>
<td>A2</td>
<td>N3 47.083</td>
<td>E12 15.383</td>
<td>Swamp</td>
<td>94.69</td>
<td>12.36% (11/89)</td>
<td>94.53</td>
<td>8.11% (3/37)</td>
</tr>
<tr>
<td>A3</td>
<td>N3 46.316</td>
<td>E12 14.440</td>
<td>Stream</td>
<td>90.82</td>
<td>2.56% (1/39)</td>
<td>90.41</td>
<td>0% (0/20)</td>
</tr>
<tr>
<td>A4</td>
<td>N3 58.464</td>
<td>E12 14.796</td>
<td>River</td>
<td>55.12</td>
<td>5.13% (2/39)</td>
<td>32.17</td>
<td>5.26% (1/19)</td>
</tr>
<tr>
<td>A5</td>
<td>N4 02.255</td>
<td>E12 15.620</td>
<td>Stream</td>
<td>83.64</td>
<td>10% (4/40)</td>
<td>75.30</td>
<td>5.26% (1/19)</td>
</tr>
<tr>
<td>A6</td>
<td>N3 43.483</td>
<td>E12 16.466</td>
<td>Swamp</td>
<td>90.31</td>
<td>17.24% (15/87)</td>
<td>85.85</td>
<td>2.94% (1/34)</td>
</tr>
<tr>
<td>A7</td>
<td>N3 38.889</td>
<td>E12 15.986</td>
<td>River</td>
<td>79.52</td>
<td>10.26% (4/39)</td>
<td>79.13</td>
<td>0% (0/13)</td>
</tr>
<tr>
<td>A8</td>
<td>N3 38.980</td>
<td>E12 14.696</td>
<td>River</td>
<td>64.86</td>
<td>9.09% (4/44)</td>
<td>73.89</td>
<td>7.69% (1/13)</td>
</tr>
<tr>
<td>A9</td>
<td>N3 29.912</td>
<td>E12 06.425</td>
<td>Swamp</td>
<td>82.40</td>
<td>4.4% (4/91)</td>
<td>91.67</td>
<td>2.7% (1/37)</td>
</tr>
<tr>
<td>A10</td>
<td>N3 29.912</td>
<td>E12 06.425</td>
<td>Swamp</td>
<td>93.62</td>
<td>7.5% (3/40)</td>
<td>89.90</td>
<td>0% (0/20)</td>
</tr>
<tr>
<td>A11</td>
<td>N3 23.271</td>
<td>E12 07.870</td>
<td>River</td>
<td>50.12</td>
<td>4.65% (4/86)</td>
<td>59.56</td>
<td>8.82% (3/34)</td>
</tr>
<tr>
<td>A12</td>
<td>N3 28.788</td>
<td>E12 07.255</td>
<td>River</td>
<td>52.87</td>
<td>2.44% (1/41)</td>
<td>66.19</td>
<td>15% (3/20)</td>
</tr>
<tr>
<td>A13</td>
<td>N3 32.322</td>
<td>E11 57.643</td>
<td>Swamp</td>
<td>81.92</td>
<td>2.22% (2/90)</td>
<td>79.39</td>
<td>8.11% (3/37)</td>
</tr>
<tr>
<td>A14</td>
<td>N3 38.032</td>
<td>E11 59.695</td>
<td>River</td>
<td>64.77</td>
<td>15.38% (6/39)</td>
<td>63.37</td>
<td>20% (4/20)</td>
</tr>
<tr>
<td>A15</td>
<td>N3 32.288</td>
<td>E11 55.239</td>
<td>Flooded</td>
<td>94.33</td>
<td>11.11% (4/36)</td>
<td>97.54</td>
<td>0% (0/15)</td>
</tr>
<tr>
<td>A16</td>
<td>N3 32.276</td>
<td>E11 55.181</td>
<td>Stream</td>
<td>89.85</td>
<td>2.86% (1/35)</td>
<td>93.40</td>
<td>0% (0/15)</td>
</tr>
</tbody>
</table>

16 sites were sampled for 12 months, sampling from different types of water bodies. The dominant members of the aquatic biota were Diptera, Hemiptera, Coleoptera, Odonata and Ephemeroptera. These made up the majority of the community in both seasons; the percentage of the biotic sampled community composed of these five groups is reported as Relative abundance of the 5 Orders in Table 1. These communities were normally positive of *M. ulcerans*, the percentage of positive samples (number of positive samples/total samples for the 5 Orders from that site in that season) describes the PCR positive samples of the 5 Orders. This table is a summary of a subset of the data presented in [30].

doi:10.1371/journal.pntd.0003298.t001
drain lowland forest, savannah and swamp”; PCAws, as “small watersheds that drain urban and agricultural environments in hilly lowlands”; and PCAws9 as “small watersheds that drain wet swamps in areas that reach from low to high elevations” (Table S1).

Secondly, for the local 5 km circular buffer, 6 principal components (PCA5, km) explained 95% of the variance in the data as described in SM2. Translating these to ecologically meaningful terms, we describe PCA5, km1 as representing “sites surrounded by flat lowland areas with urban, agriculture and the flood plains of large rivers”; PCA5, km2 as representing “sites surrounded by sloped highland areas with urban, agriculture and small rivers”; PCA5, km3 as representing “sites surrounded by sloped highland areas with savannah and large swampy rivers”; PCA5, km4 as representing “sites surrounded by flat lowland areas with small rivers and many small basins in unforest environment”, (Table S2).

Model fitting and evaluation

We allow model selection to choose which of these principal components are most informative in the species distribution, Ywet and Ydry. The dry season general linear models (GLMs) and wet season GLMs were fitted separately with glmulti in the glmulti library in R. Glmulti finds the best set of GLMs among all possible combinations of explanatory variables; so for example all possible Ydry*PCA5, km models were fitted, and each was evaluated with the Akaike information criterion corrected for small sample sizes (AICc). Low AICc scores indicate good performance and reduced overfitting [49]. The best set of these binomial GLMs (within 2 AICc scores of the best model) are selected, and the model within this range with the lowest sum of absolute residuals (best performance) is selected as the final model (Figure S1).

The response variable changed seasonally, resulting in two response variables, Ydry and Ywet. Along with the PCA5, km and PCAws, inputs this resulted in four models; Ydry~PCA5, km and Ywet~PCAws in the dry season, and Ydry~PCA5, km and Ywet~PCAws in the wet season. This reduces our variables by retaining those that are important. Then, to compare the importance of PCA5, km (local) and PCAws (regional watershed) we include in the final models, Ydry~PCA5, km+PCAws in the dry season, and Ywet~PCA5, km+PCAws in the wet season. In this way, by allowing glmulti to retain or drop these variables we can compare the importance of the watershed and local 5 km area variables in the distribution of M. ulcerans.

Potential effects of multicolinearity were explored but were deemed minimal, as all pairwise Pearson correlation coefficient R values in the principal components were below 0.75 (Tables S3 and S4).

In the initial screen of variables, Ydry~PCA5, km and Ydry~PCAws retained PCAws4, “small watersheds that drain swamp and forest at flat intermediate elevations”, PCAws9, “small watersheds that drain wet swamps in areas that reach from low to high elevations”, and PCA5, km2, “sites surrounded by sloped highland areas with urban, agriculture and small rivers”. These were included in the model of interest, Ydry~PCA5, km+PCAws.

For the wet season Ywet~PCA5, km and Ywet~PCAws retained PCAws1, “large watersheds that drain flood plains”, PCAws5, Table 2. M. ulcerans distribution at sample sites in French Guiana, South America.

<table>
<thead>
<tr>
<th>Site Code</th>
<th>Latitude</th>
<th>Longitude</th>
<th>Relative abundance of 5 Orders (%)</th>
<th>PCR positive samples of the 5 Orders,% (positive/samples)</th>
</tr>
</thead>
<tbody>
<tr>
<td>FG10</td>
<td>N4 44.170 W-52 19.618</td>
<td>58.62</td>
<td>44.12% (15/34)</td>
<td></td>
</tr>
<tr>
<td>FG11</td>
<td>N4 50.284 W-52 21.195</td>
<td>52.54</td>
<td>32.26% (10/31)</td>
<td></td>
</tr>
<tr>
<td>FG19</td>
<td>N5 17.773 W-53 03.085</td>
<td>62.50</td>
<td>10.00% (1/10)</td>
<td></td>
</tr>
<tr>
<td>FG2</td>
<td>N5 37.888 W-53 42.433</td>
<td>70.83</td>
<td>11.76% (6/51)</td>
<td></td>
</tr>
<tr>
<td>FG23</td>
<td>N5 21.724 W-53 2.0200</td>
<td>29.27</td>
<td>16.67% (2/12)</td>
<td></td>
</tr>
<tr>
<td>FG28</td>
<td>N5 36.328 W-53 46.960</td>
<td>56.60</td>
<td>20.00% (6/30)</td>
<td></td>
</tr>
<tr>
<td>FG34</td>
<td>N4 50.068 W-52 18.126</td>
<td>85.00</td>
<td>32.35% (11/34)</td>
<td></td>
</tr>
<tr>
<td>FG38</td>
<td>N5 23.646 W-52 59.521</td>
<td>73.74</td>
<td>13.70% (10/73)</td>
<td></td>
</tr>
<tr>
<td>FG41</td>
<td>N5 25.725 W-53 05.326</td>
<td>41.07</td>
<td>8.70% (2/13)</td>
<td></td>
</tr>
<tr>
<td>FG43</td>
<td>N5 22.632 W-52 57.232</td>
<td>75.47</td>
<td>2.50% (1/40)</td>
<td></td>
</tr>
<tr>
<td>FG44</td>
<td>N4 20.052 W-52 09.148</td>
<td>25.42</td>
<td>0.00% (0/15)</td>
<td></td>
</tr>
<tr>
<td>FG45</td>
<td>N4 18.025 W-52 07.397</td>
<td>61.95</td>
<td>2.86% (2/70)</td>
<td></td>
</tr>
<tr>
<td>FG46</td>
<td>N5 02.121 W-52 30.989</td>
<td>74.24</td>
<td>2.04% (1/49)</td>
<td></td>
</tr>
<tr>
<td>FG47</td>
<td>N4 55.744 W-52 24.229</td>
<td>65.00</td>
<td>0.00% (0/26)</td>
<td></td>
</tr>
<tr>
<td>FG48</td>
<td>N4 51.616 W-52 16.518</td>
<td>16.67</td>
<td>0.00% (0/1)</td>
<td></td>
</tr>
<tr>
<td>FG49</td>
<td>N5 39.996 W-52 46.794</td>
<td>36.54</td>
<td>0.00% (0/19)</td>
<td></td>
</tr>
<tr>
<td>FG53</td>
<td>N5 36.136 W-52 50.182</td>
<td>67.86</td>
<td>0.00% (0/57)</td>
<td></td>
</tr>
<tr>
<td>FG7</td>
<td>N4 51.648 W-52 15.405</td>
<td>29.41</td>
<td>0.00% (0/10)</td>
<td></td>
</tr>
</tbody>
</table>

18 sites were sampled in the wet season. The dominant members of the aquatic biota were Diptera, Hemiptera, Coleoptera, Odonata and Ephemeroptera, as in Akonolinga. These made up the majority of the community, the percentage of the biotic sampled community composed of these five groups is reported as Relative abundance of the 5 Orders. These communities were normally positive of M. ulcerans, the percentage of positive samples (number of positive samples/total samples for the 5 Orders from that site in that season) describes the PCR positive samples of the 5 Orders. This table is a summary of a subset of the data presented in [28].

doi:10.1371/journal.pntd.0003298.t002
Watersheds and *M. ulcerans*
“small watersheds that drain highland urban and savannah”, PCA5km 6, “small watersheds that drain highland urban and forest”, PCAws 8, “small watersheds that drain urban and agricultural environments in hilly lowlands”, PCA5km2, “sites surrounded by sloped highland areas with urban, agriculture and small rivers” and PCA5km4, “sites surrounded by flat lowland areas with savannah and small rivers”, which were included in Y_{wet} = \text{PCA5km} + \text{PCAws}.

Predicting the spatial distribution of suitable habitat for \textit{M. ulcerans} in the model training region, Akonolinga

We interpolate the Akonolinga model within the region of Akonolinga to predict the distribution of suitable habitat, the reservoir, of \textit{M. ulcerans}. To achieve this, points where streams (defined using STRM) flow under or across roads (defined using satellite images) were selected. These were termed ‘pour points’ in this article. Selection of the point where streams cross roads was based on the hypothesis that these environments, where contact between humans and the aquatic environment will be high, may be important in infection. This does not mean that infection does not occur in other locations, nor do we speculate on the importance of relative routes of transmission. This will not characterise all the environmental reservoir of the bacterium, but will describe an important part of it. The topography and land cover of the watershed and 5 km buffer of these pour points was characterised, transformed into PCA5km and PCAws format, and the GLM was predicted. As a summary to describe this distribution, we use Moran’s Index of spatial autocorrelation, which describes the extent to which the distribution is random, and is here used to describe the distribution of suitable sites. This is implemented using the tool Spatial Autocorrelation Global Moran’s I in ArcMap10.1 [41].

Y_{wet} \sim 1 + \text{PCAws}9 + \text{PCA5km}2

The final GLM suggested that both local and regional effects are substantially correlated to \textit{M. ulcerans} distribution. Regional effects were represented by PCAws9, “small watersheds that drain wet swamps in areas that reach from low to high elevations”, and was negatively correlated to \textit{M. ulcerans} abundance (correlation coefficient $-0.37$, $p = 0.007$). This means we expect less \textit{M. ulcerans} in small watersheds that drain swamps near highlands.

Relative importance of local and regional effects on the distribution of \textit{M. ulcerans} in wet season

The final fitted wet season Binomial logit GLM, after stepwise AICc selection, was

Results

Relative importance of local and regional effects on the distribution of \textit{M. ulcerans} in dry season

The final fitted dry season Binomial logit GLM, after stepwise AICc selection, is

Y_{dry} \sim 1 + \text{PCAws}1 + \text{PCA5km}2 + \text{PCA5km}4

The final models on the dry season found that both regional and local effects were substantially correlated to presence of \textit{M. ulcerans}. Regional effects were represented by PCAws1, “large watersheds that drain flood plains”, which was marginally negatively correlated to \textit{M. ulcerans} abundance (correlation coefficient $-0.26$, $p = 0.05210$). PCA5km2, “sites surrounded by areas with urban, agriculture and small rivers” was positively correlated to \textit{M. ulcerans} abundance (correlation coefficient 0.09, $p = 0.15709$) though the $p$ value suggests this is not significant, and finally PCA5km4, “sites surrounded by areas with savannah and small rivers”, was positively correlated to \textit{M. ulcerans} abundance, (correlation coefficient 0.38, $p = 0.007$). The spatial distribution of \textit{M. ulcerans} suitable habitat in the dry season predicted at the pour points is non-random, based on...
Moran’s I spatial autocorrelation (Moran’s Index: 0.33, z-score: 14.32, p = 0.00001) positive sites tend to cluster together (Figure 3).

Model performance when interpolated in Akonolinga

Spatial autocorrelation of model residuals can be an issue in GLMs, but this was explored, and it was not the case here. Model residuals were not significantly spatially autocorrelated in the wet season (Moran’s Index: 0.28, z-score: 1.04, p = 0.29) nor in the dry season (Moran’s Index: 0.07, z-score: 0.65, p = 0.51).

The AICc of the final dry season Binomial model was 49.6, the absolute sum of the residuals was 11.0. The AICc of the final wet season Binomial model was 67.8, the absolute sum of the residuals was 11.95.

We note that Gaussian models had significantly better performance. The AICc of the final dry season Gaussian model
Figure 4. Model validation in French Guiana. Sample sites were as in [28]. A wet season Gaussian niche model based on data collected in Cameroon was predicted into French Guiana (3rd row, left hand side). The model under-predicted, *M. ulcerans* was present in more sites than expected (bottom row, model residuals). A similar Binomial model predicted all sites to be negative.

doi:10.1371/journal.pntd.0003298.g004

was −39.8, the absolute sum of the residuals was 0.53. The AICc of the final wet season Gaussian model was −65.5, the absolute sum of the residuals was 0.24. Model performance is presented in Figure S2, model residuals were normally distributed (Figure S3).

Model performance when extrapolated in French Guiana

The Akonolinga wet season model was predicted into 18 sample sites in French Guiana (Figure 4, 2nd row). The model predicted sites to be positive or negative, and the results of qPCR corroborated these predictions (Figure 4). Performance of the Binomial model was notably poor, all sites were predicted negative. In contrast, performance of the Gaussian model was better, but accuracy was still poor at 0.39 (Table S3). Sensitivity and negative predictive values are high, indicating that the predictions of presence of the bacterium are likely to be true, specificity and positive predictive values are low; indicating predictions of absence of the bacterium are likely to be incorrect. This is a result of a bias towards Type II errors (false negatives) in the Gaussian model. Overall, the model predicts *M. ulcerans* in Akonolinga, but is sensitive to extrapolation. Extrapolation tends to result in false negative predictions of presence.

Discussion

Here, we have demonstrated that in addition to local variables around the sample site, the distribution of *M. ulcerans* correlates to regional variables, i.e. the topography and land cover of the watershed of the sample site. This spatial distribution of suitable habitat was described, allowing the production of environmental hazard maps for the distribution of the pathogen. *M. ulcerans* presence in the wet season correlates with lowland areas surrounded by few agricultural or urban areas, particularly if the sample site has a large watershed. We expect more *M. ulcerans* in the dry season in sites surrounded by urban and agricultural areas, with many small streams, particularly if the sample site has a small watershed.

Many of the findings are in accord with what little we already understand about this bacterium. *M. ulcerans* has been previously associated with flat wetland areas [9,35]. A similar association with Buruli ulcer has been reported [51], which found that high standard deviation of the wetness index was a risk factor for Buruli ulcer. These three variables are normally strongly correlated to each other and ecologically similar entities. In this study these are negatively correlated to PCA9, here termed “small watersheds that drain wet swamps in areas that reach from low to high elevations” which negatively correlated to *M. ulcerans* abundance: these studies appear to be describing the same ecological entity, but with different variables.

Our study was limited in certain regards, as we focused it on the prevalence of *M. ulcerans* in the biotic community, and on how topography and land cover in the region could influence that prevalence. We do not consider abiotic conditions testing positive for *M. ulcerans*. Potentially the abiotic distribution may respond differently to these variables, future work will aim to explore this. However, given that *M. ulcerans* is commonly detected in the biotic environment and appears to be at lower prevalence in the abiotic environment, we believe our results are still applicable to an understanding of *M. ulcerans* distribution. We had a relatively low positivity rate (Table 1). A potential limitation is that low positivity can bias a model towards false negatives, while this is possible we are unable to test this further with our current data.

The Akonolinga wet season model was extrapolated into French Guiana, where sampling was in the wet season. Despite good performance in Akonolinga, the model performed poorly in French Guiana, under-predicting the bacterium’s distribution (Figure 4). There are a number of points to be drawn from this. First, there were differences in sampling effort between the two sites, as the Akonolinga sampling regime consisted of 12 time points in the year, while the French Guiana regime consisted of 2 time points. This would be consistent with the idea that the bacterium is transiently present in different regions, and under-prediction would be expected in this case. Secondly, a potential complication results from differences in the ability of the SRTM database to delineate watersheds due to dense rainforest canopies in French Guiana [52]. The shape of a watershed is sensitive to the quality of the elevation data used, errors in the digital elevation model, or man-made drainage structures, can have effects not captured by this model. Finally, we cannot rule out that the differences are a result of differences in *M. ulcerans*. We used qPCR to detect *M. ulcerans*, however the species is known to have multiple ecovars [53,54] and subspecies, distributed differently throughout the globe. If it is the case that we are predicting the ecological niche of one *M. ulcerans* ecovar species into French Guiana, and testing it against a separate French Guiana species, one would expect the model to under-predict if the French Guiana subspecies occupies a larger ecological niche.

Regardless of error structure, selection of both types of models (Gaussian and Binomial) retained watersheds as important variables. These findings will impact future research on Buruli ulcer and *M. ulcerans*; future sampling regimes would benefit by consideration of the local hydrology before beginning sampling, and selecting sample sites along these lines. We also postulate the importance of watersheds as a barrier to dispersal for the bacterium. A recent key study found a strong relationship between *M. ulcerans* population structure and the greater West African hydrological watersheds [53], with populations being bound to watersheds. These are the drainage areas of large rivers such as the Nyong, Mbam and Ouenmé rivers, a much larger scale than our study. However, given our results herein, it seems the bacteria may drift downstream. This is inferred by the difference in the effect of watershed size from dry to wet seasons.

This is consistent with the idea of a ‘flushing’ effect of rainfall in the wet season, carrying bacteria downstream [38], which will influence their genetic population structure. This has notable consequences for the epidemiology of Buruli ulcer. If the watersheds are barriers to movement for the bacteria it implies that *M. ulcerans* may be common in the environment, but in certain areas hydrological conditions facilitate concentration of the bacterium, as is the case with anthrax [55].

Conclusion

The distribution of environmental pathogens needs to be understood to facilitate control. Commonly, local effects in the microhabitats are considered to describe the ecological niche of a pathogen. However our study demonstrates that regional effects are important factors to be considered. Future research on the *M. ulcerans* would benefit by considering the watershed of potential sample sites, particularly as such data is often quite simple to acquire. The shape, size, and land cover of the watershed correlates with changes in the distribution of *M. ulcerans*, and useful information is lost if watersheds are ignored. The distribution of

Watersheds and *M. ulcerans*
swamp in a watershed was found to be an important factor in the suitability of the site for \textit{M. ulcerans}; though a sample point in the field may be at a location normally considered unsuitable for the bacteria (e.g. a small swift lentic stream), the area upstream may contain an abundance of lotic swamps and be quite suitable for the bacterium, which may be ‘washed out’ downstream towards the sample site. This is an example of the useful information we gain by placing pathogens in an environmental context, rather than regarding them solely in an epidemiological sense.

**Supporting Information**

**Figure S1** GLMulti output, for binomial and Gaussian models. Sum of absolute model residuals are plotted against AICc. Within the region of 2 AICc scores of the best model (vertical lines) we select the model with the lowest residuals (highlighted in red).

**(DOC)**

**Figure S2** Observed against predicted values for each model. Note that Gaussian models have a much better fit.

**(DOC)**

**Figure S3** Quantile-quantile plots of normality. The Gaussian and Binomial are both similarly normally distributed, though the Binomial displays a larger variance of residuals.

**(DOC)**

**Table S1** Results of principle component analysis for topographical and land cover variables in a watershed buffer. 95% of the variance in the data was described with 9 components, the eigenvalue of each component is given at the bottom of the table. Each component correlates differently to different variables, red highlights negative correlations, blue highlights positive correlations. PCA5km1 describes large watersheds that drain flat lowlands and swamps, with few urban and agricultural areas. These are high elevation areas with variable slopes. PCA5km2 describes large watersheds that drain flat lowlands and swamps, with few urban and agricultural areas. These are high elevation areas with variable slopes. PCA5km3 represents sites surrounded by sloped highland areas and urban, agriculture and the flood plains of large rivers. PCA5km4 represents sites surrounded by sloped highland areas with savannah, and large swampy rivers. PCA5km5 represents sites surrounded by flat lowland areas with urban and agriculture, and large rivers. PCA5km6 represents sites surrounded by lowland hills, with small rivers and many small basins, in un forested environment. PCA5km7 is larger watersheds that drain forest, savannah flood plain and swamp, in areas with flat, wet, lowlands. PCA5km8 represents small watersheds that drain urban & agriculture, flood plain and savannah. These are wet lowlands with lots of small hills. PCA5km9 represents small watersheds that drain wet swamps in areas that reach from low to high elevations.

**(DOC)**

**Table S2** Results of principle component analysis for topographical and land cover variables in a 5 km buffer around the sample site. 95% of the variance in the data was described with 6 components. Each component correlates differently to different variables, red highlights negative correlations, blue highlights positive correlations. Surface area is constant, at \( \pi s^2 = 79 \text{ km}^2 \). PCA5km1 represents sites surrounded by flat lowland areas and urban, agriculture and the flood plains of large rivers. PCA5km2 represents sites surrounded by sloped highland areas and urban, agriculture, and small rivers. PCA5km3 represents sites surrounded by sloped highland areas with savannah, and large swampy rivers. PCA5km4 represents sites surrounded by flat lowland areas with savannah and small rivers. PCA5km5 represents sites surrounded by flat highlands with urban and agriculture, and large rivers. PCA5km6 represents sites surrounded by lowland hills, with small rivers and many small basins, in un forested environment.

**(DOC)**

**Table S3** Pearson product R correlation coefficients in the wet season model. Stepwise selection selected 3 components, none of which were correlated.

**(DOC)**

**Table S4** Pearson product R correlation coefficients in the dry season model. Stepwise selection selected 6 components, none of which were correlated.

**(DOC)**

**Table S5** Contingency table describing model performance of niche models constructed in Cameroon and predicted into French Guiana. The rows ‘Prediction’ are model predictions, ‘Test’ are the results from qPCR of the sites in French Guiana. Values in blue are true positives and true negatives; values in red are false positives and false negatives.

**(DOC)**

**Acknowledgments**

We are grateful to the staff of the Centre Pasteur and IRD for their invaluable help in different phases of the study, notably during data collection. We also thank Amelie Tran of CIRAD and Benjamin Roche of IRD for invaluable discussions and insights on previous versions of the manuscript, the ISIS Spot programme for support in acquiring SPOT images, and Hervé Chevillotte (IRD Cameroon), for environmental data from the IFORA project (ANR-Biodiv grant IFORA).

**Author Contributions**

Conceived and designed the experiments: KC JFG DLS. Performed the experiments: KC AG AM. Analyzed the data: KC GEGP. Contributed reagents/materials/analysis tools: LM JL SE PLG GT AF. Wrote the manuscript, the ISIS Spot programme for support in acquiring SPOT images, and Hervé Chevillotte (IRD Cameroon), for environmental data from the IFORA project (ANR-Biodiv grant IFORA).