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- 1 Methods for studying earthworm dispersal
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- 11 Abstract
- Dispersal is a key driver of species composition and functional traits in earthworm
- communities. However, it has been largely overlooked in ecological literature on earthworms
- because it is particularly difficult to study. In this publication, we review recent developments
- that have been made in this field of research. We present methods to assess dispersal distance,
- such as Capture-Mark-Recapture and molecular tools, and methods using dispersal corridors
- or X-ray imagery aiming at identifying the mechanisms triggering dispersal in earthworms
- 18 communities.

- 20 Keywords
- 21 Movements, landscape structure, connectivity, behavior, capture mark recapture methods

Introduction

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Dispersal plays a major role in shaping biodiversity, evolution, and ecosystem functioning. It connects localities together through fluxes of individuals and alleles. The direct consequence is that species abundance and genetic composition in different places of a landscape are not independent. In other words, local population abundance, genetic structure and community structure not only depend on local factors and processes such as habitat features, demography, genetic drift or species sorting and competition, but are also dependent on the properties of neighboring populations and communities (Leibold et al., 2004). In this perspective, we need to study local community and genetic structure at both local and regional scales in order to understand the structure of local populations or communities, as well as their functional role. The magnitude of the dependence between local and regional scales directly results from dispersal rate. Theoretically, when dispersal is very high, local sites are well interconnected and tend to behave like a unique population or community (Economo and Keitt, 2008; Mouquet and Loreau, 2003). Local interactions are then a major driver of species composition and a low genetic differentiation among populations is expected. When dispersal is very low, local sites are isolated and behave like islands. Local populations and communities are well differentiated and severe genetic drift occurs. Extinction risk is then high for small populations. In nature, dispersal is generally between these two extremes, and complex behaviors such as source – sink dynamics may occur. In such case, certain sites can act as sources of dispersers (source), while others behave like sinks. Such dynamics can prevent extinction or speciation in certain sites. Local populations connected by dispersal are called "metapopulation" (Hanski and Gilpin, 1997), local communities connected by dispersal are called "metacommunity" (Leibold et al., 2004). In both cases, we need to understand the

drivers and the magnitude of dispersal in order to understand the behavior and the properties

of the system, either at local and regional scales.

Dispersal is usually defined as the movement of individuals away from their natal habitat, or

from their usual home range, to a new habitat (Clobert et al., 2012). It is usually decomposed

in three successive steps. First step is departure from the usual home range. Second is the

transfer between the departure site and the arrival site. Third is the establishment in a new

habitat (Fig. 1).

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Figure 1

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All these steps can originate from two very distinct processes: it can come from individuals'

own willing, a process called "active dispersal", or from movements driven by an external

force such as wind, water runoff, displacement by another animal, or by human activities

(Matthysen, 2012). Dispersal direction is generally controlled in active dispersal but not in

59 passive dispersal.

Studying dispersal is challenging for all organisms (Nathan, 2001), because it is hard to track

individuals. This is particularly true for earthworms, because they are subterranean and cannot

be seen from surface. Several approaches have been developed to address these difficulties,

and can be classified in two groups. The first ones focus on dispersal patterns. They intend to

measure typical dispersal distances of organisms during life span, or during a precise period

of time. This approach aims at producing a histogram of distances travelled over a period of

time, the so called "dispersal kernel" (Nathan et al., 2012). It is thus centered on the spatio-

temporal aspects of the dispersal. The second group of methods focuses on the factors that

- drive dispersal, such as habitat quality or conspecific density. The products of this approach 68 are dispersal rules, like positive density dependent dispersal. This approach is often framed in 69 game theory and evolutionary ecology. It aims at predicting dispersal behavior and 70 understanding the reasons why different dispersal behaviors evolved. 71
- 72 In this work, we present techniques that are currently available to study earthworms' dispersal from these two angles: 1) dispersal distances and 2) the factors that drive dispersal. 73

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1. Methods for studying earthworm dispersal distance

- One of the most basic question regarding dispersal of organisms is to determine how far they can move. In order to address this point, we need to estimate the distribution of the distances travelled for a given period of time. For this, two kinds of movements are often defined: the usual and most frequent movements, related to foraging, and the rare long distances movements (Nathan et al., 2008). In this framework, true dispersal usually refers to the rare and long distances dispersal events (LDD), located in the tail of the dispersal distribution(Nathan et al., 2012). These rare events play a critical role in colonization processes and invasion processes (Trakhtenbrot et al., 2005).
- The estimation of traveling distances is challenging for earthworms, but can be done in two 84 broad ways: by tagging and by studying the genetic profiles of individuals or populations. 85

1.1 Tags

Methods based on tags usually share the same principle: a group of individuals is captured, 87 88 marked with a tag, then released at a known location in the field or in an experimental device. After a period of time, the marked individuals are searched at proximity from the release point. The distribution of the recapture distances from the release point (Fig. 3A) gives the

"dispersal kernel" of the individuals over the respective period of time. These methods are generally referred to as Capture Mark Recapture (CMR) methods (Seber, 1982). They are better suited for short distances estimation than for LDD. In theory, individuals can be released and recaptured several times, but this is difficult to achieve in the field on earthworms. Variations in CMR methods include the type of tags, the capacity to use tags specific to individuals, and the number of recapture events. In earthworms, several types of tags have been used so far. Early attempts used food colorants such as E102 + E132 (green) or E124 (red) and red biological stains such as Safranin and Phloxine (Mazeaud, 1979). More recently, ¹⁵NH4+ and U-¹³C glucose (Dyckmans et al., 2005), Rubidium (Ben Hamou et al., 2007), or ⁶⁰Co (Capowiez et al., 2001) have been used with success. However, all these methods are difficult to apply because they require a significant equipment or lab work to detect the tags, because they are not visible with naked eyes. We now present in detail two recent methods that circumvent these caveats.

1.1.1 Visual Tags for earthworm: VIE

Visible Implant Elastomers (VIE, Fig. 2A&B and 3A) are colored tags that are injected below the skin (Butt and Lowe, 2007; Gonzalez et al., 2006). They are injected in a liquid state but soon become solid through a polymerization process. They are available from Northwest Marine Technology, Inc (http://www.nmt.us/) and are relatively affordable.

Figure 2

This kind of tags has been used with success on a variety of organisms such as fish, frogs, turtles and seeds, in order to assess population size or dispersal distance. They are visible

through the skin of the animals, particularly with a UV light. The tags are well supported by earthworms (Butt et al., 2009). They can stay up to 27 months in their body, which is much longer than the previously mentioned tags (Butt et al., 2009).

Figure 3

The main drawback of this method is the impossibility of monitoring individuals in a continuous manner. It is not possible to track individuals between two capture events.

Marking and recapturing require a lot of time, and tags are hardly specific to individuals, even though combining tags of different colors is possible for large individuals (>1.5g, Fig. 2B).

This method requires releasing a high number of marked animals for statistical reasons, which can trigger density dependence dispersal (McCrea and Morgan, 2014). Indeed, in many species, high level of conspecific density leads to active dispersal (Caro et al., 2013; Mathieu et al., 2010). Finally, statistical methods for capture recapture data are complex (Amstrup et al., 2005; Lee and Chao, 1994). Despite these difficulties, this approach is very useful to get an estimation of the dispersal kernel of individuals.

1.1.2 Electronic tags: RFID tags

A new promising technique is the use of miniaturized RFID tags (Radio Frequency IDentification. These tags (Fig.4) offer the possibility to mark each individual specifically, with a unique barcode ID. A scanner (Fig. 4B) retrieves the ID of an individual when the tag is close enough from the receptor, typically less than 0.5 cm.

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Detecting individuals in a continuous fashion can be achieved by installing antennas at the surface of the ground, which detect any individual with a tag that comes close to the antenna. This offers the possibility to track individuals continuously and at different sites. This method suffers from several limitations that still impede its use on a regular base. First, RIFD tags are more harmful than VIE tags. They are bigger (1.4 x 8mm) and are frequently ejected out of the body by earthworms. During preliminary tests, only larger individuals, typically >1g, supported them. Second, detection can only be done at very short distance from the antenna (<0.5cm), and cannot be done through the soil. Even if the antenna can be buried, it is much more efficient to detect individuals on the surface. Thus, this technique is better suited for large epigeic and anecic species, which crawl on surface, but not for endogeic ones. Third, the tags are completely invisible once injected, which makes it difficult to find their location in the body during manual scanning, or to determine if an individual is tagged or not. A VIE can be injected in addition, close to the RIFD tag, however this increases the risk of mortality. At the moment, miniaturized RFID can only be detected by the scanners provided by the RFID manufacturers, which are expensive and have a limited efficiency. However, in the future it should be possible to build its own system of scanners or antenna and data logger.

1.2 Molecular approach

Recent progresses in molecular biology offer new opportunities for the estimation of dispersal distance of earthworms from genetic data (Dupont, 2009; Torres-Leguizamon et al., 2012; Zeller et al., 2012). As dispersal leads to gene flow, genetic information can be used to infer dispersal patterns. They can be assessed in two ways: first by comparing observed

populations' genetic structure to theoretical ones under no dispersal, and second by statistical assignment methods that allow identifying the parents or the population of origin of individuals, based on their genetic profile (Broquet and Petit, 2009). In addition, the effect of landscape structure on dispersal patterns can be assessed with the tools developed within the framework of landscape genetics (Manel et al., 2003). The principle is to determine if there is any relationship between the spatial patterns of populations' genetic structure, or individuals' genetic profile, and landscape features. This approach was applied at the plot scale (100x50m) on the earthworm *Allolobophora chlorotica* (Dupont et al., 2015). It suggested that soil properties influence dispersal in the field, and estimated that gene flow occurs at approximately 7 m.year-1. At the landscape level, it showed the role of landscape features on the dispersal of two earthworm species, *A. chlorotica* and *Aporrectodea icterica* (Dupont et al., 2017).

Figure 5

The genetic approach is based on the variation of highly variable DNA sequences (i.e. molecular markers) which are selectively neutral (i.e. the variation is not selected by the environment). Microsatellites, Amplified fragment length polymorphisms (AFLPs), and single nucleotide polymorphisms (SNPs) are typical markers used for this kind of study (Dupont et al., 2011; Torres-Leguizamon et al., 2012). Because population genetic models are dedicated to diploid (2N) species, this kind of approach can hardly be used for studying dispersal of earthworm species with a ploidy level different to 2N. Once the molecular data are acquired, they can be matched with landscape data in two ways. First, they can be compared to landscape structure in the neighborhood of the sampling sites (Fig. 5). For this,

an area (a "buffer") is defined around each sampling site, and the relevant features of landscape structure are described with the help of a GIS. For instance, the distance to the nearest forest, the percentage of crops or the length of roads can be computed (McGarigal and Marks, 1995). The result of this process is a table describing different landscape features for each sampling site. These features can then be matched with genetic descriptors of populations, such as allelic richness and heterozygote frequency, using, for instance, traditional regression techniques. The second way to proceed is through connectivity analysis. In this case, the landscape is considered in its totality, not only around a certain radius of each sampling site. In this approach, all sites are compared by pairs, both in terms of genetic difference and geographical distance (Fig. 5). This is done by correlating the geographical distance between pairs of sites to their genetic distance. If there is any significant relationship between the two matrixes, it suggests that landscape feature, and thus dispersal, has an influence on spatial genetic structure. Geographical distance is calculated by estimating the cumulated costs of movements between each pairs of sites. For this, specific costs are attributed to the different elements of the landscape. The product of this step is a "resistance map". This map is used to calculate the cumulative traveling costs between each pair of sites, using different algorithms of movements. Finally, either the cost of the shortest path, or the weighted mean of the different paths, is used to estimate the distance between all pairs of sites (McRae et al., 2008). The final product of this step is a pairwise squared distance matrices giving the traveling costs between pairs of sites. This process can be realized with different scenario of costs. By comparing different scenarios of costs, this technique allows identifying which elements are most likely to play a role on dispersal, either by facilitating it (corridors) or by impeding it (barriers) (Dupont et al., 2017). In order to match the distance matrix with genetic features, the genetic differentiation between all sampling sites must also be estimated. This can be done through different indices, but the most widespread is the fixation index Fst

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(Wright, 1951). The result of this step is a pairwise squared dissimilarity matrix between sites. In order to correlate the matrix of genetic differentiation to the matrixes of traveling costs, specific statistical techniques must be used, because the same sites are involved in several comparisons, and there is non independency in the residuals. The historical technique used for this was Mantel test, but it has been criticized and the recommend test is now MLPE (Maximum Likelihood Population Effect). This technique identifies the scenario of costs that match with differences in genetic structure (Van Strien et al., 2012)

2. Methods for studying mechanisms of earthworm dispersal

Another important aspect in the study of dispersal is to determine the conditions that drive individuals to disperse. These drivers can be internal, such as body size, hormonal status and age, or external, such as habitat quality or biotic interactions (Clobert et al., 2012). In earthworms, only external drivers of dispersal have been studied so far. Two techniques are now well established: the dispersal corridors (Mathieu et al., 2010), and the X ray imagery (Caro et al., 2012).

2.1 Dispersal corridors

The principle of dispersal corridors is to expose individuals to a stimuli, to wait a certain amount of time, and then to compare the movements of individuals in the stimuli treatment to the control (Clobert et al., 2012). In this kind of experiment, it is necessary to be able to make the difference between active dispersal and random diffusion. Active dispersal results from a decision to leave a patch, without knowing in advance where to go, and what type of environment will be available after leaving the initial patch. Random diffusion results from

usual movements, such as foraging, that do not imply a decision to leave. To be able to make the distinction between active dispersal and random diffusion, a specific experimental setup must be used (Fig. 6).

Figure 6

The experimental device is elongated and separated in three parts: 1) the "departure section"; 2) the "crossing section"; 3) the "settlement section". The "departure section" is the section where the stimuli or the control treatment is implemented, and where the earthworms are inoculated. The treatment can consist for example in soil quality, litter quality, conspecific density, or interspecific density. The crossing section must be an unhospitable soil, which individuals would cross only if the conditions in the departure section are really unhospitable. This is the key point to make the distinction between random diffusion and active dispersal. The "settlement section" must contain a very hospitable soil, which individuals will not leave once they reached it. In top of that, it is required to put litter on the surface of the departure and of the settlement zones, even for endogeic species. After a fixed period of time after inoculation, the soil and the litter of the tree sections are separated and the number of earthworms in each section is retrieved. The percentage of dispersal is compared between control and stimuli treatment with a logistic model (Caro et al., 2013).

1.3 X ray imagery

Another promising tool is X ray imagery. This method takes advantage of recent progress in imagery machines. It consists in using X ray imaging device such as Philips Diagnostic C

generator to scan the soil (Fig. 7). This kind of device is capable of differentiating earthworms and their gallery in the soil, if the soil is less than 6 cm thick. The soil must not be too mineral otherwise X ray cannot penetrate. Adding soft material, such as peat, greatly increase image quality. Introducing by incision a small metal tag (1mm) below the skin of individuals can also help to localize them on the X ray shots. With this kind of setup, it is possible to determine in real time the location of individuals in the ground (Caro et al., 2012), to tell if it is located inside or outside a gallery, and to measure its movement speed [VIDEO].

Figure 7

Conclusion

The several techniques that now exist to study earthworm dispersal (Table 1) can give us insights about dispersal distances and mechanisms leading to active dispersal, either in the field or in experimental devices. They offer new opportunities to quantify this process, and to estimate its contribution to population dynamics, community assembly, and finally ecosystem functioning.

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273 Soil Ecology

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Figure captions

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- Fig. 1 Dispersal is usually decomposed in three steps: Emigration from the source, transfer,
- and immigration to a new site.
- Fig. 2 A/ Tagging of a small *Allolobophora chlorotica* B/ An individual of *Aporrectodea*
- 366 *longa* tagged with two yellow bands.
- Fig. 3 Summary of the steps required to study earthworm dispersal with tags.
- 368 Fig. 4 A/ Picture of a RFID tag ready to be injected. B/ The scanner recognizes a tagged
- 369 *Allolobophora chlorotica* individual.
- Fig. 5 Overview of the workflow to perform landscape genetics on earthworm.
- Fig. 6 Example of a dispersal corridors with its three sections. The triangle indicates the
- inoculation point of earthworms.
- Fig. 7 Experimental setup to take X ray shots of earthworms in the soil.

Table 1 Comparison of the methods currently available for studying earthworm dispersal.

Method	Focus on dispersal	Type of method	Estimation of dispersal	Type of dispersal	Spatial extent †	Place	Cost	Computing complexity	Time required ‡	References §
Landscape genetics	Distance	Molecular	indirect	passive	++ to +++	Field	\$\$\$\$	+++	+++	1)
VIE Tags	Distance	Capture Mark	direct	active	+ to ++	Field &	\$\$	++	++	2)
		Recapture				Lab				
RFID Tags	Distance	Capture Mark	direct	active	+ to ++	Field &	\$\$\$	++	++	3)
		Recapture				Lab				
Dispersal corridors	Mechanisms	Experimental	direct	active	+	Lab	\$	+	+	4)
X rays	Mechanisms	Experimental	direct	active	+	Lab	\$\$	+	+	5)

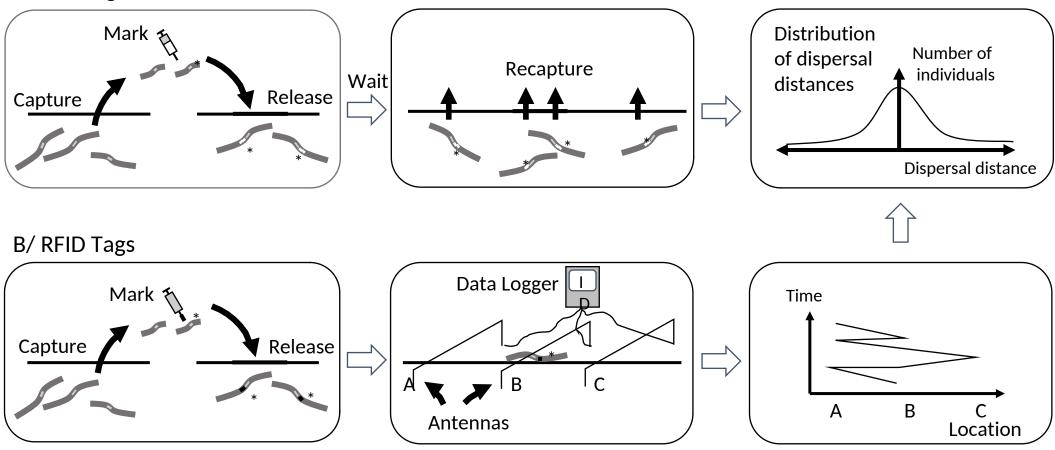
^{†: + =} small extent, +++ = large extent, ‡: + = least, +++ = most, § References : 1) Dupont, L. and al., 2017. Mol Ecol 26, 3128-3140. 2) Butt, K.R., Lowe, C.N., 2007. Appl. Soil Eco. 35, 454-457, 3) unpublished, 4) Mathieu J. and al., 2010. Soil Biol. Biochem. 42, 203-209. 5) Caro and al., 2012. Biol. Fert. Soils 48, 961-

DISPERSAL 1° Emigration 2° Transfer mmigration





A/ VIE Tags







Molecular analysis

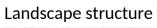
Earthworm sampling





Genetic structure of populations Or Genetic profile of individuals

Landscape Analysis



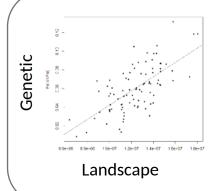


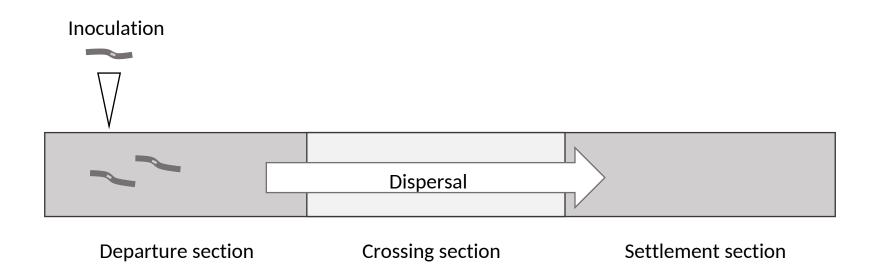
Landscape connectivity

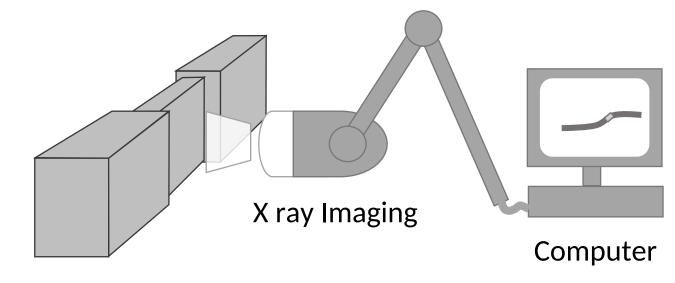


Statistical analysis









Experimental device