

Levels and limits in artificial selection of communities

Manuel Blouin, Battle Karimi, Jérôme Mathieu, Thomas Lerch

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1 TITLE

2 Levels and limits in artificial selection of communities.

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4 AUTHORS

5 Manuel Blouin, Battle Karimi, Jérôme Mathieu, Thomas Z. Lerch

6

7 AFFILIATIONS

- 8 Manuel Blouin : Université Paris-Est Créteil Val-de-Marne (UPEC, UPMC, CNRS, IRD, INRA, Paris
- 9 Diderot), Institute of Ecology and Environmental Sciences of Paris (UMR 7618), 61 avenue du
- 10 Général de Gaulle, 94010 Créteil, France, <u>blouin@u-pec.fr</u>
- 11 Battle Karimi: Laboratoire Chrono-environnement, UMR 6249, Université de Franche-Comté, 16

12 route de Gray, 25000 Besançon, France, <u>karimi.battle@gmail.com</u>

- 13 Jérôme Mathieu : Université Pierre et Marie Curie (UPEC, UPMC, CNRS, IRD, INRA, Paris Diderot),
- 14 Institute of Ecology and Environmental Sciences of Paris (UMR 7618), 7 quai Saint Bernard, 75005
- 15 Paris, France, jerome.mathieu@upmc.fr
- 16 Thomas Z. Lerch : Université Paris-Est Créteil Val-de-Marne (UPEC, UPMC, CNRS, IRD, INRA,
- 17 Paris Diderot), Institute of Ecology and Environmental Sciences of Paris (UMR 7618), 61 avenue du
- 18 Général de Gaulle, 94010 Créteil, France, thomas.lerch@u-pec.fr

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20 AUTHOR CONTRIBUTIONS

- 21 M.B. conceived the study, directed research, performed some statistical analyses and wrote the article,
- 22 and all authors contributed substantially to revisions. B.K. performed the selection experiment under

- the supervision of T.Z.L. and some statistical analyses under the supervision of J.M. and M.B.; T.Z.L.
- 24 performed the T-RFLP analysis.
- 25

26 CORRESPONDING AUTHOR

- 27 Dr Manuel BLOUIN
- 28 Université Paris Est Créteil
- 29 Institute of Ecology and Environmental Sciences of Paris
- 30 61 avenue du Général De Gaulle
- 31 94010 CRETEIL cedex FRANCE
- **32** Tel : +33 1 45 17 16 14
- 33 Courriel : <u>blouin@u-pec.fr</u>
- 34

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45 Abstract

46 Artificial selection of individuals has been determinant in the elaboration of the Darwinian theory of natural selection. Nowadays, artificial selection of ecosystems has proven its efficiency and could 47 contribute to a theory of natural selection at several organization levels. Here, we were not interested 48 in identifying mechanisms of adaptation to selection, but in establishing the proof of principle that a 49 specific structure of interaction network emerges under ecosystem artificial selection. We also 50 51 investigated the limits in ecosystem artificial selection to evaluate its potential in terms of managing ecosystem function. By artificially selecting microbial communities for low CO₂ emissions over 21 52 53 generations (n = 7,560), we found a very high heritability of community phenotype (52%). Artificial selection was responsible for simpler interaction networks with lower interaction richness. Phenotype 54 55 variance and heritability both decreased across generations, suggesting that selection was more likely 56 limited by sampling effects than by stochastic ecosystem dynamics.

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63 INTRODUCTION

64 Because of the time required for natural selection to occur in nature, artificial selection has been a major argument in the development of the theory of heredity with modification by Darwin, with the 65 deep analysis of pigeon breeding genealogy (Darwin 1859). In the line of Darwin, several 66 67 experimental studies have investigated the effects of artificial selection at the group level, to feed the 68 debate on the level of natural selection (Williams 1966; Lewontin 1970; Wilson 1997). Experiments 69 testing group artificial selection involved beetle populations (Wade 1976, 1977; Craig 1982), plant populations (Goodnight 1985), chicken populations (Craig & Muir 1996), but also two-species beetle 70 71 communities (Goodnight 1990a, 1990b) or multiple-species microbial ecosystems (Swenson et al. 72 2000a, 2000b). Recent research also focusses on the ecological consequences of selection of plant 73 trait-associated microbiomes (Lau & Lennon 2012; Panke-Buisse et al. 2015). Consequences of these 74 results for natural selection in nature have already been discussed (Goodnight & Stevens 1997).

75 In parallel, much of the work done in modern molecular genetics is focused on the genetic basis of 76 organism phenotypes and changes in alleles frequencies associated with selected phenotypes. 77 However, this cannot be the unique focus when dealing with community or ecosystem artificial 78 selection: in addition to variations in gene frequencies, changes in ecosystem or community phenotype 79 could be due to changes in intraspecies interactions among individuals and species composition 80 (Goodnight 2000). A simulation model even demonstrate that ecosystem artificial selection can occur "without genetic changes", i.e. only because of changes in species composition (Penn & Harvey 81 82 2004). Before asking the question of genetic mechanisms involved in the modification of the 83 ecosystem phenotype, it is important identifying the level at which phenotype variance occurs: community, population or individual genes? A first objective of this study was to bring an 84 experimental proof of principle that community structure, especially the structure of interaction 85 networks of communities, are significantly affected during the artificial selection procedure. 86

A second objective was to document how far we can go in changing ecosystem phenotype by artificialselection. Whereas limits in artificial selection have been well described and formalized at the

individual level (Robertson 1960; Hill 1982), the degree to which ecosystem properties may be
improved by artificial selection remains unclear (Goodnight 2000). Two different interpretations of the
nature of variation in ecosystem artificial selection are leading to opposite predictions on the link
between variance and heritability and their consequences on the limits in ecosystem artificial selection.

93 Firstly, it has been well established that, during artificial selection of individual organisms, directional 94 selection by truncation leads to a reduction in phenotypic variance, depending on the intensity of 95 selection i. i = z/p, where z is the ordinate of the normal curve at the truncation point and p is the 96 percentage of selected individuals (or the selection rate). In directional selection, the variance of the 97 individuals selected in the parental generation decreases by a factor of 1-i(i-x), where x is the abscissa 98 of the truncation point of the normal curve (Cochran 1951). Genetic variance can be "used up" by 99 selection in a manner that is proportional to the relative reduction in parental phenotypic variance via 100 the fixation of favorable alleles and the elimination of unfavorable alleles as well as rare alleles by 101 drift (i.e. sampling effect). Exceptions occur when a selected trait involves a very large number of loci 102 (Bulmer 1976). To summarize, a decreased genetic variance should lead to decreased phenotypic 103 variance and consequently to a decrease in heritability and selection efficiency. This could be true for other level of organization such as the ecosystem, in which genetic variance could be "used up" 104 105 through the successive loss of rare alleles, individuals or species. In this case, an observed decrease in 106 the variance of ecosystem phenotype should be interpreted as a loss of genetic diversity by sampling 107 effect; ecosystem phenotype variance and heritability should thus decrease along generations and be 108 positively correlated. The limits in ecosystem artificial selection would thus be determined by the 109 initial genetic diversity, size of the population and intensity of the sampling effect.

Another argument leads to an opposite prediction on nature of the limits in ecosystem selection and the sign of the correlation between variance and heritability. According to Lewontin, there are three conditions needed for selection to occur (Lewontin 1970): (i) there must be phenotypic variance among the different individuals experiencing selection; (ii) this phenotypic variance must be heritable; and (iii) phenotypic differences must be linked with different fitness values. In artificial selection experiments, this third condition is always true, as the breeder/experimenter selects individuals based 116 on phenotypic differences. But Penn et al. pointed out that Lewontin's first and second conditions 117 (variance and heritability) could be at odds as far as ecosystem artificial selection is concerned (Penn 118 2003; Penn & Harvey 2004). Indeed, Swenson et al. did not observe any effect of the size of the 119 sample used to create the offspring generation (6.0 vs. 0.06 g of soil) on ecosystem phenotype 120 variance; they interpreted this result as a proof that the intensity of the sampling effect was not 121 determinant in the variance of the ecosystem phenotype, because its importance should have been 122 lower with large samples than with small ones. Consequently, they proposed that ecosystem variance was determined by the stochastic dynamics of ecosystem or butterfly effect (Lorenz 1993), which 123 124 occurs whatever the importance of initial differences due to the sampling effect (Swenson et al. 125 2000b). For Penn et al. (2003; 2004), the stochastic ecosystem dynamics could potentially reduce the 126 heritability of ecosystem phenotypes because it leads to differences between parental and offspring communities. As a consequence, a high variance in ecosystem phenotype due to the stochastic 127 dynamics of ecosystem would be associated with a low heritability and vice versa. If true, a negative 128 129 correlation should be observed between ecosystem phenotypic variance and heritability. In that case, 130 artificial selection might involve more than a search for ecosystems with desired phenotypic traits; it might also be a selection of ecosystems quickly arriving at stable local equilibria, such that their 131 properties can be reliably transmitted to the next generation. In ecosystem artificial selection, the 132 133 limits to transmission of selected variations would thus rely on the ability for ecosystem dynamics to 134 reach quickly a stable equilibrium, not on sampling effect and consequences on initial genetic 135 diversity.

To test our hypotheses, we repeatedly selected for ecosystems with low CO₂ emissions, over 21 selection events hereafter referred to as "generations". The control (random selection) and selection treatment (selection for low CO₂ emissions) each contained 6 independent lines of 30 communities apiece, which allowed us to test the effects of ecosystem selection in a statistically sound way (Fig. 1). For each generation, we also determined community biomass and carbon assimilation yield. To determine if a part of the community phenotype variance could be due to changes in community structure, we looked for non-random changes in community composition and ecological network 143 structure in lines of the control and selection treatment. In this aim, microbial community composition was determined by characterizing the T-RFLP-defined genetic units present in the last generation. To 144 145 reveal ecological interaction network structure, we used co-occurrence patterns (see Faust & Raes 146 (2012) for a detailed discussion); we built co-occurrence networks using genetic-unit-based correlation 147 matrices and were thus able to explore how the structure of ecological interactions responded to 148 ecosystem selection. To identify if the limits in ecosystem artificial selection were due to sampling 149 effect or ecosystem dynamics, we confronted the two opposite predictions on the sign of the 150 correlation between variance and heritability. In this aim, we calculated the heritability of CO₂ emissions at the ecosystem level by (i) regressing emissions by offspring communities against mean 151 emissions by artificially selected parental communities and (ii) using the breeder's equation 152 153 (Goodnight 2000).

155 MATERIALS AND METHODS

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157 Artificial selection experiment

An initial source ecosystem was obtained from the outlet of the Valenton water treatment plant 158 (France) in April 2012. It was stored at 4 °C for 3 weeks and then, before the experiment, a 250-ml 159 sample was incubated at 25 °C for 12 h to allow for acclimation to experimental conditions. Terminal 160 161 restriction fragment length polymorphism (T-RFLP) analysis (see below) revealed that this sample contained an initial microbial community composed of 55 genetic units (Shannon index = 1.61). The 162 experiment involved a control and a selection treatment, which were each made up of 6 independent 163 lines of 30 communities apiece (Fig. 1). In the control, three communities were randomly chosen from 164 165 each generation of each line; these parental communities were then pooled to produce the offspring communities of the next generation. In the selection treatment, the three communities with the lowest 166 CO₂ emissions were selected, and the same procedure was followed. The experiment spanned 21 167 generations, producing a total of 7,560 communities. Our micro-ecosystems consisted of 50 µl inocula 168 169 taken from the initial source ecosystem (about 10⁵ CFU ml⁻¹) to which 750 µl of sterile liquid medium (1/20 diluted LB: 10 g 1⁻¹ Trypton, 15g 1⁻¹ TSB, 500 mg 1⁻¹ of yeast extract, pH=7.3) was added; they 170 were cultivated in 96-deep-well microplates. In each plate, six control wells were filled with 800 µl of 171 172 sterile liquid medium to detect any contamination. The microplates were incubated at 25 °C in the 173 dark for 24 h; this time period was defined as the generation time. CO₂ emissions and microbial biomass were measured at the end of each 24-hour period. The new generation was created by taking 174 50 µl from each pool of control or artificially selected communities. 175

176

177 Quantification of ecosystem properties

178 As stated above, communities were selected based on their CO_2 emissions. CO_2 measurements were 179 performed using the MicroRespTM system (Campbell *et al.* 2003), adapted for use with aquatic

communities (Tlili et al. 2011). The CO₂ emissions of each community were quantified by examining 180 the relative change in the cresol red indicator dye suspended above each well of the 96-well plates. 181 182 After a 24-hour incubation period, the absorbance of the indicator dye was measured at 570 nm using a microplate reader (Synergy HT, BioTek, USA); the values obtained were converted into estimates of 183 CO₂ emissions using a calibration curve previously established at 25 °C (Lerch et al. 2013). Each 184 community's microbial biomass was estimated using a 200-µl sample. Samples were placed in the 185 186 wells of a 300-µL microplate and their absorbance at 600 nm (Synergy HT, BioTek, USA) was measured to assess the density of bacterial cells (cells ml⁻¹). Cell density was converted to C biomass 187 (referred to as C_{Bio} below, $\mu g \ C \ ml^{-1}$) assuming $10^{-12} \ g$ wet weight per cell, a water content of 70%, 188 and a C content as 40% of dry weight (Bratbak & Dundas 1984). The metabolic efficiency of the 189 microbial communities was estimated using carbon assimilation yield (Y), which was calculated as 190 191 follows:

$$192 \qquad Y = \frac{C_{Bio}}{C_{Bio} + C_{Min}}$$

where C_{Bio} and C_{Min} are the amounts of assimilated and mineralized carbon, respectively (Lerch *et al.* 2007a).

195

196 Community composition

197 The composition of the bacterial communities was determined on the last generation (21th) using T-198 RFLP analysis. The details of the T-RFLP analysis are provided in Supplementary Information and the 199 length of the restriction fragments is provided in **SI Tab. 1**. The richness of the T-RFLP profiles was 200 expressed as the total number of T-RFs, and the evenness of the profiles was estimated using the 201 Shannon index (H') (Shannon 1948), which was calculated as follows:

$$H' = \sum_{i=1}^{s} p_i \ln(p_i)$$

where p_i represents the relative abundance of a given T-RF *i*. Past work has found that these indices are significantly correlated with true soil bacterial community richness and diversity (R² = 0.71, P = 0.05) only when communities contain less than 1,200 species (Blackwood *et al.* 2007), as was the case for our experimental communities.

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208 Statistical Analysis

Statistical analyses were performed using the following variables quantified for each generation: (i) absolute CO_2 emissions of the control and treatment lines; (ii) relative CO_2 emissions of the six selection treatment lines (mean of each treatment line minus the overall control mean); (iii) mean relative CO_2 emissions of the selection treatment (overall treatment mean minus the overall control mean). Biomass and carbon assimilation yield were similarly quantified, transformed, and analyzed.

Data normality was tested using the *nortest* package in R (Gross & Ligges 2012). We found that the residuals of relative CO₂ emissions, biomass, and carbon assimilation yield were normally distributed, but those of the absolute values of these variables were not. The former were consequently analyzed using linear models (*stat* package in R), while the latter were examined using generalized linear mixed models employing restricted maximum likelihood (REML, *nlme* package in R (Pinheiro *et al.* 2013) Time (generation), treatment, and their interaction were the fixed effects; line and microplate position in the incubator were the random effects.

221 The T-RFLP profile data were analyzed using correspondence analysis (ade4 package in R (Chessel et 222 al. 2004) An intergroup constrained analysis coupled with a Monte Carlo permutation test was 223 performed to examine differences in community composition between the control and the selection 224 treatment. The data were also used to build co-occurrence networks (i.e., comprising direct and 225 indirect interactions). Using the abundance of the different genetic units, we calculated Spearman 226 correlation coefficients for different genetic unit pairs and constructed a correlation matrix. Only 227 significantly correlated pairs were included in the networks ($P \le 0.05$) (Barberan *et al.* 2012). Distinct positive and negative co-occurrence networks were built for the control and selection treatment by 228

229 converting the correlation matrix into an adjacency matrix with either positive or negative correlation coefficients. A global network that combined positive and negative co-occurrences was also 230 231 constructed to provide an overview of ecological network structure. In this network, modularity, also 232 called compartmentalization, equals the number of isolated sub-networks. The adjacency matrix was 233 resampled via bootstrapping (boot package (Ripley 1999) in R) with a view to quantifying the 234 variance associated with the estimation of the four following interaction network indices (statnet 235 package (Handcock et al. 2008) in R): (1) average degree (D) is the average number of interactions engaged in by one genetic unit (equal to 0 for an unconnected unit)—it is a good estimate of network 236 237 complexity; (2) average betweenness (B) is the average number of shorter chains going through one 238 node-it can signal the presence of keystone species in the network (from a topological standpoint, 239 i.e., a node with many links); (3) connectance (C) is the proportion of possible links between species 240 that are actually realized-this index links the ecological network's overall structure to the behavior of 241 the genetic units; (4) connectedness (Cd) is the probability that at least one chain exists between any 242 pair of units—it quantifies all the direct and indirect interactions within the network. Index values for 243 the control and selection treatment were compared using a Wilcoxon rank sum test with continuity 244 correction for non-parametric data.

248 Artificial selection

Artificial selection was successful: ecosystems from the selection treatment emitted significantly less CO₂ than those of the control. Over the experiment's 21 generations, CO₂ emitted in selection treatment decreased by 0.253 μ g C ml⁻¹ (-60%) and only by 0.142 μ g C ml⁻¹ (-38%) in the control (SI Fig. 1). Three of the six treatment lines emitted significantly less CO₂ over time relative to the control (SI Fig. 2). Mean overall CO₂ emitted by the selection treatment decreased significantly more than that emitted by the control (R² = 0.23, P = 0.015) (Fig. 2a).

255

256 Ecosystem function

Relative carbon assimilation yield did not change over time in any of the treatment lines (SI Fig. 3) or 257 in the group as a whole ($R^2 = 0.009$, P = 0.29) (Fig. 2b). In contrast, relative biomass production 258 declined over time in each of the selection treatment lines (SI Fig. 4) and in the selection treatment as 259 260 a whole ($R^2 = 0.91$, P < 0.001) (Fig. 2c). The lower CO₂ emissions of the treatment communities were thus due to lower biomass production. The greater R² value for biomass (0.91) as compared with the 261 selected CO_2 emissions (0.23) can be explained in two ways: (i)biomass measurements (made using 262 optical density) were more accurate than the CO₂ measurements (made using the MicroRespTM system) 263 264 (Campbell et al. 2003) or (ii)biomass values tend to integrate temporal variation, unlike CO₂ emissions. 265

266

267 Community structure

The analysis of the genetic composition of the last generation of microbial communities showed that neither the erosion of diversity nor the presence of a single, specific genetic unit could explain the stronger decrease in CO₂ emissions in the selection treatment. Indeed, genetic diversity was similar in the control and selected communities (specific richness of 20 and 18 genetic units and Shannon index of 0.87 and 0.83 for the control and selection treatment respectively). The treatment explained 23% of the variance in community composition (constrained correspondence analysis; axis 1: 27%, axis 2: 20%, Monte Carlo test: P = 0.002) (Fig. 3). Despite an important variation in the different lines within the control and selection treatment, species composition differed significantly between the control and the treatment.

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278 Ecological network structure

279 Microbial interaction networks can only be analyzed in terms of co-occurrence networks (see Faust & 280 Raes (2012) for a detailed discussion of the difference between interaction and co-occurrence networks). . To build the co-occurrence networks in the control and selection treatment, we calculated 281 282 the correlation coefficients for pairs of genetic units and selected those that were significantly positively or negatively correlated (P < 0.05). The total number of such pairs, often referred to as 283 "interaction richness" (Tylianakis et al. 2010), is a measure of interaction diversity, which is positively 284 285 correlated with the rate of ecosystem processes (Snyder et al. 2006; Hoehn et al. 2008). Interaction 286 richness was equal to 54 and 28 for the control and selection treatment, respectively. Connectance, i.e. 287 realized interaction richness reported to potential interaction richness, was also lower in the selection 288 treatment network than in the control for the overall network (-23%, P < 0.0001, Fig. 4e) as for the positive and negative co-occurrence networks (SI Fig. 5e, 6e). Interaction diversity can additionally 289 be expressed by the average degree which reveals the average number of interactions involving each 290 291 node (i.e., an individual genetic unit with at least one significant co-occurrence). Genetic units were 292 involved in 3.2 interactions in the overall control network versus 2.0 in the overall selection treatment network (-35%, P = 0.038) (Fig. 4c); this difference was not significant for the positive and negative 293 294 networks (P = 0.10 and 0.11, respectively, SI Fig. 5c, 6c). Approximately 48 and 46% of co295 occurrences were negative in the control and treatment networks, respectively. Thus, artificial
296 selection decreased interaction diversity but did not affect co-occurrence direction.

297 Another useful information can be obtained from the number of clusters or sub-networks, a network indicia called compartmentalization or modularity. Modularity was similar for the positive and 298 299 negative co-occurrence networks (SI Fig. 5a,b, 6a,b), but the overall control network contained a 300 single cluster and the overall treatment network contained four clusters (Fig. 4a,b), which indicates the 301 presence of isolated microbial groups. Average betweenness specifies the average proportion of 302 centrally located nodes, which are viewed as "hubs" or "keystone species" from a network 303 perspective. The overall selection treatment network contained far fewer hubs than the overall control 304 network (-87%, P = 0.003, Fig. 4d); this also hold for the positive and negative networks (SI Fig. 5d, 305 6d). Finally, connectedness is the probability that at least one chain exists between any pair of units, 306 i.e. it quantifies all direct and indirect interactions within the network. Once again, the overall 307 selection treatment network had lower connectedness than the overall control network (-71%, P < P308 0.0001, Fig. 4f); positive and negative networks showed similar patterns (SI Fig. 5f, 6f). Taken 309 together, the results for these three indices show that the treatment network was formed by the juxtaposition of several small networks, characterized by many isolated compartments and contained 310 311 fewer hubs.

312

313 Ecosystem-level heritability

Heritability has been defined as the covariance between average effects and average excess (Fisher 1930), or as the proportion of total variance that can contribute to a response to selection (Goodnight 2000). Heritability at the ecosystem level can be calculated using the slope of the regression between the mean trait values of the selected parental ecosystems and the trait values of the offspring ecosystems (Goodnight 2000). Accounting for the decrease in CO₂ emissions observed in the control, we found a very high overall heritability (h²) of 51.9±1.4% (mean±s.e.) (P < 2 x 10⁻¹⁶, N = 3600) (Fig. 5). For lines 1 to 6 (N = 600), h² equaled 61.7±3.9% (P < 2 x 10⁻¹⁶), 70.4±4.3% (P < 2 x 10⁻¹⁶), 321 $81.6\pm3.8\%$ (P < 2 x 10⁻¹⁶), $41.3\pm3.8\%$ (P < 2 x 10⁻¹⁶), $32.0\pm3.3\%$ (P < 2 x 10⁻¹⁶) and $32.5\pm3.2\%$ (P < 2 322 x 10⁻¹⁶), respectively (SI Fig. 7). The dramatic differences in h² among lines (32 to 82%) suggested 323 that independent trajectories of the lines resulted in large differences in their selection potential.

To determine the sign of the correlation between heritability and ecosystem phenotype variance, we 324 estimated h² at each generation using the standard breeder's equation: $h^2 = R/S$, where R is the 325 326 response to selection and S is the selection differential (Goodnight 2000). This calculation can yield results greater than 1 when population sizes are small. We found a significant positive correlation 327 between heritability and variance for all the parental ecosystems within a given generation (P = 0.006, 328 329 Fig. 5), even if the relationship was weak ($R^2 = 0.07$). This positive correlation may be due to a time 330 effect because both the variance and heritability of the parental generation decreased over time ($R^2 =$ 331 0.10, P = 0.0007 and R² = 0.04, P = 0.0316, respectively; N = 115,).

332

334 **DISCUSSION**

335

336 Statistically sound evidence that ecosystems can be artificially selected

337 Artificial selection of ecosystems was successful. Decreased CO₂ emissions were observed in both the selection treatment and the control because biomass declined over time, probably due to a dilution 338 339 effect. However, the decrease in emissions was significantly greater in the selection treatment than in 340 the control. Previous studies demonstrating artificial selection of complex ecosystem properties were lacking statistical robustness because of an absence of line replication (Swenson et al. 2000b) or 341 synthetic regression analysis (Swenson et al. 2000a). Here, with a large data set obtained using 7,560 342 ecosystems (12 independent lines of 30 ecosystems each allowed to run for 21 generations), we have 343 344 clearly demonstrated that artificial selection can lead to a statistically significant difference in 345 ecosystem properties.

346

347 Levels of selection

348 Evolution at the ecosystem level has mainly been discussed in terms of sources of variation related to 349 target ecosystem phenotypes. Changes in ecosystem phenotype could be due to modifications in intra-350 or inter-species interactions or species composition (Goodnight 2000; Penn & Harvey 2004). An 351 artificial selection experiment shows that, in two-species communities, artificial selection can lead to 352 the emergence of correlated interspecific responses, which suggests that genetically based interactions among individuals are involved (Goodnight 1990a, 1990b). An individual-based evolutionary 353 simulation model has also confirmed that artificial selection can act on complex communities by the 354 355 way of ecological interactions (Williams & Lenton 2007). By showing that the group's response was 356 not simply the sum of the responses of its individual components, the model revealed that many communities are selected because of selective pressures acting on ecological interactions, not on 357 individual species. 358

359 In our study, we were not able to analyze all possible sources of variance, but our results provide 360 several new pieces of knowledge regarding the nature of variance of ecosystem phenotype. Firstly, 361 microbial ecosystem selection does not select for a single genetic species; indeed, at the end of the 362 experiment, selected communities were not comprised of one or a few species. Instead, species richness was equally high in the selection treatment and the control. Mean numbers of genetic units 363 and Shannon index values were similar for the control and selection treatment. However, rare species 364 365 that could not be detected with our molecular methods might have been impacted. Secondly, changes 366 in community composition are an ecosystem response to selection. Although a large amount of 367 variance resulted from sampling effects and the lines' different evolutionary trajectories, the control and treatment groups nonetheless differed dramatically in community composition and structure (Fig. 368 369 3). A significant effect of artificial selection on microbial community composition has also been 370 observed for plant trait-associated microbiomes obtained in multigenerational selection experiments (Lau & Lennon 2012; Panke-Buisse et al. 2015). Thirdly, specific ecological network patterns, 371 372 especially interaction richness, can be selected for: co-occurrence network analysis showed that 373 ecological network structure was very different in the control versus the selection treatment (Fig. 4), 374 especially with regards to interaction richness. Since lower interaction richness may be associated with 375 lower rates of ecosystem function (Snyder et al. 2006; Hoehn et al. 2008; Tylianakis et al. 2010), it is 376 likely that selection resulted in reduced CO₂ emissions by reducing interaction richness. It is thus 377 unlikely that individual organism-level selection could have acted to shape ecological co-occurrence 378 patterns, which emerge at the community level.

Our results call for epistemological considerations. First, in ecosystem artificial selection experiments, the nature of the selected entities (either community or ecosystem) depends on the amount of abiotic (even if biogenic) material transmitted with the inoculum, and its functional consequences. However, it was not possible to assess the role played by this abiotic material in our experiment. Secondly, the level of the selected unit was methodologically defined by the generation time and the volume of ecosystems. Reproduction is a step of the experimental protocol. This is responsible for a disjunction between the generation time and size of the selected unit as compared with the range of generation times of microorganisms belonging to different species (Fig. 3) and the size of their populations. Nevertheless, artificial selection operates. This argues in favor of an enlarged vision of evolution based on operationally defined functional entities (CO₂ emitting ecosystems in our case), rather than on organism reproduction and their populations (see Bouchard 2014 and Doolittle 2014 for a detailed analysis).

391

392 Limits on the artificial selection of ecosystems

The results of our experiment indicate that variance in the parental generation decreased over time. This finding illustrates that genetic diversity is progressively lost through the fixation of favorable alleles (see Introduction). Variance in ecosystem phenotype could be due to the sampling effect of the community or to the stochastic ecosystem dynamics (see Introduction). Because a decreased variance is expected in the case of an important sampling effect and not in the case of an important stochastic dynamics, the decrease in variance over time suggests that sampling effect is at the origin of ecosystem phenotype variance.

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Heritability calculated as the slope of the regression of CO₂ emissions by offspring communities 401 402 against mean emissions by artificially selected parental communities (Fig. 5) provided a very high 403 heritability value of 52% in the selection treatment, far greater than the 15% found by Goodnight 404 (2000), the only other estimate, to our knowledge. This indicated a strong potential for changing an ecosystem phenotype in the desired direction by ecosystem artificial selection (see also the review by 405 406 Goodnight & Stevens, 1997). In addition, heritability calculated using the breeder's equation declined over time, together with artificial selection efficiency. The reduced probability of improving 407 ecosystem properties over time can be understood in the context of Fisher's geometric model (Fisher 408 409 1930). At the ecosystem level, changes in species ecological interaction or community composition 410 responsible for a given amount of change in ecosystem phenotype can, in fact, be selected for when the level of adaptation is low (i.e., during the first generations of an ecosystem selection experiment). 411

412 However, as fitness increases, such changes have a reduced ability to improve future fitness. Only413 changes causing smaller changes can increase fitness in the later generations.

Both variance and heritability were positively correlated. As explained in the Introduction, if the 414 415 stochastic dynamics of ecosystem were responsible for ecosystem phenotype variance (Swenson et al. 2000a, 2000b; Penn 2003; Penn & Harvey 2004), we should observe a negative correlation between 416 417 variance and heritability. Conversely, by extrapolating Cochran hypothesis (1951) from the individual 418 organism level to the ecosystem level, if the sampling effect was dominant, we should observe a 419 positive correlation between variance and heritability. Our results show a positive correlation, which 420 suggests that stochastic ecosystem dynamics did not have a large effect on heritability and selection 421 efficiency, whereas sampling effect seems to be mainly at the origin of phenotype variance.

Goodnight found that communities that exhibited heritable variation also tended to be small, integrated 422 423 communities in which sampling effects had major consequences (Goodnight 2000, 2011). However, Swenson et al. did not observe any effect of the size of the sample used to create the offspring 424 425 generation (6.0 vs. 0.06 g of soil); they interpreted this result as meaning that initial differences that form the basis of the butterfly effect (Lorenz 1993) can be arbitrarily small. Two explanations might 426 427 account for the disparity between our results and those of Swenson et al. First, sample sizes tested in 428 Swenson et al. experiment (6.0 vs. 0.06 g of soil) could both have been too large to exhibit notable 429 differences due to sampling effects. This hypothesis could be tested by studying a larger range of 430 sample sizes. Second, population size has been identified as placing strong limits on artificial selection 431 at the individual level (Robertson 1960; Roberts 1966), because more non-optimal alleles can be fixed 432 by drift when population size decreases. In our experiment, we observed a decrease in community biomass over time (Fig. 2c), likely to result from a decrease in population size. Over time, the 433 increasing effect of drift could have taken precedence over the ecosystem's stochastic dynamics and 434 435 explain our results. This hypothesis could be tested by choosing longer generation time, to ensure a 436 constant population size and drift effect.

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448 References

449 1.

- 450 Barberan, A., Bates, S.T., Casamayor, E.O. & Fierer, N. (2012). Using network analysis to explore co-451 occurrence patterns in soil microbial communities. *ISME J.*, 6, 343–351.
- 452 2.
- Blackwood, C.B., Hudleston, D., Zak, D.R. & Buyer, J.S. (2007). Interpreting ecological diversity
 indices applied to terminal restriction fragment length polymorphism data: Insights from simulated
 microbial communities. *Appl. Environ. Microbiol.*, 73, 5276–5283.
- 456 3.
- Bouchard, F. (2014). Ecosystem evolution is about variation and persistence, not populations and
 reproduction. *Biol. Theory*, 9, 382–391.
- 459 4.
- Bratbak, G. & Dundas, I. (1984). Bacterial dry-matter content and biomass estimations. *Appl. Environ. Microbiol.*, 48, 755–757.
- 462 5.
- Bulmer, M.G. (1976). Effect of selection on genetic variability Simulation study. *Genet. Res.*, 28, 101–117.
- 465 6.
- 466 Campbell, C.D., Chapman, S.J., Cameron, C.M., Davidson, M.S. & Potts, J.M. (2003). A rapid
- 467 microtiter plate method to measure carbon dioxide evolved from carbon substrate amendments so as to
 468 determine the physiological profiles of soil microbial communities by using whole soil. *Appl. Environ.*469 *Microbiol.*, 69, 3593–3599.
- 470 7.
- 471 Chessel, D., Dufour, A.-B. & Thioulouse, J. (2004). The ade4 package-I- One-table methods. *R News*,
 472 4, 5–10.
- 473 8.
- 474 Cochran, W.G. (1951). Improvement by Means of Selection. In: *Proc. Second Berkeley Symp. Math.*475 *Stat. Probab.*, Second Berkeley Symposium on Mathematical Statistics and Probability. University of
- 476 California Press, Berkeley, pp. 449–470.
- 477 9.
- 478 Craig, D.M. (1982). Group Selection Versus Individual Selection an Experimental-Analysis.
- 479 *Evolution*, 36, 271–282.

480	10.
481 482	Craig, J. V & Muir, W.M. (1996). Group selection for adaptation to multiple-hen cages: Beak-related mortality, feathering, and body weight responses. <i>Poult. Sci.</i> , 75, 294–302.
483	11.
484	Darwin, C. (1859). On the origins of species. John Murray, London, UK.
485	12.
486 487	Doolittle, F.W. (2014). Natural selection through survival alone, and the possibility of Gaia. <i>Biol. Philos.</i> , 29, 415–423.
488	13.
489 490	Faust, K. & Raes, J. (2012). Microbial interactions: from networks to models. <i>Nat. Rev. Microbiol.</i> , 10, 538–550.
491	14.
492 493	Fisher, R.A. (1930). <i>The genetical theory of natural selection</i> . Primary so. Oxford university Press, London.
494	15.
495 496	Goodnight, C.J. (1985). The influence of environmental variation on group and individual selection in a cress. <i>Evolution</i> , 39, 545–558.
497	16.
498 499	Goodnight, C.J. (1990a). Experimental studies of community evolution 1. The response to selection at the community level. <i>Evolution</i> , 44, 1614–1624.
500	17.
501 502	Goodnight, C.J. (1990b). Experimental studies of community evolution. 2. The ecological basis of the response to community selection. <i>Evolution</i> , 44, 1625–1636.
503	18.
504 505	Goodnight, C.J. (2000). Heritability at the ecosystem level. <i>Proc. Natl. Acad. Sci. USA</i> , 97, 9365–9366.
506	19.
507 508	Goodnight, C.J. (2011). Evolution in metacommunities. <i>Philos. Trans. R. Soc. B-Biological Sci.</i> , 366, 1401–1409.
509	20.
510 511	Goodnight, C.J. & Stevens, L. (1997). Experimental studies of group selection: What do they tell us about group selection in nature? <i>Am. Nat.</i> , 150, S59–S79.
512	21.

- 513 Gross, J. & Ligges, U. (2012). Package "nortest": tests for normality. *R Packag. version 1.0-2*.
- 514 22.
- Handcock, M.S., Hunter, D.R., Butts, C.T., Goodreau, S.M. & Morris, M. (2008). Package "statnet." *R Packag.*.
- 517 23.
- Hill, W.G. (1982). Rates of change in quantitative traits from fixation of new mutations. *Proc. Natl. Acad. Sci. USA*, 79, 142–145.
- 520 24.
- Hoehn, P., Tscharntke, T., Tylianakis, J.M. & Steffan-Dewenter, I. (2008). Functional group diversity
 of bee pollinators increases crop yield. *Proc. R. Soc. B-Biological Sci.*, 275, 2283–2291.
- 523 25.
- Lau, J.A. & Lennon, J.T. (2012). Rapid responses of soil microorganisms improve plant fitness in
 novel environments. *Proc. Natl. Acad. Sci. USA*, 109, 14058–14062.
- 526 26.
- Lerch, T.Z., Coucheney, E. & Herrmann, A.M. (2013). Sensitivity of soil microbial catabolic profiles
 to a gradient of carbon inputs: Does the soil organic matter Matter? *Soil Biol. Biochem.*, 57, 911–915.
- 529 27.
- 530 Lerch, T.Z., Dignac, M.F., Barriuso, E., Bardoux, G. & Mariotti, A. (2007a). Tracing 2,4-D
- metabolism in Cupriavidus necator JMP134 with 13C-labelling technique and fatty acid profiling. J.
- 532 *Microbiol. Methods*, 71, 162–174.
- 533 28.
- Lerch, T.Z., Dignac, M.-F., Barriuso, E., Bardoux, G. & Mariotti, A. (2007b). Tracing 2,4-D
- metabolism in Cupriavidus necator JMP134 with 13C-labelling technique and fatty acid profiling. *J. Microbiol. Methods*, 71, 162–74.
- 537 29.
- 538 Lewontin, R.C. (1970). The units of selection. Annu. Rev. Ecol. Syst., 1, 1–18.
- **539** 30.
- 540 Lorenz, E.N. (1993). The Essence of Chaos . University of Washington Press, Seattle, W. A.
- 541 31.
- 542 Panke-buisse, K., Poole, A.C., Goodrich, J.K., Ley, R.E. & Kao-Kniffin, J. (2015). Selection on soil
- 543 microbiomes reveals reproducible impacts on plant function. *ISME J.*, 9, 980–989.
- 544 32.

545 546 547	Penn, A. (2003). Modelling artificial ecosystem selection: A preliminary investigation. In: <i>Advances in Artificial Life</i> , Lecture Notes in Artificial Intelligence (eds. Banzhaf, W., Christaller, T., Dittrich, P., Kim, J.T. & Ziegler, J.). Springer-Verlag, Berlin, pp. 659–666.
548	33.
549 550 551	Penn, A. & Harvey, I. (2004). The role of non-genetic change in the heritability, variation, and response to selection of artificially selected ecosystems. Artificial Life IX., M.I.T. Press, Sussex, England.
552	34.
553 554	Pinheiro, J., Bates, D., DebRoy, S. & Sarkar, D. (2013). R Development Core Team. nlme: Linear and Nonlinear Mixed Effects Models. <i>R Packag. version 3.1-111</i> .
555	35.
556	Ripley, B. (1999). Package "boot." R Package
557	36.
558 559	Roberts, R.C. (1966). The limits to artificial selection for body weight in the mouse I. The limits attained in earlier experiments. <i>Genet. Res.</i> , 8, 347–360.
560	37.
561 562	Robertson, A. (1960). A theory of limits in artificial selection. <i>Proc. R. Soc. B-Biological Sci.</i> , 153, 234–249.
563	38.
564 565	Shannon, C.E. (1948). A mathematical theory of communication. <i>Bell Syst. Tech. J.</i> , 27, 379–423,623–656.
566	39.
567 568	Snyder, W.E., Snyder, G.B., Finke, D.L. & Straub, C.S. (2006). Predator biodiversity strengthens herbivore suppression. <i>Ecol. Lett.</i> , 9, 789–796.
569	40.
570 571	Swenson, W., Arendt, J. & Sloan Wilson, D. (2000a). Artificial selection of microbial ecosystems for 3-chloroaniline biodegradation. <i>Environ. Microbiol.</i> , 2, 564–571.
572	41.
573 574	Swenson, W., Wilson, D.S. & Elias, R. (2000b). Artificial ecosystem selection. <i>Proc. Natl. Acad. Sci. USA</i> , 97, 9110–9114.
575	42.
576 577 578	Tlili, A., Marechal, M., Montuelle, B., Volat, B., Dorigo, U. & Bérard, A. (2011). Use of the MicroResp method to assess pollution-induced community tolerance to metals for lotic biofilms. <i>Environ. Pollut.</i> , 159, 18–24.

- 579 43.
- 580 Tylianakis, J.M., Laliberte, E., Nielsen, A. & Bascompte, J. (2010). Conservation of species
- 581 interaction networks. *Biol. Conserv.*, 143, 2270–2279.
- 582 44.
- Wade, M.J. (1976). Group selection among laboratory populations of Tribolium. *Proc. Natl. Acad. Sci. USA*, 73, 4604–4607.
- 585 45.
- 586 Wade, M.J. (1977). An experimental study of group selection. *Evolution*, 31, 134–153.
- **587** 46.
- 588 Williams, G.C. (1966). *Natural selection and adaptation*. Princeton University Press, Princeton.
- 589 47.
- Williams, H.T.P. & Lenton, T.M. (2007). Artificial selection of simulated microbial ecosystems. *Proc. Natl. Acad. Sci. USA*, 104, 8918–8923.
- **592** 48.
- Wilson, D.S. (1997). Multilevel selection theory comes of age. *Am. Nat.*, 150, S1–S21.
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596 Figure legends

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Experimental methodology. The control and treatment each contained six lines 598 Figure 1 599 of 30 communities, which served as replicates. To create each line, we filled 30 wells of a 96-600 well microplate with a 1/20-diluted LB medium; we then inoculated each well with stock 601 from a natural, complex microbial community. After a 24-hour incubation period, the CO₂ emitted by each community was measured. In the selection treatment lines, the three 602 communities with the lowest CO₂ emissions were then selected and used to inoculate 30 new 603 wells and thus formed the next generation. In the control lines, the three parental communities 604 were randomly selected. This cycle was repeated for 20 generations. 605

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Figure 2 Changes in mean relative community properties in response to selection. 607 608 Differences in community **a**) CO_2 emissions, **b**) carbon assimilation yield, and **c**) biomass production. To determine the effects of artificial selection, the mean values for the control 609 were subtracted from the mean values for the treatment. The R² and P-values were determined 610 using linear models. A regression line is depicted if the slope of the regression was 611 significantly different from zero. Dashed line: control; solid line: selection treatment. See the 612 Supplementary Materials and Methods section for a detailed explanation of how carbon 613 assimilation yield was calculated. 614

615

Figure 3 Community microbial diversity at the end of the selection experiment: 616 a) correlation circle for the correspondence analysis—each box corresponds to one 617 618 T-RFLP-defined genetic unit and **b**) barycenters for the control (C) and selection treatment (S) groups. Each point represents one of the six lines found in each group. The control and 619 620 treatment communities differed significantly in composition (Monte Carlo test: P = 0.002). Note that certain genetic units were present in some but not all selection treatment lines and 621 622 absent from the control lines (e.g., X197, X235, X361) or vice versa (e.g., X114, X209, X240); some were also more common in the treatment than in the control (e.g., X187, X230, 623 624 X364) and vice versa (e.g., X173, X185, X205).

Figure 4 Structure of the overall co-occurrence networks and values of the related 626 indices. The co-occurrence matrices of the T-RFLP-defined genetic units present after 21 627 generations were used to build interaction networks for **a**) the control (C) and **b**) the selection 628 treatment (S). When two dots are connected by lines, it means that the abundances of the 629 genetic units were significantly correlated (Spearman correlation coefficient; N = 6; P < 0.05). 630 The interaction networks were used to calculate c) average degree and d) average 631 betweenness for the two treatments. The networks were bootstrapped (200 random samples 632 from each group's pool of genetic units) to determine e) average connectance and f) average 633 634 connectedness. The values of these indices were compared for the control and selection treatment using Wilcoxon rank-sum tests (employing a continuity correction for non-635 parametric distributions). See the Supplementary Materials and Methods section for a full 636 description of how the indices were calculated. 637

638

Figure 5 Linear regression of the CO₂ emissions of the 30 offspring communities as a function of the mean CO₂ emissions of the three artificially selected parental communities (all generations included). The data were corrected by accounting for the decrease in CO₂ emissions in the control and standardized to get a slope, which was equal to heritability *sensu stricto* (h²). Regression equation: y = 0.52x (P < 2 x 10⁻¹⁶, N = 3600).

644

646 Figures

647 Figure 1



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Figure 3



654 Figure 4



















 $_{\mbox{Fig. 5}}$ Mean \mbox{CO}_2 emissions of the three parental communities