

Uptake of dissolved inorganic and organic nitrogen by the benthic toxic dinoflagellate Ostreopsis cf. ovata

Cécile Jauzein, Douglas Couet, Thierry Blasco, Rodolphe Lemée

▶ To cite this version:

Cécile Jauzein, Douglas Couet, Thierry Blasco, Rodolphe Lemée. Uptake of dissolved inorganic and organic nitrogen by the benthic toxic dinoflagellate Ostreopsis cf. ovata. Harmful Algae, 2017, 65, pp.9 - 18. 10.1016/j.hal.2017.04.005. hal-01513585

HAL Id: hal-01513585 https://hal.sorbonne-universite.fr/hal-01513585v1

Submitted on 25 Apr 2017

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1	Uptake of dissolved inorganic and organic nitrogen by the benthic toxic
2	dinoflagellate Ostreopsis cf. ovata
3	
4	Cécile Jauzein ^{a,b,*} , Douglas Couet ^a , Thierry Blasco ^a , Rodolphe Lemée ^a
5	
6	^a Sorbonne Universités, UPMC Univ Paris 06, INSU-CNRS, Laboratoire d'Océanographie de
7	Villefranche, Villefranche sur mer, France
8	^b IFREMER, Centre de Brest, DYNECO PELAGOS, F-29280 Plouzané, France
9	
10	
11	*Corresponding author
12	E-mail: cecile.jauzein@ifremer.fr
13	
14	
15	
16	
17	Abbreviations: Elemental Analysis – Isotope Ratio Mass Spectrometry (EA-IRMS),

particulate carbon (PC), particulate nitrogen (PN)

Abstract:

20

21

22

23

24

25

26

27

28

29

30

31

32

33

34

35

36

37

Environmental factors that shape dynamics of benthic toxic blooms are largely unknown. In particular, for the toxic dinoflagellate Ostreopsis cf. ovata, the importance of the availability of nutrients and the contribution of the inorganic and organic pools to growth need to be quantified in marine coastal environments. The present study aimed at characterizing Nuptake of dissolved inorganic and organic sources by O. cf. ovata cells, using the ¹⁵N-labelling technique. Experiments were conducted taking into account potential interactions between nutrient uptake systems as well as variations with the diel cycle. Uptake abilities of O. cf. ovata were parameterized for ammonium (NH₄⁺), nitrate (NO₃⁻) and N-urea, from the estimation of kinetic and inhibition parameters. In the range of 0 to 10 μmol N L⁻¹, kinetic curves showed a clear preference pattern following the ranking NH₄⁺ > NO₃⁻ > N-urea, where the preferential uptake of NH₄⁺ relative to NO₃⁻ was accentuated by an inhibitory effect of NH₄⁺ concentration on NO₃ uptake capabilities. Conversely, under high nutrient concentrations, the preference for NH₄⁺ relative to NO₃⁻ was largely reduced, probably because of the existence of a low-affinity high capacity inducible NO₃ uptake system. Ability to take up nutrients in darkness could not be defined as a competitive advantage for O. cf. ovata. Species competitiveness can also be defined from nutrient uptake kinetic parameters. A strong affinity for NH₄⁺ was observed for O. cf. ovata cells that may partly explain the success of this toxic species during the summer season in the Bay of Villefranche-sur-mer (France).

39

40

42

38

41

Keywords: Uptake; nitrogen; dinoflagellate; *Ostreopsis*; kinetics; interactions

1. Introduction

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

58

59

60

61

62

63

64

65

66

67

Benthic harmful algal blooms represent an increasing threat to human and environmental health worldwide (Parsons et al., 2012; Rhodes, 2011). Toxic dinoflagellates belonging to the genus Ostreopsis Schmidt are common components of tropical epibenthic microalgae communities and have also been reported in several temperate waters, including coastal waters of the Mediterranean Sea (Vila et al., 2001), New Zealand (Rhodes et al., 2000) or Japan (Taniyama et al., 2003). Along the Mediterranean coasts, massive Ostreopsis cf. ovata blooms regularly occurred during the summer season and early fall (e.g. Aligizaki and Nikolaidis, 2006; Mangialajo et al., 2011). Some of them were associated with serious cases of human health disorders (Brescianini et al., 2006; Vila et al., 2016). Symptoms of human illnesses include skin irritations, fever or broncho-constriction, partly due to exposure to toxic marine aerosols (Ciminiello et al., 2014). Blooms of O. cf ovata can also have deleterious effects on benthic marine invertebrates (Accoroni et al., 2011; Guidi-Guilvard et al., 2012; Pagliara and Caroppo, 2012; Gorbi et al., 2013). The toxicity of O. cf ovata is associated with the presence of palytoxin-like compounds that include putative palytoxin and ovatoxins-a, b, c, d, e and f (Uchida et al., 2013; Brissard et al., 2014), and mascarenotoxins-a and c (Rossi et al., 2010; Scalco et al., 2012). Palytoxin-like compounds have already been found in Mediterranean fauna (Biré et al. 2015) but no related food poisoning has been reported. The processes that shape dynamics of benthic dinoflagellate populations and facilitate

The processes that shape dynamics of benthic dinoflagellate populations and facilitate the development of specific toxic species are still poorly understood, mainly because benthic dinoflagellates have received considerably less attention than their planktonic counterparts (Parsons et al. 2012). Among potential controlling factors, temperature may represent a key factor in the seasonality of *O.* cf *ovata* blooms in temperate areas (Mangialajo et al., 2008; Accoroni et al., 2014; Accoroni and Totti, 2016). The control of bloom dynamics by water temperate has still to be clarified, however, as its appeared to vary with geographical areas

(Accoroni and Totti, 2016). Concerning other physical parameters, several studies reported higher abundances of *O*. cf *ovata* in sheltered sites compared to the ones exposed to wave action (e.g. Totti et al., 2010; Selina et al., 2014). This suggests that hydrodynamic conditions can have strong effects on bloom development and maintenance; according to Accoroni and Totti (2016), this influence of hydrodynamics on *O*. cf *ovata* bloom may be particularly pronounced under high levels of abundance (Accoroni and Totti, 2016).

68

69

70

71

72

73

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

The growth and maintenance of microalgae populations are also directly dependent on nutritive sources that are fueling the blooms. The regulation of O. cf. ovata bloom dynamics by the nutrient resource is largely unknown. Cells of *Ostreopsis* are expected to be mixotrophic, able to complete their autotrophic growth (based on photosynthesis and uptake of inorganic sources) by the use of organic matter (Burkholder et al., 2008). Among potential organic sources, the phagotrophy of preys by Ostreopsis cells was investigated (Faust et al., 1996; Barone, 2007) but is still a matter of debate (Escalera et al., 2014). The potential use of dissolved organic phosphorus by O. cf ovata cells was tested by Pistocchi et al. (2014), when the uptake of dissolved organic nitrogen sources has not been analyzed yet. Concerning the inorganic sources of nutrients, previous studies reported conflicting results regarding relationships between nutrient availability and occurrence of Ostreopsis blooms (Accoroni and Totti, 2016). Several field studies conducted in the NW Mediterranean Sea did not show any relationship between epiphytic O. cf. ovata abundances and concentrations of dissolved inorganic nutrients (dissolved inorganic nitrogen, DIN, and phosphate) (Vila et al., 2001; Accoroni et al., 2011). Conversely, Parsons and Preskitt (2007) found that Ostreopsis sp.1 abundance was positively correlated with nutrient concentrations in the waters surrounding Hawai'i. A positive correlation between phosphate concentration and O. cf. ovata abundance was also reported by Cohu et al. (2013) in the NW Mediterranean Sea. In the Northern Adriatic Sea, phosphate pulses in the bloom onset period may possibly stimulate *O* cf. *ovata* growth in these coastal waters (Accoroni et al., 2015).

The importance of the availability of nutrient sources and their contribution to *O* cf. *ovata* growth during bloom development and maintenance need to be quantified in marine coastal environments. In the present study, the control of *O* cf. *ovata* growth by several nitrogen (N) sources was investigated under controlled conditions, using cultures. The main goal of the present work was to characterize N-uptake of dissolved inorganic and organic sources, using the ¹⁵N-labelling technique and taking into account potential interactions between nutrient uptake systems as well as variations with the diel cycle.

2. Material and methods

2.1. Culture conditions

Two strains of *Ostreopsis* cf. *ovata*, MCCV 054 and MCCV 055, were obtained from the Mediterranean Culture Collection of Villefranche (MCCV). They were both isolated in 2014 from Villefranche Bay, South of France (43°41′34.83″ N and 7°18′31.66″ E), during the same bloom event. Non-axenic stock cultures were grown in modified K/10 medium (originally defined by Keller et al. (1987)), where addition of silicate and Tris base was omitted, phosphorus was added as KH₂PO₄ (final concentration of 4 μM) and ZnSO4 was added at a final concentration of 0.08 nM. Culture medium was prepared using autoclaved old seawater filtered on 0.2 μm (FSW) at salinity 38. Cultures were maintained at 23°C, under 250 μmol photons m⁻² s⁻¹, with a 16:8 h light:dark cycle. Stock cultures were grown in batch mode without bubbling, in 15 mL of culture medium. Culture flasks were maintained in flat culturing conditions in order to optimize the surface area for gas exchange and growth of benthic cells. Before each experiment, one stock culture in exponential phase of growth was successively

diluted in order to scale up the culture volume from 15 ml (in flask of 25 cm² surface area) to 350 ml (in flask of 300 cm² surface area). The final large volume culture was used to inoculate three or four replicated cultures of 350 ml. Experiments were run using a set of replicated cultures in exponential phase and characterized by a cell density higher than 1,500 cell ml⁻¹.

121

122

123

124

125

126

127

128

129

130

131

132

133

134

135

136

137

138

139

140

120

117

118

119

2.2. Micro-algal cell resuspension in low N medium

Experiments were conducted under controlled conditions of nitrogen (N) availability in order to help for a precise characterization of N-uptake capabilities of O. cf. ovata cells. Each experiment started with the resuspension of micro-algal cells in culture medium where no NH₄⁺ or NO₃⁻ addition was performed (-N medium). Concentrations of NH₄⁺ and NO₃⁻ were determined for the -N medium used for running the experiments. Full resuspension of O. cf. ovata cells was completed in about 1h. Cells were collected on an 8 or 10 µm mesh size net by gravity filtration, then rinsed with -N medium before being resuspended in -N medium. To ensure that the net was not clogged due to mucus accumulation, these collection, rinsing and resuspension steps were performed on successive aliquots of 35 or 40 mL of culture and a new piece of net was used every four aliquots. A gentle agitation of the net in -N medium did not allow for passive resuspension of O. cf. ovata cells. Thus, for each aliquot of culture, microalgal cells concentrated on the net were collected by pipetting repeatedly and carefully ~1 ml of -N medium above the net, then this volume was finally poured in a culture flask (75 cm² surface area) filled with 40 mL of –N medium. The resuspension and rinsing steps allowed for the removal of most of the bacteria present in the growth medium and limited their contribution in the resuspended cultures (Raush de Traubenberg and Soyer-Gobillard, 1990).

The resuspended culture flasks were kept aside in the culture chamber, under initial culture conditions, during 1-2h before starting the incubations. This lag reduced the potential

impact of stress associated with the resuspension step on uptake rates and also contributed to start incubations under really low N concentrations.

2.3. Kinetic experiments

Uptake kinetics of three potential N-sources, nitrate (NO₃-), ammonium (NH₄⁺) and urea, were characterized for the two *O.* cf. *ovata* strains, MCCV 054 and MCCV 055. For each strain, *O.* cf. *ovata* cells were resuspended from three replicated cultures of 350 ml in exponential phase. Each mother culture allowed for the creation of one series of eight 40 ml samples and was used to characterize the uptake kinetics of one N-source. Incubations started with the addition of ¹⁵N (¹⁵NO₃-, ¹⁵NH₄+ or ¹⁵N-urea) at eight graded concentrations (0.1, 0.2, 0.5, 1, 2, 3, 5, and 10 μmol N L⁻¹). Samples were incubated for 1h under initial culture conditions. At the end of the incubation, samples were filtered through precombusted (4 h at 450°C) A/E filters (Gelman Sciences) and rinsed with 20 mL of FSW. Filters were dried at 60°C overnight and analyzed by EA-IRMS (Elemental Analysis – Isotope Ratio Mass Spectrometry) for measurements of particulate carbon (PC), particulate nitrogen (PN) and ¹⁵N/¹⁴N isotopic ratios.

An additional experiment was conducted in order to characterize N-urea uptake capabilities of O. cf. ovata cells taking into account the potential role of preconditioning effects. Cells of O. cf. ovata were grown on a modified K/10 medium containing three potential N-sources: NO_3^- added at 28.8 μ mol N L⁻¹ and NH₄⁺ and N-urea added at 5 μ mol N L⁻¹. These growth conditions were maintained during several culture transfers in batch mode. Then, one culture of 350 mL in exponential phase was used for running a replicated kinetic experiment, in order to estimate N-urea uptake rates along a concentration gradient of $0-10~\mu$ mol N L⁻¹.

2.4. Interaction experiments

An experiment was run in order to characterize the potential interaction between NH₄⁺ and NO₃⁻ uptake. Four replicated cultures of 350 mL (MCCV 054) were used to carry out two successive series of incubations, one testing the influence of NH₄⁺ on the maximal uptake rate of NO₃, and the other the influence NO₃ on the maximal uptake rate of NH₄. During each part of the experiment, the uptake rate of one nutrient, added at a reference concentration of 10 µmol N L⁻¹, was measured as a function of the increasing concentration of the other nutrient (0, 0.1, 0.2, 0.5, 1, 2, 3, 5, and 10 µmol N L⁻¹). For each set, two series of incubations were performed in parallel in order to simultaneously assay uptake rates of both nutrients (NH₄⁺ and NO₃⁻) for all nutritive conditions. These coupled incubations were based on the same nutrient regime, with only one of the two N-sources labeled with ¹⁵N: for one series of samples, the nutrient added under various concentration (0-10 μ mol N L⁻¹) was labeled with 15 N and, for the other series of samples, the nutrient added at saturating concentration (10 µmol N L⁻¹) was labeled with ¹⁵N. Incubations started with the addition of ¹⁵NH₄⁺ or ¹⁵NO₃⁻ into 40 mL samples and lasted 1h. Incubations ended with the filtration of samples through precombusted (4 h at 450°C) A/E filters (Gelman Sciences). Filters were rinsed with 20 mL of FSW, then dried at 60°C overnight. Analyses were run using EA-IRMS in order to obtain measurements of PN, PC and ¹⁵N/¹⁴N isotopic ratios.

183

184

185

186

187

188

189

190

182

166

167

168

169

170

171

172

173

174

175

176

177

178

179

180

181

2.5. Diel cycle experiments

Variations of NH₄⁺- and NO₃⁻-uptake by *O.* cf. *ovata* cells were investigated over the diel cycle. For each N-source, three replicated cultures of 350 ml (MCCV 054) were used and allowed for the preparation of three series of ten resuspended samples, each of them containing 40 mL of –N medium. At the beginning of the incubations, all samples were spiked with a solution of ¹⁵N (¹⁵NH₄⁺ or ¹⁵NO₃⁻) at 100 μmol N L⁻¹ final concentration and were immediately replaced in the culture chamber under initial conditions. This level of concentration was used

to ensure N-sufficient conditions all along the experiment duration. Regular stops in the incubations were done during a 24h survey (every 3 h during light periods and three times during the dark phase). At each stop, three samples were taken, one originating from each of the replicated mother cultures. Cells of *O.* cf. *ovata* were collected on precombusted (4 h at 450°C) A/E filters (Gelman Sciences) and rinsed with 20 mL of FSW. Half of the samples were used to follow nutrient concentrations during the experiment, collecting 20 mL of the culture filtrate during microalgal cell collection (before the rinsing step). Filters were finally dried at 60°C overnight, then analyzed by EA-IRMS in order to obtain measurements of PN, PC and ¹⁵N/¹⁴N isotopic ratios.

2.6. Cell counts and nutrient analysis

Growth rate was estimated for each mother culture used for running experiments, from measurements of cell density done just before the resuspension step and 24h earlier. These growth rate estimations allowed to verify that *O.* cf. *ovata* cells were growing under optimal growth conditions when incubations started. For sampling, three 2 mL-aliquots of culture were taken after a gentle mixing of the culture and pooled together before counting. Samples were fixed with acidic lugol solution at 1% (vol/vol) final concentration and stored at +4°C until analysis. Cell counts were done in triplicate using a 1 mL Sedgewick rafter counting chamber.

Growth rates were calculated according to Guillard (1973), using the following formula:

210
$$\mu = \frac{\ln(C2) - \ln(C1)}{t2 - t1}$$

- where μ is the growth rate (d⁻¹), C₁ and C₂ are the cell concentrations at time 1 (t₁, d) and time 2 (t₂, d), respectively.
- Measurements of NH₄⁺ concentrations were performed few hours after sampling, using the fluorometric method (Taylor et al., 2007). Samples for estimations of NO₃⁻ concentration were immediately frozen at -20°C and stored until analysis. Concentrations of NO₃⁻ were

measured using an automated colourimetry system (Seal Analytical continuous flow AutoAnalyser III, AA3)) as described by Bendschneider and Robienson (1952).

218

219

220

221

222

223

224

225

226

227

228

229

216

217

2.7. N-uptake measurements and kinetic parameters

For determination of PN, PC and ¹⁵N/¹⁴N isotopic ratios, EA-IRMS experiments were done with an Elementar Vario Pyro Cube analyzer in CN mode (combustion oven 920°C, reduction oven 600°C) coupled to an Isoprime 100 IRMS (Isotope Resolved Mass Spectrometer). Calibration of measurements was performed with certified caffeine (AIEA-600) and other laboratory standards (commercially available glycine (Sigma), acetinalide (Merck)).

Uptake rates (V in h⁻¹) were calculated from the ¹⁵N enrichment of the samples according to Collos (1987). For kinetic and interaction experiments, relationship between uptake rates and concentrations that showed clear saturating kinetics were modeled using the original or a modified equation of the Michaelis-Menten model. When the ¹⁵N-source was added at graded concentrations, uptake data were modeled using the original Michaelis-Menten relation:

230
$$V_N = V_{\text{max-N}} \times [N] / (K_s + [N])$$
 (1)

- Where V_N (in h⁻¹) is the N-uptake rate under a nutrient concentration of [N] (in μ mol N L⁻¹),
- $V_{\text{max-N}}$ is the maximal uptake rate (in h⁻¹) and K_s is the half-saturation constant (in μ mol N L⁻¹).
- For these kinetics, the initial slope α was also calculated from the uptake rate at the
- concentration of 0.5 $\mu mol\ N\ L^{\text{--}1}$ estimated by the model equation as recommended by Hurd and
- Dring (1990), and was used as an indicator of the competitive ability of the cells at low substrate
- concentrations.
- For the interaction experiment, the exponential decrease in the uptake rate of one
- nutrient (N1) when increasing the concentration of the other (N2) was fitted to the reverse
- 239 Michaelis-Menten relation (Varela and Harrison, 1999):

240
$$V_{\text{N1}} = V_{\text{max-N2=0}} \times (1 - (I_{\text{max}} \times /N2) / (K_{\text{I}} + /N2))$$
 (2)

Where the N-uptake rate of the nutrient N1, V_{N1} (in h⁻¹), is function of the maximum uptake rate without inhibition ($V_{\text{max-N2=0}}$, in h⁻¹), the concentration of the inhibitory nutrient [N2] (in μ mol N L⁻¹), the maximum inhibition I_{max} (values from 0 to 1) and of the inhibition constant K_{I} (concentration of N2 at which $I = I_{\text{max}} / 2$, in μ mol N L⁻¹).

For the diel cycle experiment, uptake rates during each light and dark period were estimated from linear regressions of isotopic ratios *vs.* time.

Values of kinetic parameters were obtained from non-linear regressions of data sets, using the Statgraphics Centurion software (Manugistics, Inc.). Statistical tests (significance and comparison of regression slopes) were performed using the same software.

3. Results

3.1. Culture medium and cellular growth

In terms of nutrient availability, low N-conditions were verified in the culture medium used for running the experiments. The medium used for resuspension and incubation of O. cf. ovata cells was characterized by an NH₄⁺ concentration of 0.82 μ mol N L⁻¹ (SD± 0.27 μ mol N L⁻¹) and NO₃⁻ concentration of 1.00 μ mol N L⁻¹ (SD± 0.01 μ mol N L⁻¹).

All cultures used for running the experiments were growing exponentially, under similar growth conditions. At resuspension time, the growth rates of replicated cultures were, on average, 0.39 d⁻¹, 0.45 d⁻¹ and 0.51 d⁻¹ for the kinetic experiment, the interaction experiment and the diel cycle experiment, respectively. Quantities of particulate organic nitrogen (PN) and carbon (PC) allowed for estimations of C:N (atomic) ratio. Mean C:N ratios of 12.3 (SD± 1.0) and 12.5 (SD± 0.8) were estimated for the strains MCCV 054 and MCCV 055, respectively, from the compilation of data sets obtained during short-term experiments (1h-incubations). Extra-cellular mucilage might have interfered in the precision of these estimations. Trends

observed in C:N ratios did not allow for a detailed characterization of the coupling between N-and C-fluxes over the diel cycle.

3.2. Uptake rates during kinetic experiments

Variations of 15 N-enrichment over 1h-incubations showed that O. cf. ovata cells were able to use dissolved inorganic N-sources (NH₄⁺ and NO₃⁻) and dissolved organic nitrogen (Nurea). Similar saturating kinetic curves were observed for the two strains of O. cf. ovata tested, when NH₄⁺, NO₃⁻ or N-urea was added as a unique N-source along a gradient of $O - 10 \mu mol$ N L⁻¹ (Figure 1). Out of the three potential N-sources tested, O. cf. ovata cells showed a clear preference pattern following the ranking: NH₄⁺ > NO₃⁻ > N-urea. No potential preconditioning effect influenced this ranking because O. cf. ovata cells were grown in K/10 medium with NH₄⁺ and NO₃⁻ added as N-sources and an acclimation of cells to the presence of N-urea in the culture medium did not induce a clear modification of N-urea kinetics (Figure 1C). For both strains, on average along the whole gradient, NH₄⁺ uptake rate was 4 to 5 times higher than NO₃⁻ uptake rate, when NO₃⁻ uptake rate was 8 to 9 times higher than N-urea uptake rate.

Relationships between uptake rates and nutrient concentrations were characterized by the Michaelis-Menten model. Similar estimations of kinetic parameters (V_{max} , K_s , α) were obtained for the two strains (Table 1). This similarity allowed to characterize abilities of O. cf. ovata cells isolated in the Bay of Villefranche-sur-mer by a maximal uptake rate (V_{max}) of 0.021 h^{-1} (SD± 0.001 h^{-1}), 0.008 h^{-1} (SD± 0.003 h^{-1}), 0.0005 h^{-1} (SD± 0.0001 h^{-1}) for NH₄⁺, NO₃⁻ and N-urea, respectively. The associated K_s values were 0.5 μ mol N L⁻¹ (SD± 0.1 μ mol N L⁻¹), 2.3 μ mol N L⁻¹ (SD± 2.1 μ mol N L⁻¹) and 0.3 μ mol N L⁻¹ (SD± 0.1 μ mol N L⁻¹) for NH₄⁺, NO₃⁻ and N-urea, respectively. For the characterization of abilities under low nutrient concentrations, the initial slope of kinetic curves (α) was 0.011 L μ mol N⁻¹ h^{-1} (SD± 0.001 L μ mol N⁻¹ h^{-1}),

 $0.002 \text{ L } \mu\text{mol N}^{-1} \text{ h}^{-1} \text{ (SD\pm 0.001 L } \mu\text{mol N}^{-1} \text{ h}^{-1} \text{)} \text{ and } 0.0003 \text{ L } \mu\text{mol N}^{-1} \text{ h}^{-1} \text{ (SD\pm 0.0001 L}}$ 291 $\mu\text{mol N}^{-1} \text{ h}^{-1} \text{)} \text{ for NH}_4^+, \text{NO}_3^- \text{ and N-urea, respectively.}}$

3.3. Uptake rates during interaction experiments

The influence of NH_4^+ on the maximal uptake rate of NO_3^- ($V_{max-NO3-}$) was analyzed from estimations of $^{15}NH_4^+$ and $^{15}NO_3^-$ -uptake rates after an addition of 10 µmol N L⁻¹ of NO_3^- and an NH_4^+ concentration varying from 0 to 10 µmol N L⁻¹ (Figure 2A). In the presence of 10 µmol N L⁻¹ of NO_3^- , the relationship between NH_4^+ uptake rate and NH_4^+ concentration displayed a saturating kinetic (Figure 2A) with a high V_{max} of 0.034 h⁻¹ and a K_8 value (0.7 µmol N L⁻¹) close to estimations done when NH_4^+ was added as the only N-source (Table 1). Variations of $V_{max-NO3-}$ along the NH_4^+ gradient showed an exponential decrease that could be characterized by fitting the data set to the reverse Michaelis-Menten model (2) (Figure 2A). The inhibition parameters generated by the model (Table 1) showed a strong NH_4^+ inhibition of NO_3^- uptake rate, with a maximum inhibition value of 67% and a K_1 value of 6.2 µmol N L⁻¹

Under reverse nutrient conditions (addition of 10 μ mol N L⁻¹ of NH₄⁺ along a NO₃⁻ gradient of 0-10 μ mol N L⁻¹), variations of NO₃⁻ uptake rates were characterized by kinetic parameters ($V_{\text{max}} = 0.008 \text{ h}^{-1}$, $K_{\text{s}} = 2.8 \mu$ mol N L⁻¹) after fitting the data set to the Michaelis-Menten model (1) (Figure 2B, Table 1). Concerning variations of $V_{\text{max-NH4+}}$ with increasing NO₃⁻ concentration, the reverse Michaelis-Menten model did not converge, however, even if the NH₄⁺ uptake rate obtained after addition of 10 μ mol N L⁻¹ of NH₄⁺ and 10 μ mol N L⁻¹ of NO₃⁻ appeared to be slightly low Figure 2B). Globally, NH₄⁺ uptake rate was stable along the NO₃⁻ gradient, with $V_{\text{max-NH4+}} = 0.026 \text{ h}^{-1}$ (SD± 0.002 h⁻¹).

3.4. Diel cycle experiment

After the addition of 100 μ mol N L⁻¹ of ¹⁵N (¹⁵NH₄⁺ or ¹⁵NO₃⁻), linear decreases in nutrient concentrations were observed with time for both series of incubations (data not shown). Despite the consumption of nitrogen sources, final concentrations of NH₄⁺ and NO₃⁻ in replicated flasks were 84 μ mol N L⁻¹ (SD± 9 μ mol N L⁻¹) and 82 μ mol N L⁻¹ (SD± 4 μ mol N L⁻¹), respectively, after 24h of incubation. These estimations showed the maintenance of N-replete conditions over the whole experiment duration (24h).

According to estimations of 15 N-enrichment of O. cf. ovata cells over the diel cycle, microalgal cells were capable of using both NH₄⁺ and NO₃⁻ during the light and the dark periods (Figure 3). For both NH₄⁺ and NO₃⁻, linear increases in 15 N-atom (%) of O. cf. ovata cells were observed with time during each consecutive light and dark periods. During the first light period, trends were highly significant for both NH₄⁺ and NO₃⁻ (linear regressions with $r^2 = 0.99$, p < 0.001, Table 2 and Figure 3). Slopes of linear regressions allowed for precise estimations of N-uptake rates and indicated that NH₄⁺-uptake rate was higher (0.032 h⁻¹) than but also close to NO₃⁻-uptake rate (0.030 h⁻¹) during the light period. During the subsequent dark phase, values of 15 N-atom (%) were coherent between replicated flasks and showed an increasing trend with time for both NH₄⁺ and NO₃⁻ (Figure 3). Slopes of linear regressions that modelled dark processes were not significant ($r^2 \ge 0.78$, $p \ge 0.06$, Table 2), however, suggesting a lack of precision in dark N-uptake rate estimations. From present results, dark uptake rates corresponded to 19% and 10% of light uptake rates for NH₄⁺ and NO₃⁻, respectively.

4. Discussion

Nutrient uptake capabilities of phytoplankton cells are known to vary as a function of cell size (Litchman et al., 2007), nutritional history and physiological status of the cells (Mulholland and Lomas, 2008), growth rates (Maguer et al., 2007), N substrate interactions

(Maguer et al., 2007; Jauzein et al., 2008a) and environmental factors such as irradiance and temperature (Lomas and Glibert, 1999; Kudela and Cochlan, 2000). Consequently, it is often difficult to determine taxa-specific differences in uptake capabilities and environmental control on uptake from field studies. Characterization of uptake capabilities under controlled conditions from culture studies gives the opportunity to better understand uptake regulation. Relatively few studies have determined N-uptake kinetics from cells that are nitrogen replete. In previous culture works, kinetic parameters were often characterized for cells under N-limited conditions or after several days of N-starvation (e.g. Nishikawa et al., 2009; Kwon et al., 2013). Nutrient depletion can lead to transient or surge uptake, however, when an uncoupling between nutrient uptake and growth occurs (Dortch et al., 1982; Mulholland and Lomas, 2008). In the present study, efforts were made at characterizing N-uptake of cells growing under optimal conditions. Experiments started by the resuspension of exponentially growing cells in –N medium, when physical conditions were optimized to ensure no limitation of uptake capabilities. Ammonium concentrations used to monitor uptake rates were also lower than concentrations known to potentially inhibit growth of dinoflagellate cells (Collos and Harrison, 2014; Siu et al., 1997). Thus, patterns and non-linear regressions of data sets obtained allow for the characterization and parameterization of functional responses of O. cf. ovata cells to N-sources availability. In particular, values of parameters, such as half-saturation constants and inhibition parameters, are crucial for ecological modeling and understanding of forcing functions (Tian, 2006). In the present study, estimations were done when uptake of microalgal cells was tightly coupled with growth; values of parameters obtained can be used for the definition of mechanistic formulations that simulate the function for nutrient limitation of O. cf. ovata growth.

340

341

342

343

344

345

346

347

348

349

350

351

352

353

354

355

356

357

358

359

360

361

362

363

364

Nutrient uptake by microalgal cells is an active process, whose response to extracellular nutrient concentration can generally be modelled as that of an enzyme, using Michaelis-Menten kinetics. In the Michaelis-Menten model, the maximum uptake rate ($V_{\rm max}$) and half-saturation

constant (K_s) are often said to be biological parameters, dependent on the number of carrier sites on cell membrane and specific efficiency of each transporter (Aksnes and Egge, 1991; Litchman et al., 2007). Estimations of these parameters, along with the affinity coefficient α , have been used to assess the relative preference for different N-substrates or competitive abilities between species under various nutritive conditions (Cochlan et al., 2008; Mulholland and Lomas, 2008). In the present study, kinetic curves were characterized in the range of few umol N L⁻¹, for two strains isolated during the same bloom event. They show that O. cf. ovata cells are able to use dissolved inorganic N-sources (NO₃⁻ and NH₄⁺) and dissolved organic sources (N-urea) with a clear preference pattern: this pattern follows the ranking NH₄⁺> NO₃⁻ > N-urea and was well defined all over the gradient tested (0.1 – 10 µmol N L⁻¹). Preconditioning effects did not interfere in these trends for neither of the N-sources tested. Results also show that the preferential uptake of NH₄⁺ relative to NO₃⁻ is accentuated for O. cf. ovata cells by an inhibitory effect of NH₄⁺ concentration on NO₃⁻ uptake capabilities. Conversely, no influence of NO₃⁻ availability on NH₄⁺ uptake was observed for this species. Repression of NO₃⁻ uptake by NH₄⁺ has been well studied for many decades for several phytoplankton species (Glibert et al., 2016), but never for benthic dinoflagellates. The maximal inhibition estimated for O. cf. ovata cells ($I_{\text{max}} = 67\%$) is similar to values reported for other dinoflagellates (Alexandrium minutum, Prorocentrum minimum, Gyrodinium uncatenum) (Lomas and Glibert, 1999; Maguer et al., 2007). The half-inhibition constant (K_I) estimated for O. cf. ovata (6.2 µmol N L⁻¹) appears really high compared to values reported in previous studies, however, in particular for N-sufficient microalgal cells (e.g., Lomas and Glibert, 1999; Maguer et al., 2007); this suggests a low sensitivity of NO₃ uptake of O. cf. ovata cells to NH₄⁺ concentration, in particular under low NH₄⁺ availability. On a broader point of view, NH₄⁺ is commonly found to be the preferred N-source over

NO₃⁻ and N-urea for phytoplankton uptake (Mulholland and Lomas, 2008; Glibert et al., 2016,

365

366

367

368

369

370

371

372

373

374

375

376

377

378

379

380

381

382

383

384

385

386

387

388

and references therein). The preferential use of NH₄⁺ is attributed largely to the low energetic demand for its uptake and assimilation (Syrett, 1981). Exceptions have been documented, however, such as the preference of *Pseudo-nitzschia australis* for NO₃⁻ over NH₄⁺ and N-urea reported by Cochlan et al. (2008). Most importantly, the preference for NH₄⁺ over NO₃⁻ may strongly depend on the range of nutrient considered. As well explained and conceptualized by Glibert et al. (2016), the preference for NH₄⁺/NO₃⁻ can be inverted under high nutrient conditions, due to either (i) the toxicity and growth inhibition of high NH₄⁺ concentrations and/or (ii) the potential acceleration of NO₃ uptake in the presence of NO₃ that can lead to biphasic kinetics. In the present study, results suggest that such an acceleration on NO₃ uptake occurs for O cf. ovata cells exposed to high NO₃⁻ concentrations. Indeed, NH₄⁺ uptake rates were 4 to 5 times higher than NO₃ uptake rates during kinetic experiments that were conducted in the range 0.1 – 10 µmol N L⁻¹. During the light period of the diel-cycle experiment, estimations of mean NO₃⁻ uptake rate (0.030 h⁻¹) were close to estimations of mean NH₄⁺ uptake rate (0.032 h^{-1}) after addition of 100 μ mol N L⁻¹ of NH₄⁺ or NO₃⁻, however. This mean NO₃⁻ uptake rate obtained under high nitrate concentrations was more than 3 times higher than the $V_{\rm max}$ value of 0.008 h⁻¹ estimated in the range 0.1 – 10 µmol N L⁻¹. Such variations in N-uptake capabilities are consistent with the existence of a two-component NO₃ uptake system, involving a high-affinity low-capacity constitutive component and a low-affinity high capacity inducible uptake component (Glibert et al., 2016). Results obtained in the present study are consistent with the threshold of about 60 µmol N L⁻¹ that was reported for the transition between biphasic kinetics of NO₃⁻ uptake for several phytoplankton species (Collos et al., 1992; Lomas and Glibert, 2000).

390

391

392

393

394

395

396

397

398

399

400

401

402

403

404

405

406

407

408

409

410

411

412

413

414

The monitoring of N-uptake rates done over the diel cycle allows for the characterization of dark N-uptake capability of *O* cf. *ovata* under N-sufficient conditions. Dark uptake of N compounds is commonly observed in marine waters (e.g. Cochlan et al., 1991; Fan and Glibert,

2005; Maguer et al., 2015). For photosynthetic cells, nutrient uptake and assimilation in darkness occurs at the expense of previously accumulated carbon that will supply dark processes with energy (ATP), reductant (NAD(P)H) and C-skeletons (Turpin, 1991). Photosynthetic carbon can be stored in excess during the light period into C-rich and N-free macromolecules, such as carbohydrates (Clark and Flynn, 2002; Granum et al., 2002) or neutral lipids (Fabrégas et al., 2002). Detailed observations of morphological and metabolic features O. cf. ovata cells revealed that their cytoplasm is often full of neutral lipid droplets, in all stages of growth under N-sufficient conditions (Honsell et al., 2013). This suggests a potential for Cstorage strategies that could support dark processes. Various taxonomic groups of phytoplankton, including dinoflagellates, prymnesiophytes and diatoms, carry out uptake at night under N-sufficient conditions (e.g. Paasche et al., 1984; Clark et al., 2002; Needoba and Harrison, 2004), with reported dark: light (D/L) uptake ratios ranging from 1% to 75% for NO₃⁻ and from 21% to 100% for NH₄⁺ (Jauzein et al., 2011, and references therein). With D/L uptake ratios of 10% and 19% measured for NO₃⁻ and NH₄⁺ in the present study, O. cf. ovata shows low capabilities for N-uptake in darkness, at least under N-sufficient conditions. Thus, the lipid storage strategy noted by Honsell et al. (2013) has to be explored further through additional experiments to define its implications in the species competitiveness. As dark N-uptake processes have been shown to be enhanced under N-limited conditions (Paasche et al., 1984; Turpin, 1991), it could also be interesting to complete the characterization of dark N-uptake capabilities of O. cf. ovata testing N-limited conditions and/or in situ measurements.

415

416

417

418

419

420

421

422

423

424

425

426

427

428

429

430

431

432

433

434

435

436

437

438

439

According to present results, ability to take up nutrients in darkness cannot be seen as a competitive advantage for O. cf. ovata. Additional information about species competitiveness can be defined from nutrient uptake kinetic parameters (e.g. Smayda, 1997; Litchman et al., 2007). On a broad point of view, reviews of parameter values reported in both field studies and experimental works show that dinoflagellates have generally higher K_s for DIN than diatoms

(Smayda, 1997; Kudela et al., 2010). Diatoms were also characterized by higher carbon-specific V_{max} (values standardized by C units in order to diminish the effect of cell size) for NO_3^- than other taxa, including dinoflagellates (Litchman et al., 2007). Thus, a general trend of low competitive abilities for acquisition of DIN can be defined for dinoflagellates that puts emphasis on the contribution of organic matter in fueling dinoflagellate blooms in oligotrophic or mesotrophic coastal waters (e.g. Collos et al., 2004).

440

441

442

443

444

445

446

447

448

449

450

451

452

453

454

455

456

457

458

459

460

461

462

463

464

For the benthic compartment in particular, studies reporting uptake parameters are rare. Estimations of NO₃⁻ uptake kinetics were done for several species of benthic diatoms by Kwon et al. (2013) and for O. cf. ovata by Pistocchi et al. (2014). These studies report surprisingly high K_s values (ranging from 6.75 μ mol N L⁻¹ to 9.29 μ mol N L⁻¹) compared to literature on planktonic species (Kudela et al., 2010). For O. cf. ovata, K_s value reported by Pistocchi et al. (2014) (8.4 μmol N L⁻¹) is strongly higher than values (< 4 μmol N L⁻¹) obtained in the present study for strains isolated from French coastal waters. Strong differences in N-uptake abilities can be observed between strains of the same species (Jauzein et al., 2008b). It has also to be noted that some methodological choices can partly explain the high level of K_s reported in Kwon et al. (2013) and Pistocchi et al. (2014) for NO₃ uptake of benthic species. These studies were conducted on a large NO₃⁻ concentration gradient, ranging from 1 umol N L⁻¹ to 100 umol N L⁻¹, when kinetic experiments of the present study were run on 0.1-10 μmol N L⁻¹. Collos et al. (2005) pooled data from 20 different studies dealing with phytoplankton uptake and showed a direct increase in K_s values with maximal NO_3 concentration used in the respective studies. This correlation can be explained by an acclimation of microalgal cells to high NO₃⁻ concentrations coming from the existence of multiphasic uptake systems (Collos et al., 2005; Glibert et al., 2016); as explained above, such a biphasic system is suspected for O. cf. ovata. Thus, kinetic parameters estimated by Kwon et al. (2013) and Pistocchi et al. (2014) are probably representative of microalgal cell responses to high nutrient availability, but might not provide an appropriate representation of uptake capabilities under low N-conditions, making these results hardly comparable to the present study.

465

466

467

468

469

470

471

472

473

474

475

476

477

478

479

480

481

482

483

484

485

486

487

488

489

Uptake kinetics are also known to potentially vary with N-history and physiological status of microalgal cells (Mulholland and Lomas, 2008). When compiling exclusively results obtained from actively growing cultures, kinetic parameter values determined for NH_4^+ , NO_3^- and N-urea uptake by O. cf. *ovata* in the present study are in accordance with values reported for other planktonic dinoflagellates (Table 3). For NO_3^- , the compilation of these data sets also highlights higher V_{max} for diatoms compared to dinoflagellates, when no pattern can be defined for K_8 (Table 3).

In the Bay of Villefranche-sur-mer (South of France), nutritive conditions can be seen as oligotrophic to mesotrophic (Selmer et al., 1993). Recent blooms of O. cf. ovata in this bay were observed when NH_4^+ , NO_3^- and N-urea concentrations ranged between $0.04 - 0.27 \mu mol$ N L⁻¹, $0.20 - 2.12 \mu mol N L^{-1}$ and $0.3 - 2.25 \mu mol N L^{-1}$, respectively (data not shown). Under such low availability in N-sources, affinity (K_s) more than velocity (V_{max}) should control species competitiveness. In the present study, values of K_s estimated for NH_4^+ , NO_3^- and N-urea uptake by O. cf. ovata are in the upper part of these ranges of in situ concentrations. This suggests a good adaptation of the cells to field conditions but not a strong affinity strategy. To go further on the definition of specific competitive abilities, affinity for N-sources can also be compared to other taxa taking into account field studies and other culture works. Indeed, characterization of N-uptake kinetics of microalgal cells as a function of N-limitation did not always show variations of K_s with N-status or cell nutritional history (Hu et al., 2014; Maguer et al., 2007). According to the review done by Kudela et al. (2010), O. cf. ovata shows relatively high K_s (low affinity) for NO_3^- , as most of the dinoflagellates do, but a low K_s for NH_4^+ compared to both dinoflagellates and diatoms. This defines strong competitive abilities of O. cf. ovata for NH₄⁺ uptake under low N-conditions, like the ones encountered during bloom seasons in the Bay of Villefranche-sur-mer. Mixotrophic abilities of *O.* cf. *ovata* have not been fully characterized yet and are still a matter of debate (Escalera et al., 2014). Current results characterize the potential use of N-urea as a source of labile DON for *O.* cf. *ovata* cells. According to uptake capabilities and ranges of *in situ* conditions, contribution of N-urea to growth of *O.* cf. *ovata* in Villefranche-sur-mer Bay is probably low compared to DIN sources, even if this source can be rapidly regenerated in the water column (Lomas et al., 2002). Out of the three N-sources tested in the present study, the main N-source fueling blooms *O.* cf. *ovata* in the Bay of Villefranche-sur-mer is probably NH₄⁺, a recycled N-source for which *O.* cf. *ovata* showed highest uptake rates and good competitive abilities.

5. Conclusions

The present study provides a detailed parameterization of N-uptake by *O.* cf. *ovata*. Kinetic and inhibition parameters can be used for the definition of mechanistic formulations in order to simulate growth limitation by nutrient availability. Ability to take up nutrients in darkness could not be defined as a competitive advantage for *O.* cf. *ovata* during exponential growth. Conversely, a strong affinity for NH₄⁺ was observed for *O.* cf. *ovata* cells and may partly explain the success of this species during the summer season in the Bay of Villefranchesur-mer (France). Further studies will be necessary to clarify the role of organic matter in growth of *O.* cf. *ovata* cells during bloom development and maintenance. It could also be interesting to better characterize links between C fluxes and other metabolic processes in *O.* cf. *ovata* cells, trying to define for example the potential role of the lipid storage strategy in terms of species competitiveness.

Acknowledgements

The authors are grateful to Sophie Marro for isolating and maintaining the algal strains. Those experiments were done with the financial assistance of the European Union under the ENPI CBC Mediterranean Sea Basin Programme, within the project M3-HABs (Project Reference Number: IIB/ 2.1/0096). The authors also wish to thank the French National Project ANR OCEAN 15 for its contribution for the financial support and the fruitful discussions with project members. Our group is part of the national French GDR PHYCOTOX (CNRS and Ifremer). The authors also have deep thoughts for Yves Collos who inspired this work.

522

523

524

References

- Accoroni, S., Romagnoli, T., Colombo, F., Pennesi, C., Di Camillo, C.G., Marini, M.,
 Battocchi, C., Ciminiello, P., Dell'Aversano, C., Dello Iacovo, E., Fattorusso, E.,
 Tartaglione, L., Penna, A., Totti, C., 2011. *Ostreopsis* cf. *ovata* bloom in the northern
 Adriatic Sea during summer 2009: Ecology, molecular characterization and toxin profile.
 Mar. Pol. Bull. 62, 2512-2519.

 Accoroni, S., Romagnoli, T., Pichierri, S., Totti, C., 2014. New insights on the life cycle stages
- of the toxic benthic dinoflagellate *Ostreopsis* cf. *ovata*. Harmful Algae 34, 7-16.
- Accoroni, S., Glibert, P.M., Pichierri, S., Romagnoli, T., Marini, M., Totti, C., 2015. A conceptual model of annual *Ostreopsis* cf. *ovata* blooms in the northern Adriatic Sea based on the synergic effects of hydrodynamics, temperature, and the N:P ratio of water column nutrients. Harmful Algae 45, 14-25.
- Accoroni, S., Totti, C., 2016. The toxic benthic dinoflagellates of the genus *Ostreopsis* in temperate areas: a review. Advances in Oceanography and Limnology 7(1), 1-15.
- Aksnes, D.L., Egge, J.K., 1991. A theoretical model for nutrient uptake in phytoplankton. Mar. Ecol. Prog. Ser. 70, 65–72.

Aligizaki, K., Nikolaidis, G., 2006. The presence of the potentially toxic genera *Ostreopsis* and 540 Coolia (Dinophyceae) in the north Aegean sea, Greece. Harmful Algae 5, 717-730. 541 Barone, R., 2007. Behavioural trait of Ostreopsis ovata (Dinophyceae) in Mediterranean rock 542 pools: the spider's strategy. Harmful Algae News 33, 1-3. 543 Bendschneider, K., Robinson, R.J., 1952. A new spectrophotometric method for the 544 determination of nitrite in sea water. J. Mar. Res. 11, 87–96. 545 Biré, R., Trotereau, S., Lemée R., Oregioni, D., Delpont, C., Krys, S., Guérin, T., 2015. Hunt 546 547 for palytoxins in a wide variety of marine organisms harvested in 2010 on the French Mediterranean coast. Mar. Drugs 13, 5425-5446. 548 Brescianini, C., Grillo, C., Melchiorre, N., Bertolotto, R., Ferrari, A., Vivaldi, B., Icardi, G., 549 Gramaccioni, L., Funari, E., Scardala, S., 2006. Ostreopsis ovata algal blooms affecting 550 human health in Genova, Italy, 2005 and 2006. Eurosurveillance 11, e060907.3. 551 552 Brissard, C., Herrenknecht, C., Séchet, V., Hervé, F., Pisapia, F., Harcouet, J., Lemée, R., Chomérat, N., Hess, P., Amzil, Z., 2014. Complex toxin profile of French Mediterranean 553 554 Ostreopsis cf. ovata strains, seafood accumulation and ovatoxins prepurification. Mar. Drugs 12, 2851–2876. 555 Burkholder, J.M., Glibert, P.M., Skelton, H.M., 2008. Mixotrophy, a major mode of nutrition 556 for harmful algal species in eutrophic waters. Harmful Algae 8, 77-93. 557 Ciminiello, P., Dell'Aversano, C., Iacovo, ED., Fattorusso, E., Forino, M., Tartaglione, L., 558 Benedettini, G., Onorari, M., Serena, F., Battocchi, C., et al., 2014. First finding of 559 Ostreopsis cf. ovata toxins in marine aerosols. Environ. Sci. Technol. 48, 3532–3540. 560 Clark, D.R., Flynn, K.J., 2002. N-assimilation in the noxious flagellate Heterosigma carterae 561

(Raphidophyceae): dependence on light, N-source, and physiological state. J. Phycol. 38,

562

563

503-512.

Clark, D.R., Flynn, K.J., Owens, N.J.P., 2002. The large capacity for dark nitrate assimilation 564 in diatoms may overcome nitrate limitation of growth. New Phytol. 155, 101–108. 565 Cochlan, W.P., Harrison, P.J., Denman, K.L., 1991. Diel periodicity of nitrogen uptake by 566 marine phytoplankton in nitrate-rich environments, Limnol, Oceanogr, 36, 1689-1700. 567 Cochlan, W.P., Herndon, J., Kudela, R.M., 2008. Inorganic and organic nitrogen uptake by the 568 toxigenic diatom *Pseudo-nitzschia australis* (Bacillariophyceae). Harmful Algae 8, 111-569 570 118. Cohu, S., Mangialajo, L., Thibaut, T., Blanfuné, A., Marro, S., Lemée, R., 2013. Proliferation 571 of the toxic dinoflagellate Ostreopsis cf. ovata in relation to depth, biotic substrate and 572 environmental factors in the North West Mediterranean Sea. Harmful Algae 24, 32-44. 573 Collos, Y. 1987. Calculations of N-15 uptake rates by phytoplankton assimilating one or several 574 nitrogen sources. Appl. Radiat. Isot. 38, 275–82. 575 576 Collos, Y., Siddiqi, M.Y., Wang, M.Y., Glass, A.D.M., Harrison, P.J., 1992. Nitrate uptake kinetics by two marine diatoms using the radioactive tracer ¹⁵N. J. Exp. Mar. Biol. Ecol. 577 578 163, 251-260. 579 Collos, Y., Gagne, C., Laabir, M., Vaquer, A., Cecchi, P., Souchu, P., 2004. Nitrogenous nutrition of Alexandrium catenella (Dinophyceae) in cultures and in Thau Lagoon, 580 Southern France. J. Phycol. 40, 96-103. 581 Collos, Y., Vaquer, A., Souchu, P., 2005. Acclimation of nitrate uptake by phytoplankton to 582 high substrate levels. J. Phycol. 41, 466-478. 583 Collos, Y., Harrison, P.J., 2014. Acclimation and toxicity of high ammonium concentrations 584 to unicellular algae. Mar. Pollut. Bull. 80, 8-23. 585 Dortch, Q., Clayton, J.R., Thoresen, S.S., Bressler, S.L., Ahmed, S.I., 1982. Response of marine 586

phytoplankton to nitrogen deficiency: Decreased nitrate uptake vs enhanced ammonium

587

588

uptake. Mar. Biol. 70, 13-19.

- Escalera, L., Benvenuto, G., Scalco, E., Zingone, A., Montresor, M., 2014. Ultrastructural
- Features of the Benthic Dinoflagellate *Ostreopsis* cf. *ovata* (Dinophyceae). Protist 165,
- 591 260-274.
- Fábregas, J., Maseda, A., Domínguez, A., Ferreira, M., Otero, A., 2002. Changes in the cell
- composition of the marine microalga, Nannochloropsis gaditana, during a light:dark
- 594 cycle. Biotechnol. Lett. 24, 1699–1703.
- Fan, C., Glibert, P.M., Burkholder, J.M., 2003. Characterization of the affinity for nitrogen,
- uptake kinetics, and environmental relationships for *Prorocentrum minimum* in natural
- blooms and laboratory cultures. Harmful Algae 2, 283-299.
- 598 Fan, C., Glibert, P.M., 2005. Effects of light on nitrogen and carbon uptake during a
- 599 *Prorocentrum minimum* bloom. Harmful Algae 4, 629-641.
- Faust, M.A., Morton, S.L., Quod, J.P., 1996. Further study of marine dinoflagellates: the genus
- Ostreopsis (Dinophyceae). J. Phycol. 32, 1053-1065.
- Glibert, P.M., Wilkerson, F.P., Dugdale, R.C., Raven, J.A., Dupont, C.L., Leavitt, P.R., Parker,
- A.E., Burkholder, J.M., Kana, T.M., 2016. Pluses and minuses of ammonium and nitrate
- uptake and assimilation by phytoplankton and implications for productivity and
- community composition, with emphasis on nitrogen-enriched conditions. Limnol.
- 606 Oceanogr. 61, 165-197.
- 607 Gorbi, S., Avio, G.C., Benedetti, M., Totti, C., Accoroni, S., Pichierri, S., Bacchiocchi, S.,
- Orletti, R., Graziosi, T., Regoli, F., 2013. Effects of harmful dinoflagellate *Ostreopsis* cf.
- ovata exposure on immunological, histological and oxidative responses of mussels
- 610 *Mytilus galloprovincialis*. Fish Shellfish Immunol. 35, 941–950.
- 611 Granum, E., Kirkvold, S., Myklestad, S.M., 2002. Cellular and extracellular production of
- carbohydrates and amino acids by the marine diatom Skeletonema costatum: diel
- variations and effects of N depletion. Mar. Ecol. Prog. Ser. 242, 83–94.

- 614 Guidi-Guilvard, L.D., Gasparini, S., Lemée, R., 2012. The negative impact of Ostreopsis cf.
- 615 ovata on phytal meiofauna from the coastal NW Mediterranean. Cryptogamie, Algologie
- 616 33(2), 121-128.
- 617 Guillard, R.R.L., 1973. Division rates. In: Stein, J. R. (Ed.), Handbook of phycological
- methods. Culture methods and growth measurements. Cambridge University Press,
- 619 Cambridge, UK, pp. 289–311.
- Herndon, J., Cochlan, W.P., 2007. Nitrogen utilization by the raphidophyte *Heterosigma*
- *akashiwo*: Growth and uptake kinetics in laboratory cultures. Harmful Algae 6, 260-270.
- Honsell, G., Bonifacio, A., De Bortoli, M., Penna, A., Battocchi, C., Ciminiello, P.,
- Dell'Aversano, C., Fattorusso, E., Sosa, A., Yasumoto, T., Tubaro, A., 2013. New
- 624 insights on cytological and metabolic features of Ostreopsis cf. ovata Fukuyo
- 625 (Dinophyceae): a multidisciplinary approach. PLoS ONE 8 (2): e57291.
- Hu, Z., Duan, S., Xu, N., Mulholland, M.R., 2014. Growth and nitrogen uptake kinetics in
- 627 cultured *Prorocentrum donghaiense*. PLoS ONE 9(4): e94030.
- doi:10.1371/journal.pone.0094030.
- Hurd, C.L., Dring, M. J., 1990. Phosphate uptake by intertidal algae in relation to zonation and
- 630 season. Mar. Biol. 107, 281–289.
- Jauzein, C., Loureiro, S., Garcés, E., Collos, Y., 2008a. Interactions between ammonium and
- urea uptake by five strains of *Alexandrium catenella* (Dinophyceae) in culture. Aquat.
- 633 Microb. Ecol. 53, 271–280.
- Jauzein, C., Collos, Y., Garcés, E., Vila, M., Maso, M., 2008b. Short-term temporal variability
- of ammonium and urea uptake by Alexandrium catenella (Dinophyta) in cultures. J.
- 636 Phycol. 44, 1136–1145.

- Jauzein, C., Collos, Y., Laabir, M., Vaquer, A., 2011. Dark metabolism and carbon–nitrogen 637 uncoupling in the toxic dinoflagellate Alexandrium catenella (Dinophyceae). Harmful 638 Algae 11, 73-80. 639 Keller, M.D., Selvin, R.C., Claus, W., Guillard, R.R.L., 1987. Media for the culture of oceanic 640 ultraphytoplankton. J. Phycol. 23 (4), 633-638. 641 Kudela, R.M., Cochlan, W.P., 2000. Nitrogen and carbon uptake kinetics and the influence of 642 irradiance for a red tide bloom off southern California. Aguat. Microb. Ecol. 21, 31–47. 643 Kudela, R.M., Seeyave, S., Cochlan, W.P., 2010. The role of nutrients in regulation and 644 promotion of harmful algal blooms in upwelling systems. Prog. Oceanogr. 85, 122-135. 645 Kwon, H.K., Oh, S.J., Yang, H-S., 2013. Growth and uptake kinetics of nitrate and phosphate 646 by benthic microalgae for phytoremediation of eutrophic coastal sediments. Bioresource 647 Technol. 129, 387-395. 648 649 Li, J., Glibert, P.M., Alexander, J.A., 2011. Effects of ambient DIN:DIP ratio on the nitrogen uptake of harmful dinoflagellate *Prorocentrum minimum* and *Prorocentrum donghaiense* 650 in turbidistat. Chin. J. Oceanol. Limn. 29 (4), 746-761. 651 Litchman, E., Klausmeier, C.A., Schofield, O.M., Falkowski, P.G., 2007. The role of functional 652 traits and trade-offs in structuring phytoplankton communities: scaling from cellular to 653 ecosystem level. Ecol. Lett. 10, 1170–1181. 654 Lomas, M.W., Glibert P.M., 1999. Interactions between NH4+ and NO3- uptake and 655 assimilation: comparison of diatoms and dinoflagellates at several growth temperatures. 656 Mar. Biol. 133, 541–551. 657
- Lomas, M.W., Glibert, P.M., 2000. Comparisons of nitrate uptake, storage, and reduction in marine diatoms and flagellates. J. Phycol. 36, 903-913.

- Lomas, M.W., Trice, T.M., Glibert, P.M., Bronk, D.A., McCarthy, J.J., 2002. Temporal and
- spatial dynamics of urea uptake and regeneration rates and concentrations in Chesapeake
- Bay. Estuaries. 25 (3), 469-482.
- Maguer, J.-F., L'Helguen, S., Madec, C., Labry, C., Le Corre, P., 2007. Nitrogen uptake and
- assimilation kinetics in *Alexandrium minutum* (Dynophyceae): effects of N-limited
- growth rate on nitrate and ammonium interactions. J. Phycol. 43, 295–303.
- Maguer, J.-F., L'Helguen, S., Waeles, M., 2015. Effects of mixing-induced irradiance
- fluctuations on nitrogen uptake in size-fractionated coastal phytoplankton communities.
- 668 Estuar. Coast. Shelf S. 154, 1-11.
- Mangialajo, L., Bertolotto, R., Cattaneo-Vietti, R., Chiantore, M., Grillo, C., Lemée, R.,
- Melchiorre, N., Moretto, P., Povero, P., Ruggieri, N., 2008. The toxic benthic
- dinoflagellate Ostreopsis ovata: Quantification of proliferation along the coastline of
- 672 Genoa, Italy. Mar. Pollut. Bull. 56, 1209-1214.
- 673 Mangialajo, L., Ganzin, N., Accoroni, S., Asnaghi, V., Blanfuné, A., Cabrini, M., Cattaneo-
- Vietti, R., Chavanon, F., Chiantore, M., Cohu, S., Costa, E., Fornasaro, D., Grossel, H.,
- Marco-Miralles, F., Masó, M., Reñé, A., Rossi, A.M., Sala, M.M., Thibaut, T., Totti, C.,
- Vila, M., Lemé e, R., 2011. Trends in *Ostreopsis* proliferation along the Northern
- Mediterranean coasts. Toxicon 57 (3), 408–420.
- Mulholland, M.R., Lomas, M.W., 2008. Nitrogen uptake and assimilation. In: Capone, D.G.,
- Bronk, D.A., Mulholland, M.R., Carpenter, E.J. (Eds.), Nitrogen in the marine
- environment, second ed. Elsevier Inc. Burlington, MA. pp. 303-384.
- Needoba, J.A., Harrison, P.J., 2004. Influence of low light and a light:dark cycle on NO3
- uptake, intracellular NO3, and nitrogen isotope fractionation by marine phytoplankton. J.
- 683 Phycol. 40, 505–516.

- Nishikawa, T., Tarutani, K., Yamamoto, T., 2009. Nitrate and phosphate uptake kinetics of the
- harmful diatom *Eucampia zodiacus* Ehrenberg, a causative organism in the bleaching of
- aquacultured *Porphyra thalli*. Harmful Algae 8, 513-517.
- Paasche, E., Bryceson, I., Tangen, K., 1984. Interspecific variation in dark nitrogen uptake by
- dinoflagellates. J. Phycol. 20, 394–401.
- Pagliara, P., Caroppo, C., 2012. Toxicity assessment of Amphidinium carterae, Coolia cfr.
- 690 monotis and Ostreopsis cfr. ovata (Dinophyta) isolated from the northern Ionian Sea
- 691 (Mediterranean Sea). Toxicon 60, 1203–1214.
- Parsons, ML., Preskitt, LB., 2007. A survey of epiphytic dinoflagellates from the coastal waters
- of the island of Hawai' i. Harmful Algae 6, 658-669.
- Parsons, M.L., Aligizaki, K., Dechraoui Bottein, M.Y., Fraga, S., Morton, S.L., Penna, A.,
- Rhodes, L., 2012. *Gambierdiscus* and *Ostreopsis*: Reassessment of the state of knowledge
- of their taxonomy, geography, ecophysiology, and toxicology. Harmful Algae. 14, 107-
- 697 129.
- 698 Pistocchi, R., Pezzolesi, L., Guidi, F., Vanucci, S., Guerrini, F., Pinna, A., 2014. Inorganic
- nutrients uptake and organic phosphorus utilization by Ostreopsis cf. ovata. In:
- Proceedings of 16th ICHA conference. pp. 147-150.
- 701 Rausch de Traubenberg, C., Soyer-Gobillard, M.O., 1990. Bacteria associated with a
- photosynthetic dinoflagellate in culture. Symbiosis 8, 117–133.
- Rhodes, L., Adamson, J., Suzuki, T., Briggs, L., Garthwaite, I., 2000. Toxic marine epiphytic
- dinoflagellates, Ostreopsis siamensis and Coolia monotis (Dinophyceae), in New
- 705 Zealand. New Zeal. J. Mar. Fresh. 34, 371–383.
- Rhodes, L., 2011. World-wide occurrence of the toxic dinoflagellate genus *Ostreopsis* Schmidt.
- 707 Toxicon 57 (3), 400–407.

- Rossi, R., Castellano, V., Scalco, E., Serpe, L., Zingone, A., Soprano, V., 2010. New palytoxin-
- 709 like molecules in Mediterranean Ostreopsis cf. ovata (dinoflagellates) and in Palythoa
- 710 tuberculosa detected by liquid chromatography–electrospray ionization time-of-flight
- 711 mass spectrometry. Toxicon 56 (8), 1381–1387.
- Scalco, E., Brunet, C., Marino, F., Rossi, R., Soprano, V., Zingone, A., Montresor, M., 2012.
- Growth and toxicity responses of Mediterranean Ostreopsis cf. ovata to seasonal
- irradiance and temperature conditions. Harmful Algae 17, 25-34.
- 715 Selina, M.S., Morozova, T.V., Vyshkvartsev, D.I., Orlova, T.Y., 2014. Seasonal dynamics and
- spatial distribution of epiphytic dinoflagellates in Peter the Great Bay (Sea of Japan) with
- special emphasis on *Ostreopsis* species. Harmful Algae 32, 1-10.
- 718 Selmer, J.-S., Ferrier-Pages, C., Celario, C., Rassoulzadegan, F., 1993. New and regenerated
- production in relation to the microbial loop in the NW Mediterranean Sea. Mar. Ecol.
- 720 Prog. Ser. 100, 71-83.
- Siu, G.K.Y., Young, M.L.C., Chan, D.K.O., 1997. Environmental and nutritional factors which
- regulate population dynamics and toxin production in the dinoflagellate *Alexandrium*
- *catenella*. Hydrobiologia 352, 117–140.
- Smayda, T.J., 1997. Harmful algal blooms: Their ecophysiology and general relevance to
- phytoplankton blooms in the sea. Limnol. Oceanogr. 42 (5), 1137-1153.
- Syrett, P.J., 1981. Nitrogen metabolism of microalgae. Can. Bull. Fish. Aquat. Sci. 210, 182-
- 727 210.
- 728 Taniyama, S., Arakawa, O., Terada, M., Nishio, S., Takatani, T., Mahmud, Y., Noguchi, T.,
- 729 2003. Ostreopsis sp., a possible origin of palytoxin (PTX) in parrotfish Scarus ovifrons.
- 730 Toxicon 42, 29–33.

731	Taylor, B.W., Keep, C.F., Hall, R.O., Jr., Koch, B.J., Tronstad, L.M., Flecker, A.S., Ulseth,
732	A.J., 2007. Improving the fluorometric ammonium method: matrix effects, background
733	fluorescence, and standard additions. J. N. Am. Benthol. Soc., 26(2), 167-17.
734	Tian, R.C., 2006. Toward standard parameterizations in marine biological modeling. Ecol
735	Model. 193, 363–386.
736	Totti, C., Accoroni, S., Cerino, F., Cucchiari, E., Romagnoli, T., 2010. Ostreopsis ovata bloom
737	along the Conero Riviera (northern Adriatic Sea): Relationships with environmental
738	conditions and substrata. Harmful Algae 9:233-239.
739	Turpin, D.H., 1991. Effects of inorganic N availability on algal photosynthesis and carbon
740	metabolism. J. Phycol. 27, 14–20.
741	Uchida, H., Taira, Y., Yasumoto, T., 2013. Structural elucidation of palytoxin analogs produced
742	by the dinoflagellate Ostreopsis ovata IK2 strain by complementary use of positive and
743	negative ion liquid chromatography/ quadrupole time-of-flight mass spectrometry. Rapid
744	Commun. Mass Spectrom. 27, 1999-2008.
745	Varela, D.E., Harrison, P.J., 1999. Effect of ammonium on nitrate utilization by Emiliania
746	huxleyi, a coccolithophore from the oceanic northeastern Pacific. Mar. Ecol. Prog. Ser.
747	186, 67–74.
748	Vila, M., Garcés, E., Masó, M., 2001. Potentially toxic epiphytic dinoflagellates assemblages
749	on macroalgae in the NW Mediterranean. Aquat. Microb. Ecol. 26 (1), 51-60.
750	Vila, M., Abós-Herràndiz, R., Isern-Fontanet, J., Àlvarez, J., Berdalet, E., 2016. Establishing
751	the link between Ostreopsis cf. ovata blooms and human health impacts using ecology
752	and epidemiology. Scientia Marina 80 (S1), 107-115.

Legend of figures

Figure 1. Kinetic curves of NH₄⁺, NO₃⁻ and N-urea uptake for *Ostreopsis* cf. *ovata*. Data points show variations of NH₄⁺, NO₃⁻ and N-urea uptake rates as a function of each respective nutrient concentration for the French strains MCCV 054 (A) and MCCV 055 (B). Detailed representations of variations of N-urea uptake rate, with or without acclimation to N-urea as a N-source, are shown for MCCV 054 (C) and MCCV 055 (D). Respective modeled curves (straight lines for NH₄⁺, dotted lines for NO₃⁻ and dashed lines for N-urea) correspond to Michaelis-Menten model and values of parameters (V_{max} , K_{s} , α), as well as associated r², are listed in Table 1.

Figure 2. Variations in NH₄⁺ and NO₃⁻ uptake rates for the strain MCCV 054 of *Ostreopsis* cf. *ovata* after an addition of 10 μmol N L⁻¹ of NO₃⁻ and along a graded NH₄⁺ concentration of 0 to 10 μmol N L⁻¹ (A), or after an addition of 10 μmol N L⁻¹ of NH₄⁺ and along graded NO₃⁻ concentration of 0 to 10 μmol N L⁻¹. The modeled curves of NH₄⁺ and NO₃⁻ uptake data correspond to solid and dotted lines, respectively. Values of parameters used for modelling these data sets are listed in Table 1 with associated r² values.

Figure 3. Variations in 15 N isotopic ratios of *Ostreopsis* cf. *ovata* cells over 24h, after addition of 100 μmol N L⁻¹ of 15 NH₄⁺ or 15 NO₃⁻. The dark period is indicated by the horizontal solid line. Vertical lines indicate standard deviations from three replicate cultures. Linear regressions of data sets are represented by solid lines and dashed lines for incubations with 15 NH₄⁺ and 15 NO₃⁻, respectively. For the first light period, equations of linear regressions are Y = 0.0318 X – 0.3111 (2 = 0.99) for 15 NH₄⁺ incubations and Y = 0.0300 X – 0.3415 (2 = 0.99) for 15 NO₃⁻ incubations. For the dark period, equations of linear regressions are Y = 0.0059 X + 0.2558 (2 = 0.99) for 15 NH₄⁺ incubations and Y = 0.0029 X + 0.2421 (2 = 0.78) for 15 NO₃⁻ incubations.

Table 1. Values of kinetic parameters (V_{max} in h⁻¹, K_{s} in μ mol N L⁻¹, α in L μ mol N⁻¹ h⁻¹) and inhibition parameters ($V_{\text{max-N=0}}$ in h⁻¹, K_{I} in μ mol N L⁻¹, I_{max}) obtained for *Ostreopsis* cf. *ovata* under various culture conditions. Uptake abilities were characterized for three potential N-sources (NH₄⁺, NO₃⁻ and N-urea), added under graded nutrient concentrations (listed in μ mol N L⁻¹), as a unique N-source or in combination with another one. Non-linear regressions were based on the Michaelis-Menten model or the reverse Michaelis-Menten relation.

Experiment (Strain)	¹⁵ N-addition	on	Kinetic parameters			r^2
	N-source	Concentration	V_{max}	$K_{\rm s}$	α	
Kinetics (MCCV 054)	$\mathrm{NH_4}^+$	0.1 - 10	0.020	0.40	0.0114	0.92
	NO_3^-	0.1 - 10	0.006	0.80	0.0023	0.94
	N-urea	0.1 - 10	0.0005	0.28	0.0003	0.82
Acclimation to N-urea	N-urea	0.1 - 10	0.0005	0.11	0.0004	0.93
Kinetics (MCCV 055)	$\mathrm{NH_4}^+$	0.1 - 10	0.021	0.58	0.0098	0.89
	NO_3^-	0.1 - 10	0.010	3.73	0.0011	0.96
	N-urea	0.1 - 10	0.0004	0.38	0.0002	0.96
Interaction NH ₄ ⁺ /NO ₃ ⁻ (MCCV 054))					
NH_4^+ gradient, $10 \mu M NO_3^-$	$\mathrm{NH_4}^+$	0.1 - 10	0.034	0.72	0.0139	0.95
NO_3^- gradient, 10 $\mu M NH_4^+$	NO_3^-	0.1 - 10	0.008	2.77	0.0012	0.99
			Inhibition p	arametei	rs	
			$V_{\text{max-N=0}}$	K_{I}	I_{max}	

 NH_4^+ gradient, $10 \ \mu M \ NO_3^ NO_3^-$ 10 0.013 6.24 0.67 0.91

Table 2. Estimations of ¹⁵N-uptake rates (in h⁻¹) of Ostreopsis cf. ovata cells in cultures under both light and dark phases of the diel cycle.

Uptake abilities were characterized for NH₄⁺ and NO₃⁻ under N-sufficient conditions.

Experiment (Strain)	¹⁵ N-addition		Mean uptake rate	r ² (p value)	
	N-source Concentration				
Diel cycle (MCCV 054)					
First light period	$N{H_4}^+$	100	0.032	$0.99 \ (p < 0.001)$	
	NO ₃ -	100	0.030	$0.99 \ (p < 0.001)$	
Dark period	$N{H_4}^+$	100	0.006	$0.99 \ (p = 0.06)$	
	NO_3^-	100	0.003	$0.78 \ (p = 0.12)$	

Table 3. Values of kinetic parameters (maximal uptake rate V_{max} in h⁻¹ and half-saturation constant K_s in μ mol N L⁻¹) reported for phytoplankton species from culture experiments with N-sufficient cells. Ranges of nutrient concentration used for kinetics are indicated in μ mol N L⁻¹. Most of the references listed reported results obtained for actively growing N-replete cells. Some experiments were conducted from recently N-depleted cultures at the end of the growth phase and are indicated with a footnote.

Species	Substrate	NH ₄ ⁺		NO ₃		N-urea		Reference	
	concentration	V_{max}	$K_{\rm s}$	$V_{ m max}$	$K_{\rm s}$	V_{max}	K_{s}		
Diatoms									
Chaetoceros sp.	0.01 - 40			0.110	3.10			Lomas and Glibert (2000)	
Pseudo-Nitzschia australis	0.1 - 40	0.071	5.37	0.105	2.82	0.0300		Cochlan et al. (2008) ^{ab}	
Skeletonema costatum	0.01 - 40			0.100	0.40			Lomas and Glibert (2000)	
Thalassiosira weissflogii	0.01 - 40			0.170	2.80			Lomas and Glibert (2000)	
Dinoflagellates									
Alexandrium catenella	0.1 - 10	0.002 - 0.026	0.1 - 6.2			0.0004 - 0.001	0.6 - 2.3	Jauzein et al. (2008b) ^c	
Alexandirum minutum	0.1 - 30		0.33		0.28			Maguer et al. (2007) ^d	
Ostreopsis cf. ovata	0.01 - 10	0.021	0.49	0.008	2.27	0.0005	0.33	Present study	
Prorocentrum minimum	0.01 - 40			0.050	5.00			Lomas and Glibert (2000)	
	0.4 - 30		2.48		5.18		1.82	Fan et al. (2003)	
	0.2 - 20	0.046	1.25			0.0004	0.05	Li et al. (2011)	
Prorocentrum donghaiense	0.1 - 50	0.075	7.10			0.0400	0.12	Hu et al. (2014)	

Haptophyte

	Pavelova lutheri	0.01 - 40			0.120	22.70			Lomas and Glibert (2000)
	Chlorophyte Dunaliella tertiolecta	0.01 - 40			0.030	11.10			Lomas and Glibert (2000)
709	Raphydophyte Heterosigma akashiwo	0.1- 12	0.028	1.44	0.018	1.47	0.0029	0.42	Herndon and Cochlan (2007) ^b
798									

799

800

Footnotes

- Non-saturating kinetics were observed for N-urea uptake; the V_{max} value indicated for N-urea uptake of P. asutralis in this table corresponds to
- the uptake rate estimated at 36 μ mol N L⁻¹.
- b Experiments were conducted on recently N-depleted cultures, in late growth phase.
- ^c Reported values correspond to incubations done just after and 3h after resuspension of N-replete cells in -N medium.
- 805 d Reported values correspond to results obtained for the higest growth rate, when μ/μ max = 0.42.





