



HAL
open science

Resource use by three juvenile scarids (*Cryptotomus roseus*, *Scarus iseri*, *Sparisoma radians*) in Caribbean seagrass beds

Charlotte R. Dromard, Amandine Vaslet, Françoise Gautier, Yolande Bouchon,
Mireille Harmelin-Vivien, Claude Bouchon

► **To cite this version:**

Charlotte R. Dromard, Amandine Vaslet, Françoise Gautier, Yolande Bouchon, Mireille Harmelin-Vivien, et al.. Resource use by three juvenile scarids (*Cryptotomus roseus*, *Scarus iseri*, *Sparisoma radians*) in Caribbean seagrass beds. *Aquatic Botany*, 2017, 136, pp.1-8. <10.1016/j.aquabot.2016.08.003>. <hal-01375546>

HAL Id: hal-01375546

<https://hal.sorbonne-universite.fr/hal-01375546v1>

Submitted on 3 Oct 2016

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



HAL Authorization

14 **ABSTRACT**

15 The bucktooth parrotfish *Sparisoma radians*, the striped parrotfish *Scarus iseri* and the
16 bluelip parrotfish *Cryptotomus roseus* are three herbivorous fishes commonly found at
17 juvenile stages in Caribbean seagrass beds. While the diet of the three species as adults is
18 relatively well known, few studies have been conducted on the feeding patterns of juveniles.
19 In this study, the resource use of the juveniles of three scarid species were studied using two
20 complementary methods: gut content and stable isotope analyses (^{13}C : ^{12}C and ^{15}N : ^{14}N ratios).
21 Bayesian mixing model approaches were used to calculate the contribution of each food item
22 to fish diets (SIAR, mixing models). The three parrotfish species appeared to rely essentially
23 on the consumption of fleshy macrophytes. *Cryptotomus roseus* consumed more benthic
24 invertebrates and presented a higher trophic level than the two other scarid species. *Scarus*
25 *iseri* presented a higher assimilation of benthic biofilm, in accordance with the high
26 percentage of sediment in its gut content, and *Sparisoma radians* assimilated more *Thalassia*
27 *testudinum* leaves. This research highlighted a food resources partitioning among the
28 juveniles of the three herbivorous fishes, probably to avoid inter-specific competitive
29 interactions for the most palatable food at a critical stage of their life.

30

31 *Keywords:* Gut content analyses, trophic niche, stable isotopes analyses.

32 1. Introduction

33 Herbivorous fishes have been widely studied on coral reefs due to their major role in the
34 control of algae and are known to mitigate their competitive interactions with corals
35 (Burkepile and Hay, 2010 and references therein). On tropical seagrass beds, herbivorous
36 fishes, such as parrotfishes (Scarinae), have received less attention (Del Moral *et al.*, 2016). In
37 the Caribbean, three species of herbivorous fishes are commonly found on seagrass beds: the
38 bucktooth parrotfish *Sparisoma radians* (Valenciennes, 1840), the bluelip parrotfish
39 *Cryptotomus roseus* (Cope, 1870) and the striped parrotfish *Scarus iseri* (Bloch, 1789). The
40 first two species are exclusively found on seagrass beds, whereas *Scarus iseri* can also be
41 found on coral reefs at all life stages. During their juvenile stages, these three species cohabit
42 in shallow seagrass meadows which represent their primary nursery habitat (Bouchon-Navaro
43 *et al.*, 2004; Kopp *et al.*, 2010; Layman and Silliman, 2002; Weinstein and Heck, 1979).
44 Indeed, studying seagrass meadows is particularly relevant due to their role as nursery areas
45 for nearshore fishes (Tuya *et al.*, 2014).

46 Possibly due to anthropic pressures on seagrass meadows, the stocks of these three juveniles
47 have dramatically decreased over the last ten years in Guadeloupe, although they were
48 formerly common in this habitat (Y. Bouchon-Navaro, Pers. com. 2016).

49 Information on the diet of these herbivorous fishes in the literature is principally available for
50 adults. *Sparisoma radians* mainly ingests turtlegrass *Thalassia testudinum* Banks & Sol. ex
51 König, 1805, fleshy macroalgae like *Dictyota* sp. or *Acanthophora spicifera* (M. Vahl)
52 Børgesen, 1910 and calcified macroalgae such as *Halimeda* sp. (Lobel and Ogden, 1981;
53 Randall, 1967; Targett and Targett, 1990). Depending on the region, *Scarus iseri* is described
54 as a feeder on microalgae from dead coral pavements, fleshy macroalgae or epiphytic
55 filamentous microalgae (McAfee and Morgan, 1996; Nagelkerken *et al.*, 2006; Randall,
56 1967). *Cryptotomus roseus* is described as strictly herbivorous, feeding on seagrass

57 (Carpenter, 2002), but its diet has not been studied in detail. However, little information is
58 available for the juveniles of these three parrotfish species, although ontogenetic shifts in diet
59 are common for tropical fishes (Cocheret de la Morinière *et al.*, 2003) or temperate
60 herbivorous fishes (Havelanche *et al.*, 1997). Dietary changes are useful to understand their
61 ecological role in seagrass beds, such as the regulation of the algal biomass.

62 Trophic niches of fishes can be described on the basis of gut content analyses (Ogden, 1976;
63 Randall, 1967) or direct observation of their feeding behaviour in the field by the counting of
64 “bites” (Cardoso *et al.*, 2009; Lobel and Ogden, 1981; McAfee and Morgan, 1996;
65 Overholtzer and Motta, 1999). However, these methods provide a description of a species’
66 diet at a specific time and present several practical problems (Bearhop *et al.*, 2004). With gut
67 content analyses, the principal difficulty results from the ability of herbivorous fishes, such as
68 Scarinae, to grind the ingested matter into small fragments with their fused beak and their
69 pharyngeal mills (Bellwood and Choat, 1990; Randall, 1967). The relative proportions of
70 food items in gut contents are estimated with varying degrees of accuracy by different authors
71 and on the basis of different parameters (occurrence, weight or volume percentages). With
72 field observations, it is difficult to discriminate between potential dietary targets (e.g. between
73 seagrass leaves and epiphytes) and the ingestion of some food items, such as detritus, is
74 difficult to quantify. For these reasons, determining the trophic niches of herbivorous fishes
75 remains challenging.

76 More recently, stable isotope ratios of fish muscles ($^{13}\text{C}:^{12}\text{C}$ and $^{15}\text{N}:^{14}\text{N}$) have been used to
77 determine their trophic niche. Isotopic ratios measured in consumer tissues are closely linked
78 to those of their diet and increase in a stepwise fashion with each trophic level.

79 The mean trophic isotopic enrichment, or fractionation factor ($\Delta^{15}\text{N}$ and $\Delta^{13}\text{C}$), was estimated
80 at $+ 3.4 \pm 1.1\text{‰}$ for nitrogen (Minagawa and Wada, 1984) and $+ 0.4 \pm 1.3\text{‰}$ for carbon (Post,
81 2002). However, Mill *et al.* (2007) demonstrated a higher $\Delta^{15}\text{N}$ for herbivorous fishes (4.7‰

82 for *Acanthurus sohal* and 4.1‰ for *Sparisoma* spp.) and Sweeting *et al.* (2007) recommended
83 a $\Delta^{13}\text{C}$ ranging between 1.5‰ and 2‰ for marine fishes.

84 Stable isotope analysis is considered to be a powerful tool to reflect the feeding behaviour of
85 individuals over long periods (approximately three months in the muscles of adult fishes) for
86 fish muscles in adults), corresponding to the turnover of the tissues of consumers (Bearhop *et*
87 *al.*, 2004).

88 Several methods using C and N stable isotopes ratios, expressed as $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$, have been
89 developed to understand and interpret fish ecology. Isotopic signatures of the consumer and
90 those of its potential food sources can be used in mixing models to calculate and estimate the
91 contribution of several food sources to the diet of a consumer (see Phillips, 2012 and Phillips
92 *et al.*, 2015 for reviews). Results of mixing models can be used to describe and compare
93 trophic niches between fish species (Dromard *et al.*, 2015; Plass-Johnson *et al.*, 2013).
94 Despite the fact that numerous potential organic matter sources occur in seagrass beds, mixing
95 models have been used before for fish species living in this habitat (Benstead *et al.*, 2006;
96 Hyndes and Lavery, 2005; Loneragan *et al.*, 1997). However, few studies have been done in
97 the Caribbean seagrass beds (Nagelkerken and van der Velde, 2004).

98 In the present study, we analysed the diet of juveniles of three parrotfish species in a
99 Caribbean seagrass bed to describe and compare their trophic niche, combining stable isotope
100 and gut content analyses.

101

102 **2. Materials and methods**

103 The study was carried out in the Bay of the Grand Cul-de-Sac Marin, located in the northern
104 part of Guadeloupe Island (Lesser Antilles) (Fig. 1). In the south, the bay is bordered by
105 mangroves and the northern part is partially enclosed by a coral barrier reef. Shoals of the bay
106 are colonized by seagrass meadows dominated by the turtlegrass, *Thalassia testudinum*.

107 Sampling was performed in a shallow seagrass bed (< 2 m depth), at mid-distance between
108 the coast and the barrier reef, far from the influence of mangroves, at the end of the wet
109 season (October to December 2010). The sampling area covers 1 km² approximately.

110

111 *2.1. Sampling protocol*

112 Three herbivorous fish species (*C. roseus*, *S. iseri* and *S. radians*) were sampled during 6
113 purse seine samples (Table 1). The total length of fish was measured to the nearest millimetre
114 and individuals were weighed to the nearest milligram (Table 2). All specimens sampled were
115 below their minimum size at first maturity (Bouchon-Navaro *et al.*, 2006). The main potential
116 organic matter (OM) sources in the seagrass beds were sampled and treated for stable isotope
117 analysis (Table 1). The dominant species or genera of fleshy macroalgae occurring in seagrass
118 beds were collected and cleaned with distilled water: *Anadyomene stellate* (Wulfen) C.
119 Agardh, 1823, *Caulerpa cupressoides* (Vahl) C. Agardh, 1817, *Chaetomorpha* sp., *Dictyota*
120 *cf pulchella* Hörnig & Schnetter, 1988 and *Padina* sp., along with the calcified macroalgae
121 *Halimeda incrassata* (J. Ellis) J.V. Lamouroux, 1816 and *Udotea flabellum* (J. Ellis &
122 Solander) M.A. Howe, 1904. Two seagrasses were collected, *Thalassia testudinum* and
123 *Syringodium filiforme* Kützinger, 1860. Samples of *T. testudinum* were sorted into two
124 categories: old leaves (O) and young leaves (Y), both selected without epiphytes. Around
125 100g (wet weight) of each species of macroalgae and seagrass were collected on field.
126 Epiphytes colonising old leaves of *Thalassia* were gently scrapped with a scalpel blade and
127 stored apart. Benthic invertebrates (amphipods, copepods, decapods, gastropods) collected
128 with seagrass and macroalgae samples were sorted. When collected, macroalgae and
129 *Thalassia* leaves were preserved in plastic bags in order to retain the detritus, composed of
130 organic matter and bacteria (Crossman *et al.*, 2001), deposited on algal thalli and seagrass
131 leaves. Plastic bags were then opened in plastic boxes, macroalgae and *Thalassia* were gently

132 shacked on water and detritus settled were collected at the bottom of the boxes. Due to the
133 small amounts, invertebrates were pooled together to constitute another type of food source.
134 Gastropods were acidified to remove their hard shells before analyses. Particulate organic
135 matter (POM) present in the water column above seagrasses was collected by filtering
136 subsurface water on preweighed Whatman GF/F filters precombusted for 4 hours at 450°C.
137 The first centimetre of surface sediment was collected to analyse the biofilm mainly
138 composed of benthic diatoms, bacteria, detritus and settling POM (Belicka *et al.*, 2012). To
139 ensure a sufficient quantity of material for the analyses, around 10g (wet weight) of epiphytes,
140 invertebrates, detritus, POM and biofilm were sampled.

141

142 2.2. Gut content analyses

143 Ten individuals of each species had full stomach and were studied for gut content analyses.
144 Two methods were used according to the fish species. For *C. roseus*, food categories were
145 sorted, oven dried and weighed separately as described by Hyslop (1980), because it was
146 possible to isolate the different food items. Diets of the two other fish species were studied
147 using the method of point-intercepts, described by Jones (1968). Food categories were
148 expressed as percentages of the total dry-weight for *C. roseus* and as percentages of point-
149 intercepts for *S. iseri* and *S. radians*.

150

151 2.3. Stable isotope analysis

152 Dorsal fish muscles cut into small pieces and all seagrass carbon sources were oven dried at
153 50°C to a constant weight and ground into a homogenous fine powder. Carbon and nitrogen
154 stable isotope ratios were performed on two subsamples for food sources that may contain
155 carbonates: POM, biofilm, detritus, epiphytes, invertebrates and calcified macroalgae. For
156 $\delta^{13}\text{C}$, a subsample was acidified with 1N HCl to remove calcified material that presents a less

157 negative $\delta^{13}\text{C}$ than organic material (De Niro and Epstein, 1978). For $\delta^{15}\text{N}$, a non-acidified
158 subsample was used, because acidification can distort $\delta^{15}\text{N}$ values (Pinnegar and Polunin,
159 1999). For each sample (and each subsample), 1mg of dry weight was used for analysis.
160 Nitrogen and carbon isotope ratios were determined by a continuous flow mass spectrometer
161 (Thermo Fisher™, delta V Advantage). Elemental concentrations of carbon and nitrogen
162 ([C]% and [N]%) were measured with an elementary analyser (Thermo Fisher™, Flash EA
163 1112). Isotopic ratios were expressed in standard delta notation [δ values (‰)] according to
164 the following formula: $\delta_X = [(R_{\text{sample}}/R_{\text{standard}} - 1)] \times 1000$, where X is ^{13}C or ^{15}N and R the
165 ratio $^{15}\text{N}:^{14}\text{N}$ or $^{13}\text{C}:^{12}\text{C}$ of samples or international standards (Vienna Pee Dee belemnite
166 limestone carbonate for carbon and atmospheric air for nitrogen). The measurement precision
167 was $<0.1\%$ for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$. Percentages of organic C and organic N were measured on
168 acidified sub-samples for those treated by HCl, using the elemental analyser and were used to
169 calculate the sample C/N ratio. Low C/N ratio indicated higher nutritional value and
170 digestibility of food sources.

171

172 2.4. Mixing models

173 The Bayesian stable isotope mixing model SIAR v4.0 (Stable Isotope Analysis in R)
174 developed by Parnell *et al.* (2010) was used to estimate the proportional contribution of food
175 sources to the diet of the three herbivorous fishes. As there were too many sources of carbon
176 in seagrass beds, we reduced their number to eight. Some were not taken into account in
177 mixing models, as they were not or hardly consumed by the juvenile scarids: MOP,
178 calcareous macroalgae (*Halimeda incrassata* and *Udotea flabellum*) and the seagrass
179 *Syringodium filiforme*. Even if some juvenile scarids scraped calcified *Halimeda* (Overholzer
180 and Motta, 1999), they probably mainly ingest epiphytes. Other prey were combined when
181 presenting not significantly different C and N stable isotopic values. Epiphytes, detritus and

182 *Dictyota* were considered as one food source, as were old and young leaves of *Thalassia*
183 combined. Three models were run according to each fish species. In each model, we entered
184 the individual isotopic values of fish muscles and the mean carbon and nitrogen signatures (\pm
185 SD) of the eight potential food sources. Mixing models took into account carbon and nitrogen
186 fractionation factors ($\Delta^{13}\text{C}$ and $\Delta^{15}\text{N}$). In this study, we fixed mean enrichments (\pm SD) of 1.5
187 $\pm 0.5\text{‰}$ for carbon and $3.9 \pm 0.5\text{‰}$ for nitrogen as the mean difference between $\delta^{15}\text{N}$ values
188 of all sources taken into account and juvenile scarids. This $\Delta^{15}\text{N}$ is slightly lower than those
189 given for adult herbivorous fishes by Mill *et al.* (2007) (4.3 to 5.7‰), and higher than those
190 indicated by Wyatt *et al.* (2010) (1.7 to 2.5‰). Mean contributions (Bayesian Credibility
191 Interval 95%) of food sources to fish diets were expressed as percentages.

192

193 2.5. Statistical analyses

194 Data were tested for normality with Shapiro tests. When the data were normal, nitrogen and
195 carbon signatures among fish muscles and among food sources were compared using one-way
196 analyses of variance (ANOVA), followed by *post hoc* Tukey tests. Otherwise a non-
197 parametric Kruskal-Wallis analysis of variance was used. Hierarchical clustering based on
198 normalized Euclidean distance and Ward's criterion was performed on mean isotopic ratios to
199 identify groups of species with similar C and N isotopic ratios. All tests were performed using
200 the program R.

201

202 3. Results

203

204 3.1. Gut content analyses

205 Due to the difficulty of accurately identifying all fragments of seaweeds and seagrasses in gut
206 contents, it was not possible to calculate the proportion of each plant species ingested by the

207 three fish species. Gut contents were sorted into four categories: sediment, vegetal matter
208 (including microalgae, turfing algae, macroalgae and seagrass), benthic invertebrates and
209 unidentified organic matter (Table 3). Benthic invertebrates (amphipods, isopods and
210 decapods) were found only in the gut contents of *Cryptotomus roseus*. *Scarus iseri* and
211 *Sparisoma radians* presented similar diets with a dominance of vegetal matter (Table 3).

212

213 3.2. Stable isotope analyses

214 Nitrogen and carbon signatures of fish muscles and food sources are represented as a bi-plot
215 (Fig. 2, Table 1). *Cryptotomus roseus*, *Scarus iseri* and *Sparisoma radians* presented the most
216 enriched nitrogen isotopic signatures ($\delta^{15}\text{N} \pm \text{SD}$), with values of $7.0 \pm 0.7\text{‰}$, $5.7 \pm 0.8\text{‰}$ and
217 $5.7 \pm 0.6\text{‰}$, respectively. $\delta^{15}\text{N}$ of fish muscles were significantly different between *C. roseus*
218 and the two other fish species ($F = 12.30$, $p < 0.001$), while *S. radians* differed significantly
219 from *C. roseus* and *S. iseri* by its higher $\delta^{13}\text{C}$ value ($-9.6 \pm 0.8\text{‰}$ vs $-11.5 \pm 1.6\text{‰}$ and $-11.9 \pm$
220 0.9‰ respectively; $F = 14.01$, $p < 0.001$). Nitrogen signatures of OM sources ranged from 5.1
221 $\pm 0.2\text{‰}$ for invertebrates to $1.2 \pm 0.1\text{‰}$ for *Udotea flabellum* and their carbon signatures
222 ($\delta^{13}\text{C} \pm \text{SD}$) varied between $-17.2 \pm 0.7\text{‰}$ for *Chaetomorpha* sp. and $-4.7 \pm 0.1\text{‰}$ for
223 *Syringodium filiforme* (Fig. 2, Table 1). Significant differences in C and N isotopic signatures
224 were found among OM sources ($\delta^{13}\text{C}$: $X^2 = 69.7$, $p < 0.001$; $\delta^{15}\text{N}$: $X^2 = 58.3$, $p < 0.001$).
225 However, no significant difference was observed between old and young leaves of *Thalassia*,
226 and between epiphytes, *Dictyota* and detritus ($p > 0.05$ for both groups). They were thus
227 pooled together into two food categories for mixing models.

228

229 3.3. Mixing models

230 Three mixing models were performed, i.e. one per fish species. Eight types of food sources
231 were introduced into each model: invertebrates, biofilm, *Thalassia testudinum* (O and Y

232 leaves combined), the group epiphytes-detritus-*Dictyota* and four macroalgae likely to be
233 consumed by juvenile scarids (*Anadyomene stellata*, *Caulerpa cupressoides*, *Chaetomorpha*
234 sp. and *Padina* sp.). The results of the mixing models were largely undetermined, with a
235 broad overlap in the 95% confidence intervals, due to the number of sources and the wide
236 range of isotopic values in the macroalgae (Table 4). However, macroalgae constituted the
237 largest part of the food assimilated by juvenile scarids (mean contribution: 42%, 53% and
238 57% in *C. roseus*, *S. iseri* and *S. radians* respectively), followed by the group epiphytes-
239 detritus-*Dictyota* between 10% and 15% (Fig. 3). The three other food sources discriminated
240 better the three fish species. *C. roseus* assimilated a higher proportion of invertebrates (mean
241 contribution: 21%) than the two other species (*Scarus iseri* S.i: 4% and *Sparisoma radians*
242 S.r.: 5%), *S. iseri* a higher proportion of biofilm (mean contribution: 22% vs *Cryptotomus*
243 *roseus* C.r.:15% and S.r.: 10%) and *S. radians* a higher proportion of *Thalassia testudinum*
244 (mean contribution: 17% vs C.r.: 8% and S.i.: 11%) (Fig. 3, Table 4). While seagrass,
245 macrophytes and epiphytes presented high C/N ratios (10 – 38), invertebrates, biofilm and
246 detritus had lower C/N values (6 - 9), testifying their higher nutritional quality (Table 1).
247 These high quality food items contributed around 50% of the food assimilated by *C. roseus*,
248 36% by *S. iseri* and 25% by *S. radians*, epiphytes and *Dictyota* being included.

249

250 **4. Discussion**

251 Coupling gut contents and stable isotope analyses allows the description of fish diets with
252 greater accuracy. These two complementary methods have been used before in *Thalassia*
253 seagrass beds to study fish feeding behaviours (Harrigan *et al.*, 1989; Vaslet *et al.*, 2015), but
254 few such studies have been conducted on fish juveniles (Lugendo *et al.*, 2006; Nagelkerken
255 and van der Velde, 2004) or on herbivorous fishes (Cocheret de la Morinière *et al.*, 2003).

256

257 4.1. Gut content and isotopic mixing model problems

258 Observations of gut contents showed that the three species principally ingested vegetal matter
259 (including micro- and macroalgae, seagrass and their epiphytes). Gut content analyses give
260 interesting results but present several limitations, as previously discussed, especially for
261 Scarinae which grind their food. As these three parrotfish species differed significantly in
262 their stable isotopic values, this suggested that they might have different feeding patterns.
263 Results of stable isotope mixing models (SIAR) showed that juveniles of these three species
264 indeed assimilated mainly macroalgae, but presented nevertheless some food partitioning,
265 which was in accordance with the results of gut content analysis. *C. roseus* presented a more
266 enriched $\delta^{15}\text{N}$ than the two other species and assimilated more benthic invertebrates, such as
267 small crustaceans, in accordance with gut content analysis. *S. iseri* had a more depleted $\delta^{13}\text{C}$
268 and assimilated a higher proportion of biofilm, which might explain the higher proportion of
269 sediment in the gut content of this species. *S. radians* presented an enriched $\delta^{13}\text{C}$ and a higher
270 assimilation of *Thalassia testudinum* leaves. However, the results of the mixing models were
271 largely undetermined, with a wide range of mean contributions (95% CI, Table 4) and should
272 be taken with caution. This lack of resolution was due, firstly, to the wide range of food
273 sources taken into account in the mixing models, even in selecting the most susceptible ones
274 to be eaten by juvenile scarids and in combining those with close isotopic values; and
275 secondly to the wide range of $\delta^{13}\text{C}$ in macroalgae and their rather close $\delta^{15}\text{N}$ (cf Table 1).
276 *Chaetomorpha* presented a highly negative $\delta^{13}\text{C}$ value close to the biofilm, while *Caulerpa*
277 and *Padina* exhibited enriched $\delta^{13}\text{C}$ values close to those of *Thalassia* leaves. Grouping these
278 food categories together to reduce the number of sources used in the mixing models would
279 have resulted in more determined models, but would not have had any ecological meaning.
280 Juvenile parrotfishes consume a wide range of food resources on Caribbean coral reefs
281 (Overholtzer and Motta, 1999) and seagrass beds (this study). Taking into account too many

282 food resources leads to inconclusive mixing model results (cf. Brett, 2014; Fry, 2013 and
283 Philipps *et al.*, 2015, for critical reviews on stable isotope mixing models). Thus the selection
284 of a reduced number of food sources improve the results of mixing models but may lead to an
285 oversimplification of fish diets that would be inconsistent with field observations. Combining
286 gut content and stable isotope analyses with DNA analysis of the food ingested, field
287 observations and feeding trial experiments may be necessary to fully resolve the problem of
288 parrotfish feeding.

289

290 4.2. Interspecific differences

291 Few data are available on the diet of *Cryptotomus roseus*, which is assumed to feed mostly on
292 *Thalassia testudinum* (Carpenter, 2002). However, the morphology of its teeth (separated
293 teeth) could explain its greater consumption of invertebrates compared to the other species.
294 The use of benthic invertebrates by young herbivorous fishes could also result from a relic of
295 their previous carnivorous stage (early post-settlement stage). Ontogenetic shifts in diet from
296 carnivorous to herbivorous have been described before for Scarinae in the Pacific region
297 (Bellwood, 1988; Chen, 2002).

298 Adult individuals of *Scarus iseri* are described as feeders on microalgae and consumers of
299 algal turf, especially when found on coral reefs (Dromard *et al.*, 2015; McAfee and Morgan,
300 1996; Nagelkerken *et al.*, 2006; Randall, 1967; van Rooij *et al.*, 1996), on macroalgae
301 (Cardoso *et al.*, 2009; Randall, 1967; Wolf, 1985), or epiphytes from seagrass leaves
302 (Nagelkerken *et al.*, 2006). In Guadeloupean seagrass beds, *S. iseri* juveniles were
303 characterized by a higher assimilation of biofilm compared to the two other fish species. The
304 shift of diet between juvenile and adult stages could result from the ontogenetic change of
305 habitat, from seagrass beds to coral reefs. Algal turf, constituted by filamentous macroalgae
306 heavily grazed by adults, was not found in the studied seagrass bed because of the lack of

307 dead pavement support. The greater quantity of sediment in *S. iseri* gut content could be due
308 to the consumption of biofilm developing at sediment surface.

309 Previous studies showed that adult *Sparisoma radians* consume between 88 and 95% of
310 *Thalassia testudinum*, including epiphytes from leaves (Lobel and Ogden, 1981; Montague *et*
311 *al.*, 1995; Randall, 1967), because this fish is undeterred by chemical and physical defences of
312 *T. testudinum* (Goecker *et al.*, 2005; Targett and Targett, 1990). In the present study, the
313 proportion of *T. testudinum* assimilated by juveniles of *S. radians* was estimated to be 17%,
314 most of the vegetal matter ingested coming from macroalgae. *S. radians* seems thus to exhibit
315 an ontogenetic diet shift, possibly due to a gradual physiological adaptation allowing *S.*
316 *radians* to consume and digest *T. testudinum*. Several authors have shown the importance of
317 epiphytes of *Thalassia* leaves in the diet of *S. radians* and *S. iseri* (Cocheret de la Morinière *et*
318 *al.*, 2003; Montague *et al.*, 1995; Nagelkerken *et al.*, 2006). In the present study, the
319 respective assimilation of detritus, epiphytes and *Dictyota* could not be estimated because of
320 their close isotopic values. The high proportion of detritus in the diets of herbivorous fishes
321 has been underlined by several authors (e.g. Crossman *et al.* 2001; Dromard *et al.*, 2013) and
322 related to their high nutritional quality estimated by their low C/N ratio, as also observed in
323 this study. The proportion of food sources with low C/N ratios (invertebrates, biofilm, detritus
324 and epiphytes) differed among the three juvenile parrotfishes, in the following decreasing
325 order, *C. roseus* (50%), *S. iseri* (36%) and *S. radians* (25%). These differences suggested that
326 they differed in their ability to consume, process and assimilate plant material. Further studies
327 focussed on the physiological processes involved in nutrient extraction and utilisation in
328 herbivorous fishes, as highlighted by Clements *et al.* (2009), should be conducted in order to
329 fully understand interspecific differences in scarid diets.

330 Thus, the juveniles of the three scarid species seem to present a high overlap in their use of
331 food resources, with high consumption of macroalgae, but nevertheless displayed different

332 preferences for specific food sources: invertebrates for *C. roseus*, biofilm for *S. iseri* and
333 *Thalassia* leaves for *S. radians*. These results indicated some early food partitioning among
334 juvenile scarids of different species, differences which would increase with individual growth.

335

336 **5. Conclusions**

337 The lack of dietary redundancy observed for adult scarids on Indian Ocean coral reefs (Plass-
338 Johnson *et al.*, 2013) seems to be also observed for juvenile scarid species in Caribbean
339 seagrass beds. At a juvenile stage, the three fish species live together in seagrass meadows,
340 their primary nursery habitat, forming multispecific shoals. The present study indicated that,
341 while macroalgae formed a large part of their diets, some differences in the use of food
342 resource occur between these juveniles, in spite of the lack of robustness of mixing model
343 results. The difference of diet between the three herbivorous scarid species could contribute to
344 their coexistence in seagrass beds without inter-specific competitive interactions during their
345 juvenile stage, as observed among other juvenile parrotfishes on coral reefs in Florida Keys
346 (Overholtzer and Motta, 1999). The comparison between juveniles (present study) and adults
347 (from the literature) also suggested that ontogenetic dietary changes occurred in these three
348 scarid fish species, resulting in higher trophic niche differentiation in adults.

349

350 **Acknowledgements**

351 The authors express their thanks to P. Richard and G. Guillou, Université de la Rochelle -
352 CNRS UMR LIENSs, for performing stable isotope analyses. We thank S. Cordonnier,
353 Université des Antilles, for his help in field sampling, and Michael Paul for improvement of
354 the English text. We are grateful to K. Clements, University of Auckland, for his useful
355 comments on an earlier draft of the manuscript.

356

357 **References**

- 358 Bearhop, S., Adams, C.E., Waldron, S., Fuller, R.A., Macleod, H., 2004. Determining trophic
359 niche width: A novel approach using stable isotope analysis. *J Animal Ecol*, 73, 1007-
360 1012.
- 361 Belicka, L.L., Burkholder, D., Fourqurean, J.W., 2012. Stable isotope and fatty acid
362 biomarkers of seagrass, epiphytic, and algal organic matter to consumers in a pristine
363 seagrass ecosystem. *Mar Freshwater Res*, 63, 1085-1097.
- 364 Bellwood, D.R., 1988. Ontogenetic changes in the diet of early post-settlement *Scarus* species
365 (Pisces: Scaridae). *J Fish Biol*, 33, 213-219.
- 366 Bellwood, D.R., Choat, J.H., 1990. A functional analysis of grazing in parrotfishes (family
367 Scaridae): the ecological implications. *Environ Biol Fishes*, 28, 189-214.
- 368 Benstead, J.P., March, J.G., Fry, B., Ewel, K.C., Pringle, C.M., 2006. Testing IsoSource:
369 stable isotope analysis of a tropical fishery with diverse organic matter sources.
370 *Ecology*, 87, 326-333.
- 371 Bouchon-Navaro, Y., Bouchon, C., Louis, M., 2004. L'ichtyofaune des herbiers de
372 phanérogames marines des Antilles françaises : intérêt de leur protection. *Revue*
373 *d'Ecology*, 59, 253-272.
- 374 Bouchon- Navaro, Y., Bouchon, C., Kopp, D., Louis, M., 2006. Weight-length relationships
375 for 50 fish species collected in seagrass beds of the Lesser Antilles. *J Appl Ichthyol*,
376 22, 322-324.
- 377 Brett, M.T., 2014. Resource polygon geometry predicts Bayesian stable isotope mixing model
378 bias. *Mar Ecol Prog Ser*, 514, 1-12.
- 379 Burkepile, D.E., Hay, M.E., 2010. Impact of herbivore identity on algal succession and coral
380 growth on a Caribbean Reef. *PLoS ONE*, 5, e8963.

- 381 Cardoso, S.C., Soares, M.C., Oxenford, H.A., Côté, I.M., 2009. Interspecific differences in
382 foraging behaviour and functional role of Caribbean parrotfish. *Mar Biodivers Rec*, 2,
383 e148.
- 384 Carpenter, K.E., 2002. Scaridae. In: Carpenter, K. (ed.), *The living marine resources of the*
385 *Western central Atlantic, Volume 2: Bony fishes Part 2 (Opistognathidae to Molidae),*
386 *sea turtles and marine mammals*. Rome: FAO, pp. 1723-1739.
- 387 Chen, L.S., 2002. Post-settlement diet of *Chlorurus sordidus* and *Scarus schlegeli* (Pisces:
388 Scaridae). *Zool Stud*, 41, 47-58.
- 389 Clements, K.D., Raubenheimer, D., Choat, J.H., 2009. Nutritional ecology of marine
390 herbivorous fishes: ten years on. *Funct Ecol*, 23, 79-92.
- 391 Cocheret de la Moriniere, E., Pollux, B., Nagelkerken, I., Hemminga, M., Huiskes, A., van
392 der Velde, G., 2003. Ontogenetic dietary changes of coral reef fishes in the mangrove-
393 seagrass-reef continuum: stable isotopes and gut-content analysis. *Mar Ecol Prog Ser*,
394 246, 279-289.
- 395 Crossman, D.J., Choat, J.H., Clements, K.D., Hardy, T., McConochie, J., 2001. Detritus as
396 food for grazing fishes on coral reefs. *Limnol Oceanogr*, 46, 1596-1605.
- 397 Del Moral, L., Vidal, J., Betancor, S., Tuya, F., 2016. Differences in herbivory intensity
398 between seagrass *Cymodocea nodosa* and the green alga *Caulerpa prolifera* inhabiting
399 the same habitat. *Aquat Bot*, 128, 48-57.
- 400 De Niro, M.J., Epstein, S., 1978. Influence of diet on the distribution of carbon isotope ratios
401 in animals. *Geochim Cosmochim Acta*, 42, 495-506.
- 402 Dromard, R.C., Bouchon-Navaro, Y., Cordonnier, S., Fontaine, M.F., Verlaque, M.,
403 Harmelin-Vivien, M., Bouchon, C., 2013. Resource use of two damselfishes, *Stegastes*
404 *planifrons* and *Stegastes adustus*, on Guadeloupean reefs (Lesser Antilles): Inference
405 from stomach content and stable isotopes analysis. *J Exp Mar Biol Ecol*, 440, 116-125.

- 406 Dromard, R.C., Bouchon-Navaro, Y., Harmelin-Vivien, M., Bouchon, C., 2015. Diversity of
407 trophic niches among herbivorous fishes on a Caribbean reef (Guadeloupe, Lesser
408 Antilles), evidenced by stable isotope and gut content analyses. *J Sea Res*, 95, 124-
409 131.
- 410 Fry, B., 2013. Alternative approaches for solving undetermined isotope mixing problems.
411 *Mar Ecol Prog Ser*, 472, 1-13.
- 412 Goecker, M., Heck, Jr K., Valentine, J., 2005. Effects of nitrogen concentrations in turtlegrass
413 *Thalassia testudinum* on consumption by the bucktooth parrotfish *Sparisoma radians*.
414 *Mar Ecol Prog Ser*, 286, 239-248.
- 415 Harrigan, P., Zieman, J.C., Macko, S.A., 1989. The base of nutritional support for the gray
416 snapper (*Lutjanus griseus*): An evaluation based on a combined stomach content and
417 stable isotopes analysis. *Bull Mar Sci*, 44, 65-77.
- 418 Havelanche, S., Lepoint, G., Dauby, P., Bouquegneau, J.M., 1997. Feeding of the sparid fish
419 *Sarpa salpa* in a seagrass ecosystem: diet and carbon flux. *Mar Ecol*, 18, 289-297.
- 420 Hyndes, G.A., Lavery, P.S., 2005. Does transported seagrass provide an important trophic
421 link in unvegetated, nearshore areas? *Estuar Coast Shelf S*, 63: 633-643.
- 422 Hyslop, E.J., 1980. Stomach contents analysis - A review of methods and their application. *J*
423 *Fish Biol*, 17, 411-429.
- 424 Jones, R.S., 1968. A suggested method for quantifying gut contents in herbivorous fishes.
425 *Micronesica*, 4, 369-371.
- 426 Kopp, D., Bouchon-Navaro, Y., Louis, M., Legendre, P., Bouchon, C., 2010. Spatial and
427 temporal variation in a Caribbean herbivorous fish assemblage. *J Coast Res*, 28, 63-
428 72.
- 429 Layman, C.A., Silliman, B.R., 2002. Preliminary survey and diet analysis of juvenile fishes of
430 an estuarine creek on Andros Island, Bahamas. *Bull Marine Sci*, 70, 199-210.

- 431 Lobel, P., Ogden, J., 1981. Foraging by the herbivorous parrotfish *Sparisoma radians*. *Mar*
432 *Biol*, 64, 173-183.
- 433 Loneragan, N., Bunn, S., Kellaway, D., 1997. Are mangroves and seagrasses sources of
434 organic carbon for penaeid prawns in a tropical Australian estuary? A multiple stable-
435 isotope study. *Mar Biol*, 130, 289-300.
- 436 Lugendo, B.R., Nagelkerken, I., van der Velde, G., Mgya, Y.D., 2006. The importance of
437 mangroves, mud and sand flats, and seagrass beds as feeding areas for juvenile fishes
438 in Cwaka bay, Zanzibar: gut content and stable isotope analyses. *J Fish Biol*, 69,
439 1639-1661.
- 440 McAfee, S., Morgan, S., 1996. Resource use by five sympatric parrotfishes in the San Blas
441 Archipelago, Panama. *Mar Biol*, 125, 427-437.
- 442 Mill, A., Pinnegar, J., Polunin, N., 2007. Explaining isotope trophic- step fractionation: why
443 herbivorous fish are different. *Funct Ecol*, 21, 1137-1145.
- 444 Minagawa, M., Wada, E., 1984. Stepwise enrichment of ^{15}N along food chains: Further
445 evidence and the relation between $\delta^{15}\text{N}$ and animal age. *Geochim Cosmochim Acta*,
446 48, 1135-1140.
- 447 Montague, J.R., Carballo, J.L., Valdes, L.M., Chacken, M., 1995. Analyses of decay and
448 parrotfish grazing along attached leaves of turtle grass (*Thalassia testudinum*) from
449 two sites in Biscayne Bay. *Florida Scientist*, 58, 206–215.
- 450 Nagelkerken, I., van der Velde, G., 2004. Relative importance of interlinked mangroves and
451 seagrass beds as feeding habitats for juvenile reef fish on a Caribbean island. *Mar Ecol*
452 *Prog Ser*, 274, 153-159.
- 453 Nagelkerken, I., van der Velde, G., Verberk, W.C., Dorenbosch, M., 2006. Segregation along
454 multiple resource axes in a tropical seagrass fish community. *Mar Ecol Prog Ser*, 308,
455 79-89.

- 456 Ogden, J.C., 1976. Some aspects of herbivore-plant relationships on Caribbean reefs and
457 seagrass beds. *Aquat Bot*, 2, 103-116.
- 458 Overholtzer, K.L., Motta, P.J., 1999. Comparative resource use by juvenile parrotfishes in the
459 Florida Keys. *Mar Ecol Prog Ser*, 177, 177-187.
- 460 Parnell, A.C., Inger, R., Bearhop, S., Jackson, A.L., 2010. Source partitioning using stable
461 isotopes: coping with too much variation. *PLoS ONE*, 5, e9672.
- 462 Phillips, D.L., 2012. Converting isotope values to diet composition: the use of mixing models.
463 *J Mammal*, 93, 342-352.
- 464 Phillips, D.L., Inger, R., Bearhop, S., Jackson, A.L., Moore J.W., Parnell, A.C., Semmens,
465 B.X., Ward E.J. 2015. Best practices for use of stable isotope mixing models in food-
466 web studies. *Can J Zool*, 92, 823-835.
- 467 Pinnegar, J.K., Polunin, N.V.C., 1999. Differential fractionation of $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ among fish
468 tissues: implications for the study of trophic interactions. *Funct Ecol*, 13, 225-231.
- 469 Plass-Johnson, J.G., McQuaid, C.D., Hill, J.M., 2013. Stable isotope analysis indicates a lack
470 of inter- and intra-specific dietary redundancy among ecologically important coral reef
471 fishes. *Coral Reefs*, 32, 429-440.
- 472 Post, D.M., 2002. Using stable isotopes to estimate trophic position: models, methods, and
473 assumptions. *Ecology*, 83, 703-718.
- 474 Randall, J.E., 1967. Food habits of reef fishes of the West Indies. *Studies in Tropical*
475 *Oceanography*, 5, 665-847.
- 476 Sweeting, C., Barry, J., Barnes, C., Polunin, N., Jennings, S., 2007. Effects of body size and
477 environment on diet-tissue $\delta^{15}\text{N}$ fractionation in fishes. *J Exp Mar Biol Ecol*, 340, 1-
478 10.

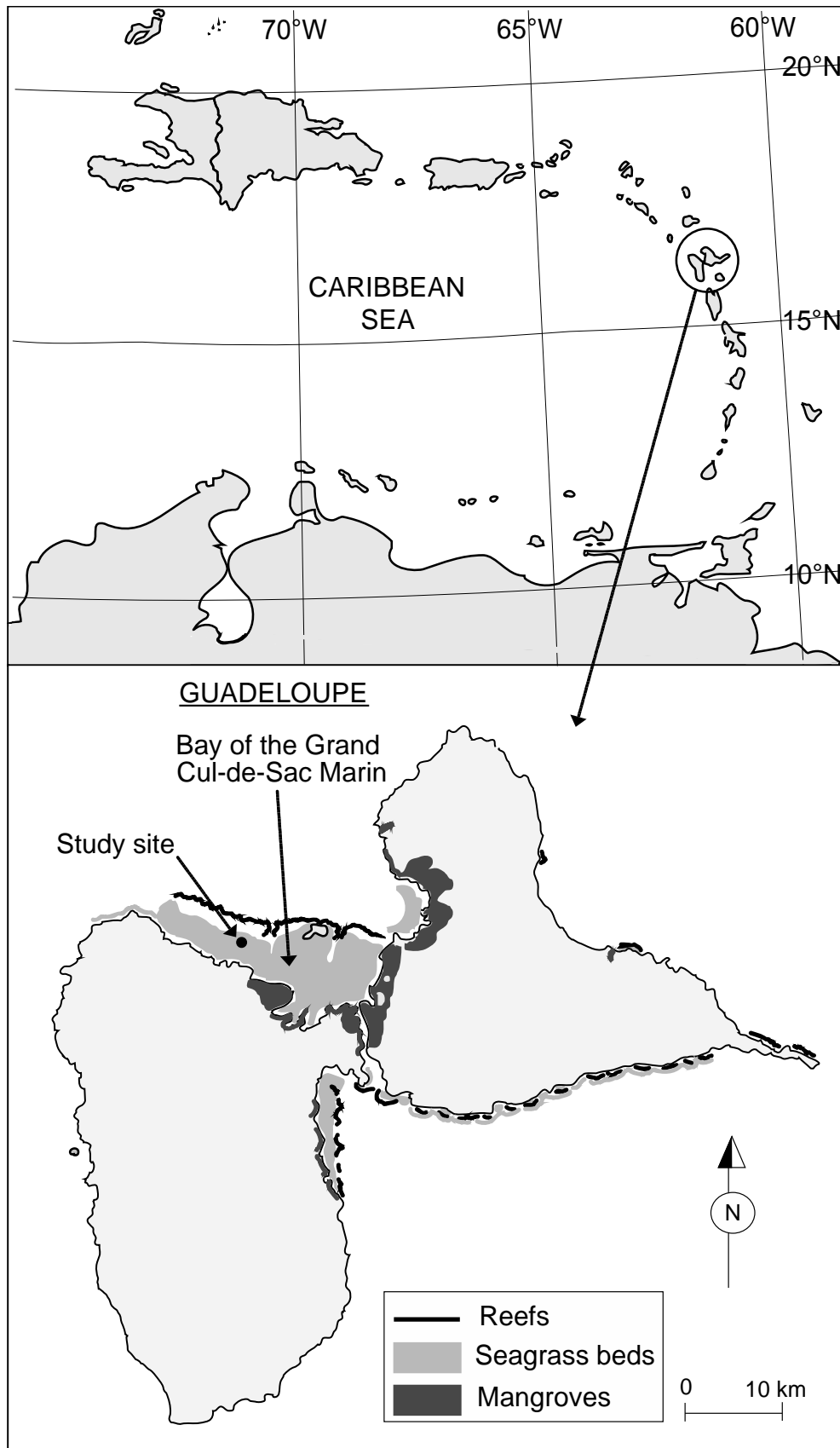
- 479 Targett, T.E., Targett, N.M., 1990. Energetics of food selection by the herbivorous parrotfish
480 *Sparisoma radians*: roles of assimilation efficiency, gut evacuation rate, and algal
481 secondary metabolites. *Mar Ecol Prog Ser*, 66, 13-21.
- 482 Tuya, F., Haroun, R., Espino, F., 2014. Economic assessment of ecosystem services:
483 Monetary value of seagrass meadows for coastal fisheries. *Ocean Coast Manage*, 96,
484 181-187.
- 485 van Rooij, J.M., Jong, E., Vaandrager, F., Videler, J.J., 1996. Resource and habitat sharing by
486 the stoplight parrotfish, *Sparisoma viride*, a Caribbean reef herbivore. *Environ Biol*
487 *Fishes*, 47, 81-91.
- 488 Vaslet, A., Bouchon-Navaro, Y., Harmelin-Vivien, M.L., Lepoint, G., Louis, M., Bouchon,
489 C., 2015. Foraging habits of reef fishes associated with mangroves and seagrass beds
490 in a Caribbean lagoon: a stable isotope approach. *Cienc Mar*, 41 (3), 217-232
- 491 Weinstein, M., Heck, Jr K., 1979. Ichthyofauna of seagrass meadows along the Caribbean
492 coast of Panama and in the Gulf of Mexico: composition, structure and community
493 ecology. *Mar Biol*, 50, 97-107.
- 494 Wolf, N., 1985. Food selection and resources partitioning by herbivorous fishes in mixed-
495 species groups. *Proceedings of the 5th International Coral Reef Congress* (Tahiti), 4,
496 23-28.
- 497 Wyatt, A.S.J., Waite, A.M., Humphries, S., 2010. Variability in isotope discrimination factors
498 in coral reef fishes: implications for diet and food web reconstruction. *PLoS ONE*, 5,
499 e13682.

500

501 **List of figures**

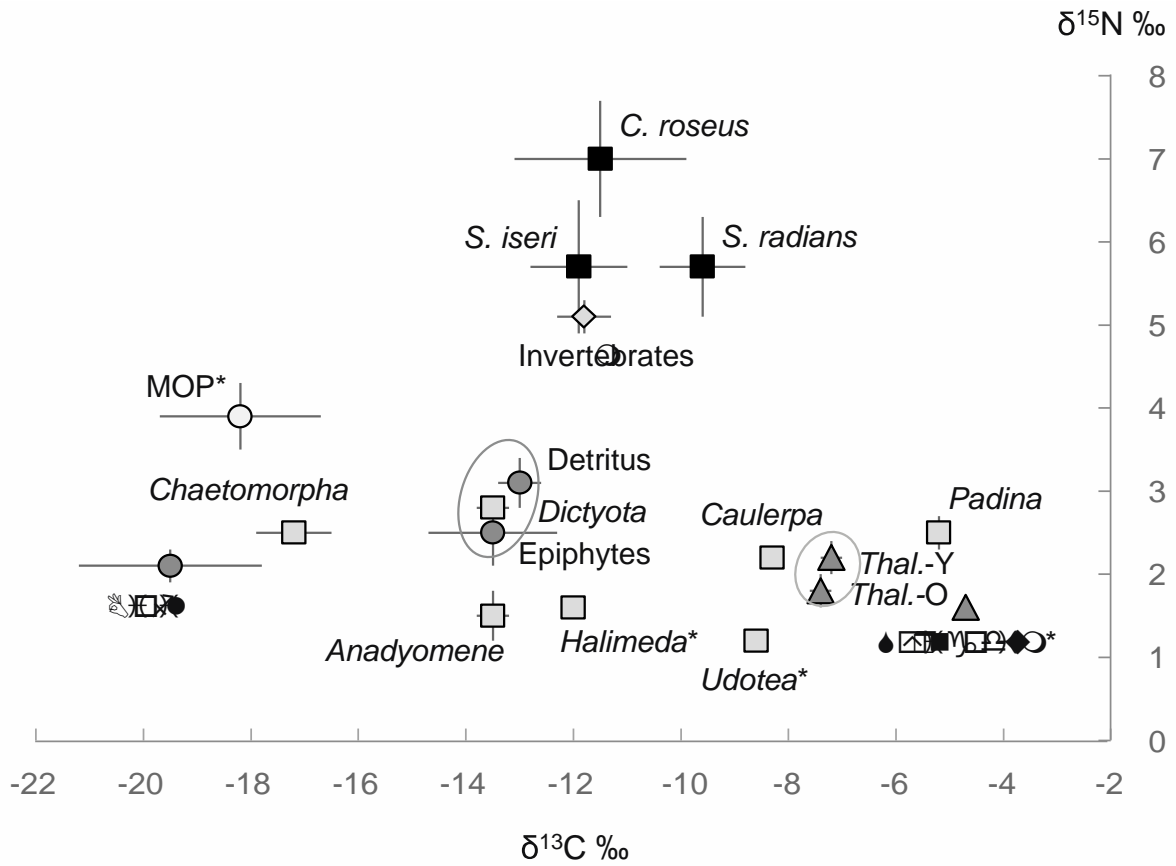
502

503 Figure 1. Location of the study site in the Bay of the Grand Cul-de-Sac Marin in Guadeloupe
504 (Lesser Antilles).

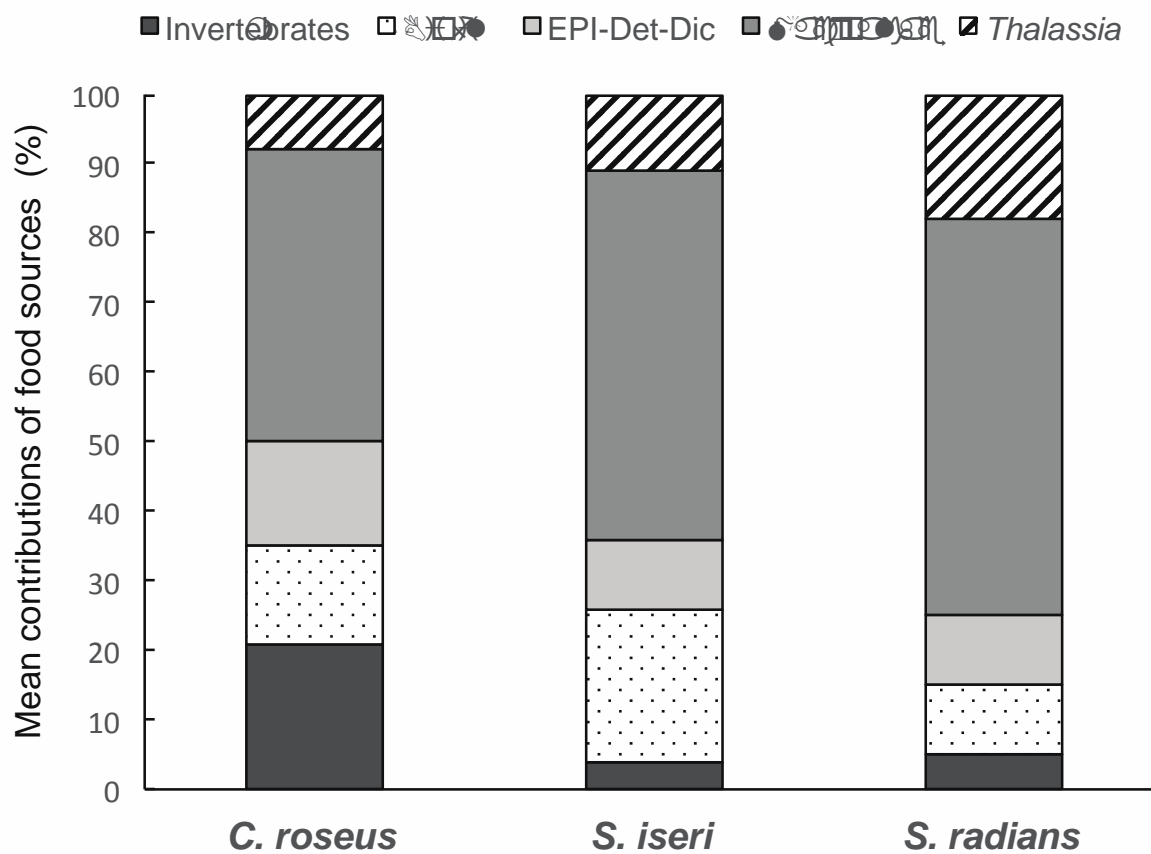


505

506 Figure 2. Mean isotopic signatures of carbon ($\delta^{13}\text{C} \pm \text{SD}$) and nitrogen ($\delta^{15}\text{N} \pm \text{SD}$) of the
 507 three juvenile fish species and their potential food sources in Guadeloupean seagrass beds.
 508 *Thal.-O*: *Thalassia*-old leaves; *Thal.-Y*: *Thalassia*- young leaves; * indicates sources not
 509 taken into account in the mixing models.



510
 511 Figure 3. Mean contribution of food sources to the diet of fishes calculated with SIAR mixing
 512 models. 95% Bayesian Credibility Intervals of these values are given in Table 3. Fleshy
 513 macroalgae: *Anadyomene stellata*, *Caulerpa cupressoides*, *Chaetomorpha* sp. and *Padina* sp.
 514 EPI-Det-Dic: epiphytes, detritus and *Dictyota* sp.



515

516 **Tables**

517

518 Table 1. Mean \pm SD values of $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ signatures and C/N ratio of juvenile scarids and

519 organic matter sources collected in Guadeloupe seagrass beds. N: number of samples. *

520 Sources not taken into account in mixing models

Sample types	N	$\delta^{13}\text{C}$ ‰ \pm SD	$\delta^{15}\text{N}$ ‰ \pm SD	C/N
Fish species				
<i>Cryptotomus roseus</i>	13	-11.5 \pm 1.6	7.0 \pm 0.7	3.3 \pm 0.1
<i>Scarus iseri</i>	11	-11.9 \pm 0.9	5.7 \pm 0.8	3.3 \pm 0.03
<i>Sparisoma radians</i>	13	-9.6 \pm 0.8	5.7 \pm 0.6	3.1 \pm 0.02
Invertebrates	5	-11.8 \pm 0.5	5.1 \pm 0.2	5.6 \pm 0.5
MOP*	9	-18.2 \pm 1.5	3.9 \pm 0.4	7.1 \pm 1.2
Biofilm	6	-19.5 \pm 1.7	2.1 \pm 0.2	9.4 \pm 0.2
Detritus	5	-13.0 \pm 0.04	3.3 \pm 0.03	8.6 \pm 0.1
Fleshy macroalgae				
<i>Anadyomene stellata</i>	3	-13.5 \pm 0.3	1.5 \pm 0.3	20.0 \pm 0.4
<i>Caulerpa cupressoides</i>	3	-8.3 \pm 0.1	2.2 \pm 0.03	13.5 \pm 0.2
<i>Chaetomorpha</i> sp.	3	-17.2 \pm 0.7	2.5 \pm 0.03	19.5 \pm 0.7
<i>Dictyota</i> cf <i>pulchella</i>	3	-13.5 \pm 0.3	2.8 \pm 0.1	23.8 \pm 0.5
<i>Halimeda incrassata</i> *	3	-12.0 \pm 0.04	1.6 \pm 0.04	9.9 \pm 0.1
<i>Padina</i> sp.	3	-5.2 \pm 0.1	2.5 \pm 0.2	38.0 \pm 0.9
<i>Udotea flabellum</i> *	3	-8.6 \pm 0.03	1.2 \pm 0.1	10.8 \pm 0.1
Epiphytes (<i>Thalassia</i>)	6	-13.5 \pm 1.2	2.5 \pm 0.4	11.8 \pm 2.9
Seagrass				
<i>Thalassia testudinum</i> (old)	10	-7.4 \pm 0.1	1.8 \pm 0.2	23.8 \pm 2.1
<i>Thalassia testudinum</i> (young)	5	-7.2 \pm 0.1	2.2 \pm 0.1	19.7 \pm 0.2
<i>Syringodium filiforme</i> *	5	-4.7 \pm 0.1	1.6 \pm 0.07	23.0 \pm 0.3

521

522

523

524

525 Table 2. Mean and range of fish total length (TL) and wet weight (W) of the three fish
 526 species. Lengths at maturity (Lm) are taken from Bouchon-Navaro *et al.* (2006).

527

Fish species	TL (cm)	W (g)	Lm (cm)
<i>Cryptotomus roseus</i>	5.7 (4.6 – 6.8)	2.5 (1.3 – 4.2)	8.6
<i>Scarus iseri</i>	5.0 (4.5 – 5.5)	2.0 (1.5 – 2.6)	15.9
<i>Sparisoma radians</i>	6.4 (6.0 – 6.7)	5.1 (4.5 – 5.7)	12.0

528

529

530 Table 3. Mean proportion of food items (\pm SD) ingested by *Cryptotomus roseus*, *Scarus iseri*
 531 and *Sparisoma radians*.

532

Species	Principal food items	Secondary food items	Method
<i>Cryptotomus roseus</i>	Vegetal matter (72 \pm 11%)	Invertebrates (28 \pm 5%)	Gravimetric
<i>Scarus iseri</i>	Vegetal matter (58 \pm 10%)	Sediment (12 \pm 7%) and unidentified matter (30 \pm 6%)	Point-intercept
<i>Sparisoma radians</i>	Vegetal matter (60 \pm 8%)	Sediment (7 \pm 4%) and unidentified matter (33 \pm 8%)	Point-intercept

533

534

535

536

537 Table 4. Mean contribution of food sources (%) and Bayesian Credibility Interval 95% to the
 538 diet of *Cryptotomus roseus*, *Scarus iseri* and *Sparisoma radians* calculated with SIAR mixing
 539 models.

540

Food sources	<i>C. roseus</i>	<i>S. iseri</i>	<i>S. radians</i>
	%	%	%
Invertebrates	21.1 (7.8 - 33.9)	3.5 (0.0 - 9.9)	4.9 (0.0 - 13.5)
Biofilm	14.7 (0.1 - 26.6)	22.4 (1.0 - 33.8)	10.1 (0.0 - 21.3)
Epiphytes+Detritus+ <i>Dictyota</i>	14.7 (0.7 - 26.6)	9.8 (0.0 - 23.9)	10.4 (0.0 - 24.1)
Fleshy Macroalgae			
<i>Anadyomene stellata</i>	10.2 (0.0 - 23.0)	21.9 (0.1 - 43.0)	16.8 (0.1 - 33.1)
<i>Caulerpa cupressoides</i>	9.0 (0.0 - 20.8)	11.1 (0.0 - 25.2)	16.1 (0.0 - 31.0)
<i>Chaetomorpha</i> sp	14.8 (0.1 - 28.1)	13.3 (0.0 - 28.2)	9.2 (0.0 - 21.4)
<i>Padina</i> sp.	8.4 (0.0 - 19.0)	6.9 (0.0 - 17.1)	15.3 (0.1 - 28.1)
Seagrass			
<i>Thalassia testudinum</i>	8.4 (0.0 - 19.1)	11.1 (0.0 - 24.8)	17.2 (1.0 - 31.6)

541