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## High contribution of Rhizaria (Radiolaria) to vertical export in the California Current Ecosystem revealed by DNA metabarcoding

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1                                   **High contribution of Rhizaria (Radiolaria)**  
2                                   **to vertical export in the California Current Ecosystem**  
3                                   **revealed by DNA metabarcoding**

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29 ABSTRACT

30 Passive sinking of particulate organic matter (POM) is the main mechanism through  
31 which the biological pump transports surface primary production to the ocean interior.  
32 However, the contribution and variability of different biological sources to vertical export  
33 is not fully understood. Here, we use DNA metabarcoding of the 18S rRNA gene and  
34 particle interceptor traps (PITs) to characterize the taxonomic composition of particles  
35 sinking out of the photic layer in the California Current Ecosystem (CCE), a  
36 productive system with high export potential. The PITs included formalin-fixed and  
37 'live' traps to investigate eukaryotic communities involved in the export and  
38 remineralization of sinking particles. Sequences affiliated with Radiolaria dominated the  
39 eukaryotic assemblage in fixed traps (90%), with Dinophyta and Metazoa making minor  
40 contributions. The prominence of Radiolaria decreased drastically in live traps, possibly  
41 due to selective consumption by copepods, heterotrophic nanoflagellates and phaeodarians  
42 that were heavily enriched in these traps. These patterns were consistent across the water  
43 masses surveyed extending from the coast to offshore, despite major differences in  
44 productivity and trophic structure of the epipelagic plankton community. Our findings  
45 identify Radiolaria as major actors in export fluxes in the CCE.

46

47

49           The main mechanisms of the biological pump include the gravitational sinking of  
50 particles, the active transport associated with zooplankton, and the mixing and diffusive  
51 transport of dissolved and particulate organic matter (DOM and POM) (Turner, 2015).  
52 Among them, particle sinking is the main process contributing to carbon export and is  
53 responsible for 5-21 PgC y<sup>-1</sup> (Eppley and Peterson, 1979; Laws *et al.*, 2000; Henson *et*  
54 *al.*, 2011). Phytoplankton community structure and food-web processes determine the  
55 fraction of net primary production exported as well as the size and chemical characteristics  
56 of sinking material (Boyd and Newton, 1999; Guidi *et al.*, 2009; Stukel *et al.*, 2011).  
57 The fraction of exported particles that reaches the ocean interior is further controlled by  
58 biotic (mainly microbial and zooplankton) transformations during their downward transit,  
59 which affects remineralization rates and particulate organic carbon (POC) flux attenuation  
60 with depth (Lampitt *et al.*, 1990; Steinberg *et al.*, 2008; Giering *et al.*, 2014).

61           Sinking particles are composed of zooplankton fecal pellets (Steinberg and Landry,  
62 2017), organic aggregates of various source ('marine snow') including mucilaginous  
63 structures of larger plankton (Caron *et al.*, 1986; Alldredge and Silver, 1988) and intact  
64 phytoplankton cells (Martin *et al.*, 2011; Smetacek *et al.*, 2012; Agusti *et al.*, 2015).  
65 Drifting particle interceptor traps (PITs) are the most common approach for quantifying  
66 vertical fluxes of particles (Knauer *et al.*, 1979). However, the material collected in PITs  
67 is often partially degraded and heavily transformed by biological activity, which hampers  
68 its taxonomic identification based on morphological attributes and limits the ability to  
69 distinguish biological sources and export mechanisms. Moreover, export fluxes from  
70 sediment traps do not usually match estimated metabolic demands in the ocean twilight  
71 zone (Steinberg *et al.*, 2008; Burd *et al.*, 2010; Herndl and Reinthaler, 2013), suggesting  
72 that other organic fluxes besides those typically considered from phytoplankton and fecal  
73 pellets contribute significantly to vertical export.

74           Recently developed *in situ* imaging methods have proven useful for identifying  
75 larger particles (Guidi *et al.*, 2009; Bochdansky *et al.*, 2013), but they often fall short in  
76 resolving the composition of smaller particles and complex aggregates. On the other hand,  
77 DNA sequencing analysis coupled with accurate reference databases offer a powerful  
78 alternative for extracting detailed taxonomic information from partially degraded material  
79 and complex communities. Amacher *et al.* (2009) pioneered this approach in sediment  
80 traps, using clone libraries to quantify the relative contributions of protist groups to

81 downward particle fluxes in the eastern subtropical North Atlantic. More recently, a  
82 metagenomic approach based on 454 pyrosequencing technology was used to investigate  
83 microbial communities associated with sinking particles in the oligotrophic North Pacific  
84 Subtropical Gyre (NPSG) (Fontanez *et al.*, 2015). However, analogous information for  
85 productive systems with high export potential is lacking.

86 The California Current Ecosystem (CCE) is a coastal upwelling biome characterized  
87 by high production and strong advective fields that transport high-nutrient, high-biomass  
88 coastal waters to oligotrophic offshore areas (Ohman *et al.*, 2013). This gradient is  
89 reflected in phytoplankton composition and productivity (Taylor *et al.*, 2015), food-web  
90 interactions (Landry *et al.*, 2009), and export fluxes (Stukel *et al.*, 2011, 2017). The present  
91 study aims at characterizing the composition of eukaryotic communities involved in the  
92 export and remineralization of sinking particles in the CCE. We capitalize on the  
93 capability of DNA metabarcoding of the 18S rRNA gene to retrieve taxonomic  
94 information from complex environmental samples such as those collected by PITs. Our  
95 specific objectives are to assess 1) the taxonomic compositions of POM sinking out of the  
96 euphotic zone and 2) the compositional changes associated with degradation and  
97 consumption processes transforming this POM below the euphotic zone. Towards these  
98 goals, we deployed formalin-fixed (i.e. fixed) and preservative-free (i.e. live) traps at the  
99 base of the euphotic zone, on the premise that microbial activity and degradation of POM  
100 would be inhibited in formalin-fixed traps (Knauer *et al.*, 1984) but allowed in live traps  
101 (Karl and Knauer, 1984; Lee *et al.*, 1992).

## 103 MATERIALS AND METHODS

### 104 *Study area and sampling strategy*

105 Hydrographic and biological data were collected during the CCE-P1408 Process  
106 Cruise on the R/V Melville (6 Aug - 4 Sept, 2014), as part of the CCE LTER (Long Term  
107 Ecosystem Research) program. A quasi-Lagrangian strategy was adopted to sample  
108 representative water parcels from coastal to offshore conditions over 3-days sampling  
109 periods called 'Cycles'. We used a satellite-tracked drifting array with a drogue at 15 m  
110 to follow the water parcels and sampled the water column daily at the array for a suite of  
111 physical, chemical and biological measurements (Landry *et al.*, 2009). Cycles 1 to 3  
112 were initiated in more productive coastal waters around Point Conception, while Cycles 4  
113 and 5 represented typical oligotrophic offshore conditions (Figure 1). Hydrographic data

114 and water samples were acquired at 6-8 depths from a CTD-rosette system with 10-L  
115 Niskin bottles with Teflon-coated springs.

### 116 *Sediment traps and export measurements*

117 A second drifting array with VERTEX-style drifting sediment traps (Knauer *et al.*, 1979)  
118 was deployed during each cycle to collect sinking particles and assess export fluxes.  
119 The sediment trap consisted of arrays of 8-12 PITs in a cross-like layout attached to the  
120 wire at 2-3 depths below the euphotic zone (100 m, 150 m, and base of the euphotic zone if  
121 shallower than 100 m). The tubes were filled with a brine solution of 0.1  $\mu\text{m}$  filtered  
122 seawater with 50  $\text{g L}^{-1}$  NaCl and 80  $\text{mg L}^{-1}$  of  $\text{SrCl}_2$  added to prevent mixing with *in situ*  
123 water and dissolution of acantharian skeletons (Beers and Stewart, 1970), respectively.  
124 Most tubes were fixed with formaldehyde (0.4% final concentration, fixed traps), to  
125 minimize decomposition and consumption of organic matter (Knauer *et al.*, 1984), while  
126 selected tubes were not fixed (live traps) to allow these biotic processes to continue (Karl  
127 and Knauer, 1984).

128 Upon recovery, the interface of brine and *in situ* water was visually determined in  
129 each tube and the upper layer removed gently by suction. For fixed traps, mesozooplankton  
130 swimmers were removed under a dissecting scope, before the sample was mixed, split  
131 and subsampled for particulate organic carbon (POC) and nitrogen (PON), chlorophyll *a*  
132 (Chl *a*), and phaeopigment analyses as described in Stukel *et al.* (2013). Previous  
133 calibration of our PITs using  $^{238}\text{U}$ : $^{234}\text{Th}$  disequilibrium methods suggest that they are  
134 accurately collecting sinking particles (Stukel *et al.*, 2015). For live traps, swimmers were  
135 not removed, and samples were directly split and filtered to minimize processing time and  
136 potential degradation. For DNA flux and diversity analysis, a known fraction of fixed- and  
137 live- traps was vacuum filtered in parallel through 0.8 and 8  $\mu\text{m}$  Supor membrane filters  
138 (Pall Life Sciences, Port Washington, NY, USA) to obtain replicate independent samples  
139 for  $> 0.8 \mu\text{m}$  and  $> 8 \mu\text{m}$  particles. DNA fluxes from fixed traps were estimated by  
140 dividing the DNA concentration by an extraction efficiency factor of 0.153 (Amacher *et*  
141 *al.*, 2013).

### 142 *Water column chemical and biological analysis*

143 Water-column dissolved inorganic nutrients (DIN) and total and size-fractionated Chl  
144 *a* were obtained from <http://oceaninformatics.ucsd.edu/datazoo/catalogs/ccelter/datasets>,  
145 where sampling and analytical methods are described in detail. For DIN, seawater was  
146 filtered directly from the Niskin bottle using a Suporcap filter capsule (0.1  $\mu\text{m}$  pore size)

147 and major nutrient concentrations ( $\text{NO}_3^- + \text{NO}_2^-$ ,  $\text{NO}_2^-$ ,  $\text{PO}_4^{3-}$ ,  $\text{NH}_4^+$ , and  $\text{SiOH}_3$ ) analyzed by  
148 autoanalyzer using standard methods (Gordon *et al.*, 1992). For total Chl *a* analysis, the  
149 samples were filtered onto 25 mm GF/F filters (Whatman, Maidstone, UK), and pigments  
150 were extracted in 90% acetone at  $-18^\circ\text{C}$  in the dark for 24 h and quantified on a calibrated  
151 10AU fluorometer (Turner Designs, Sunnyvale, CA, USA) using the acidification method.  
152 For size-fractionated Chl *a*, 0.10-0.25 L samples from the surface mixed layer were filtered  
153 through a series of filters with different pore sizes (20  $\mu\text{m}$  Nitex mesh, 8  $\mu\text{m}$ , 3  $\mu\text{m}$  and 1  
154  $\mu\text{m}$  Nucleopore and GF/F filters, Whatman, Maidstone, UK) and analyzed  
155 fluorometrically as above.

156 Water-column samples for DNA analysis were collected in two ways: 1) seawater  
157 was collected from the Niskin spigot into an acid-washed and Milli-Q rinsed  
158 polypropylene bottle, with a known volume (1.5-3 L) then filtered on a 0.8  $\mu\text{m}$  Supor filter  
159 (i.e. called small-fraction). 2) We also emptied entire Niskin bottles into 10-L carboys by  
160 opening the bottom closure and using a large funnel to collect larger organisms  
161 (Michaels *et al.*, 1995). These samples were then concentrated to 250 mL using a 5  $\mu\text{m}$   
162 mesh plankton net and subsequently vacuum-filtered onto 8  $\mu\text{m}$  Supor filters (i.e. called  
163 large size-fraction).

#### 164 *Genomic DNA extraction, PCR amplification and library preparation*

165 DNA was extracted with the Nucleospin Plant kit (Macherey-Nagel, Düren,  
166 Germany), mini version for the water-column 0.8- $\mu\text{m}$  filters and the midi version for 8.0-  
167  $\mu\text{m}$  filters, and PIT samples (Supplementary Table S1). The V4 region of the 18S rRNA  
168 gene was amplified using the eukaryotic primers V4F\_illum (5'  
169 CCAGCASCYGCGGTAATTCC 3') and V4R\_illum (5'  
170 ACTTTCGTTCTTGATYRATGA 3') with Illumina overhang adapters (Forward 5'  
171 TCGTCGGCAGCGTCAGATGTGTATAAGAGACAG 3' and Reverse 5'  
172 GTCTCGTGGGCTCGGAGATGTGTATAAGAGACAG 3') for attaching Nextera  
173 indexes (Piredda *et al.* 2017). PCR amplifications were done in triplicate for each sample.  
174 PCR products were visualized on 1.5% agarose gel and pooled together before purification  
175 with Agencourt AMPure XP purification system (Beckman Coulter, Brea, CA, USA).  
176 Randomly selected 20 purified amplicons were sized and validated using the Agilent High  
177 Sensitivity DNA Assay in the 2100 Bioanalyzer (Agilent Technologies, Palo Alto, CA,  
178 USA). For 6 distinct samples, the purified PCR product was split into three subsamples and  
179 also included in the library to investigate the reproducibility of the sequencing step. A  
180 second PCR step to attach index and Illumina adapters was conducted with the Nextera

181 DNA library Preparation Kit (Illumina), followed by additional AMPure purification and  
182 library validation. The library was then quantified and prepared for 2 x 250 bp sequencing  
183 on a MiSeq platform. For Cycle 1, only fixed-trap samples were available for sequencing,  
184 while the  $> 0.8 \mu\text{m}$  live-trap samples from Cycle 3 were lost during processing. We  
185 sequenced a total of 54 samples, which included water-column samples from mixed-layer,  
186 deep chlorophyll maximum and below (n=22), live-trap samples from 60, 100 and 150 m  
187 (n=21), and fixed-trap samples from 100 m (n=11) (Supplementary Table S1). Raw  
188 sequence have been deposited to GenBank under Bioproject number PRJNA432581.

### 189 *Processing and taxonomic assignation of sequencing reads*

190 Fastq files were checked using FastQC on the Galaxy platform (Goecks *et al.*, 2012)  
191 for sequence length (L) and Quality Score (Q). Forward reads with  $L > 200$  and  $Q > 20$   
192 over 75% of the sequence were retained. For reverse reads, the first 85 bases were of very  
193 bad quality. These 85 bases were removed and sequences with  $L > 150$  and  $Q > 20$  in 50%  
194 of the sequence were retained. Unpaired sequences were removed. All further processing  
195 was performed using mothur v.1.33.0 (Schloss *et al.*, 2009). First, contigs were assembled  
196 from forward and reverse reads, keeping only contigs free of ambiguity. Singletons were  
197 removed and sequences were aligned to 18S rRNA Silva reference. Sequences with  
198 two nucleotide differences were pre-clustered, and chimeras were removed using  
199 UCHIME (Edgar *et al.*, 2011) as implemented in mothur. Only pre-clusters with  $> 10$   
200 sequences were retained. Pre-clustered sequences were taxonomically annotated using  
201 classify.seqs against the PR2 reference database (Guillou *et al.*, 2013) version 4.4  
202 available from [https://github.com/vaulot/pr2\\_database/releases](https://github.com/vaulot/pr2_database/releases). Sequences were clustered  
203 using the average neighbour algorithm to determine OTUs at 97% similarity level. OTUs  
204 were taxonomically assigned using classify.otu and BLASTed against GenBank to confirm  
205 the assignation provided by mothur and to get their percentage of similarity to existing  
206 sequences. Since the community composition for triplicated amplicons was not  
207 substantially different (Supplementary Figure S1), sequences from replicated amplicons  
208 were pooled for subsequent analysis. Similarly, sequences from water-column and live-  
209 trap samples obtained at different depths were combined for downstream analysis. The  
210 OTU abundance table and OTU sequences are available as supplementary material on  
211 Figshare (<https://doi.org/10.6084/m9.figshare.5944291>).

### 212 *Statistics and diversity analysis of sequence data*



213 Data analyses and statistics were done using R version 3.2.4 (R Core Team, 2016)  
214 with the *vegan* 2.4-3 (Oksanen *et al.*, 2017), *ggplot2* (Wickham, 2009), *treemap*  
215 (*Tennekes*, 2017) packages, GraphPad 5.0 (GraphPad Software, La Jolla, CA, USA) and  
216 the Paleontological Statistics software (*Hammer et al.*, 2001). OTU richness was estimated  
217 with *rarefy* function (*vegan* package) on a random subsample of size determined by  
218 the minimum number of sequences found among the samples compared (e.g. fixed vs.  
219 live traps).

## 220 RESULTS

### 221 *Water-column physical, chemical and biological conditions*

222 During the sampling cycles (C1 to C5), we explored 5 water parcels (Figure 1) with  
223 physical and chemical properties that reflected primarily their coastal vs. offshore  
224 characteristics (Table 1, Supplementary Figure S2a). A hierarchical dendrogram based on  
225 both physico-chemical and biological properties clusters coastal C1 and C2 together,  
226 separated from offshore C4 and C5, with intermediate conditions for C3 (Supplementary  
227 Figure S2b). C1 and C2 exhibited colder and saltier surface waters and shallower  
228 nitracline depth indicative of coastal upwelling (Table 1). Surface Chl *a* concentration  
229 was also higher with large phytoplankton cells (> 20  $\mu\text{m}$ ) accounting for ~40% of Chl  
230 *a* for C1 and C2, while the contribution of large cells was minor (~2%) for C4 and C5  
231 (Table 1, Supplementary Figure S2c).

232 POC and PON fluxes were higher at coastal cycles and decreased offshore (Figure  
233 2). DNA fluxes measured in fixed traps at 100 m followed the same trend (Figure 2a) and  
234 were significantly correlated with POC fluxes ( $r = 0.88$ ,  $p < 0.05$ ,  $n = 5$ ). The Chl *a* to  
235 phaeopigment ratio of sediment trap material indicated that pigment flux during all cycles  
236 was dominated by fecal-derived material (Figure 2b). The higher values observed at the  
237 coastal C1 and C2 cycles, however, indicated a higher contribution of ‘fresh’ algae  
238 compared to offshore C4 and C5 (Figure 2b).

### 239 *18S V4 OTU distribution and diversity patterns*

240 Over 18 million paired reads were obtained from the sequencing run, half of which  
241 were removed after filtering based on quality and length (Supplementary Table S2). After  
242 processing, we obtained 2,662 OTUs (97% similarity) corresponding to 2,802,466  
243 sequences (Supplementary Table S2).

244 For protists, rarefaction curves for different stations were not saturated for all  
245 samples, although the analysis of pooled sequences from different sample types suggest  
246 an adequate recovery of epipelagic protist diversity for cycles C2-C5 (Supplementary  
247 Figure S3a). Cycle C1 was excluded from this analysis because only fixed-trap samples  
248 were available. Protistan OTU richness increased from coastal to oceanic locations,  
249 being higher in water-column compared to fixed- and live-trap samples for all cycles  
250 (Figure 3a). More protistan OTUs were recovered in live compared to fixed traps for C2 and  
251 C3, while fixed traps had higher or similar OTU numbers compared to live traps for C4  
252 and C5 (Figure 3a). OTU richness for water-column protists was higher in the smaller  
253 compared to the larger size-fraction samples (Figure 3b). This difference decreased in  
254 sediment traps, particularly in fixed traps, where protistan richness was similar or higher  
255 in the larger fraction (Figure 3b). Similar diversity patterns were observed for all eukaryotic  
256 OTUs, including metazoans (Supplementary Figure S3b).

257 Non-metric multidimensional scaling (NMDS) analysis, based on protistan OTUs  
258 abundance and Bray-Curtis distances, ordinated samples into three main clusters  
259 corresponding to the different sampling methods (Figure 4). Compositional changes in  
260 water-column samples were significantly correlated with surface temperature, salinity,  
261 nitracline depth, and Chl *a*, with size fraction (NMDS<sub>1</sub>,  $r^2 = 0.61$ ,  $P = 0.001$ ) and cycle  
262 (NMDS<sub>2</sub>,  $r^2 = 0.30$ ,  $P = 0.02$ ) being the primary ordination factors (Supplementary Figure  
263 S4). Sampling cycle emerged as primary ordination factor for the correlations of trap  
264 samples with physico-chemical and biogeochemical variables (Supplementary Figure S4).

#### 265 *Taxonomic composition of the water-column eukaryotic community*

266 OTU taxonomy was assigned based on the curated 18S rRNA database PR<sup>2</sup> (Guillou  
267 *et al.*, 2013) which uses 8 different taxonomical levels from Kingdom to Species.  
268 Protistan sequences dominated the community across all cycles and size fractions ( $58 \pm$   
269  $27\%$  of eukaryotic sequences), with a substantial contribution of metazoan sequences  
270 ( $39 \pm 26\%$ , Supplementary Figure S5).

271 Among protists, Dinophyta was the most abundant group ( $72 \pm 13\%$  of protistan  
272 sequences) followed by Radiolaria ( $12 \pm 13\%$ ) and Chlorophyta ( $4.7 \pm 7.1\%$ ) (Figure 5).  
273 Ochrophyta ( $2.1 \pm 2.0\%$ ) contributed less on average but represented a substantial  
274 percentage of the protistan community at times (Figures 5 and 6). Dinophyta included  
275 sequences affiliated with Syndiniales and Dinophyceae in similar relative abundances  
276 ( $\sim 20\%$ , Figure 6), although their distributional pattern and partitioning between size

277 fractions differed markedly (Figure 7). In addition to *Ptychodiscus noctiluca*, uncultured  
278 Dinophyceae and Syndiniales OTUs were among the most abundant in water-column  
279 samples (Supplementary Table S3 and Figure S6).

280 Most radiolarian sequences were affiliated with Collodaria and Acantharia,  
281 although groups such as Spumellaria and RAD-A were occasionally important (Figure 8).  
282 Chlorophyta (green algae) were mostly picoplanktonic belonging to Chloropicophyceae  
283 (*Chloroparvula*, previously prasinophytes clade VII-B1, Lopes dos Santos *et al.*, 2017a)  
284 and Mamiellophyceae (*Bathycoccus*) (Figure 6). Both picoplanktonic groups peaked in  
285 coastal waters (e.g. Chloropicophyceae represented ca. 20% of protistan sequences in C2),  
286 and decreased offshore (Figures 6 and 7). Ochrophyta (photosynthetic stramenopiles)  
287 sequences were also more abundant in the smaller size fraction (Figure 5) and  
288 comprised mainly of sequences assigned to pelagophytes and diatoms (Figure 6), with the  
289 latter reaching ~5-fold higher relative abundance toward the coast (Figure 7). The  
290 contribution of Phaeodaria, a rhizarian group related to Radiolaria, also peaked in coastal  
291 waters, but in the larger size fraction (Figure 7).

#### 292 *Taxonomic composition of eukaryotic community in fixed traps*

293 Fixed-trap samples were overwhelmingly dominated by protistan sequences, mainly  
294 Radiolaria ( $88 \pm 8.6\%$  of eukaryotic sequences) and Dinophyta ( $8.6 \pm 6.3\%$ ) (Figures 5  
295 and Supplementary Figure S5). Metazoan sequences accounted for  $6.1 \pm 5.0\%$  and were  
296 dominated by Crustacea, with Mollusca and Cnidaria contributing less on average  
297 (Supplementary Figure S5).

298 The dominance of Radiolaria was a consistent feature across cycles (Figure 6),  
299 with contributions from a diverse suite of OTUs from the four major radiolarian groups  
300 (Acantharia, Collodaria, Nassellaria and Spumellaria) (Figure 8, Supplementary Table S3).  
301 Overall, Spumellaria was the most abundant group ( $60 \pm 22\%$  of radiolarian sequences)  
302 followed by Acantharia ( $22 \pm 28\%$ ) and Collodaria ( $4.8 \pm 11\%$ ) (Figure 8). Among  
303 Dinophyta, Dinophyceae (true dinoflagellates) were relatively more abundant than  
304 Syndiniales, and their contribution increased in offshore cycles (Figure 7) with  
305 uncultured Dinophyceae and *P. noctiluca* being the most abundant OTUs (Supplementary  
306 Table S3 and Figure S6). Diatoms comprised most Ochrophyta sequences in fixed traps  
307 from cycles C1 and C2 (Figure 6), belonging mainly to the same genera (*Pseudo-*  
308 *nitzschia* and *Chaetoceros*) as those in the water column (Supplementary Figure S6). They  
309 virtually disappeared offshore, where sequences affiliated with the heterotrophic

310 nanoflagellate (HNF) *Paraphysomonas imperforata* (Chrysophyceae) became the  
311 dominant Ochrophyta group (Figure 6).

312 Taxonomic composition of eukaryotic community in live traps

313 Metazoan sequences dominated live-trap samples ( $78 \pm 18\%$  of eukaryotic  
314 sequences, Supplementary Figure S5). Among them, Crustaceans belonging to different  
315 copepod genera, notably *Metridia* spp., contributed most to the eukaryotic community  
316 followed by Cnidaria and Mollusca (Supplementary Table S3).

317 Among protists, Dinophyta ( $51 \pm 5.7\%$  of protistan sequences) and Stramenopiles\_X  
318 (heterotrophic stramenopiles,  $24 \pm 7.2\%$ ) were the dominant groups, with Radiolaria  
319 ( $9.6 \pm 3.7\%$ ) and Phaeodaria ( $5.4 \pm 5.5\%$ ) contributing less (Figures 5 and 6). In addition  
320 to the dinoflagellate *P. noctiluca*, the HNFs *Caecitellus parvus* and *C. paraparvulus* (order  
321 Anoeales) were the most abundant protistan species in live traps (Supplementary Table S3  
322 and Figure S6), accounting for the high proportion of Stramenopiles\_X. The abundance of  
323 these HNFs decreased significantly in the  $> 8 \mu\text{m}$  fraction, suggesting that they are free  
324 living or feeding on particles and aggregates to which they are loosely attached  
325 (Figures 6 and 7). Diatoms and chrysophytes recovered from live traps showed the same  
326 spatial and size-fraction distribution as in fixed traps, although the contribution of  
327 chrysophytes in offshore cycles was higher in live traps (Figure 7). Most radiolarian  
328 sequences recovered from live traps belonged to Acantharia and Spumellaria (Figure 8).  
329 RAD-A and Collodaria were detected in all cycles but showed higher contribution in  
330 offshore cycles C4 and C5 (Figure 8). Phaeodaria, mostly affiliated with *Aulacantha* spp.  
331 (Supplementary Table S3), were more abundant in coastal cycles C1 and C2, where they  
332 represented up to 20% of protistan sequences (Figure 7).

## 334 DISCUSSION

335 Overall, the protistan communities in water-column and trap samples were markedly  
336 dissimilar (Figures 5 and 6), showing higher similarity in sinking material across cycles  
337 than between traps and the overlying euphotic zone (Figure 4). Such pattern has been  
338 observed in previous molecular-based studies (Amacher *et al.*, 2013; Fontanez *et al.*, 2015)  
339 and suggests either that a common suite of organisms is responsible for export despite  
340 dramatic differences in ambient microbial communities or that these organisms are  
341 resistant to degradation of the genetic material. Radiolaria were identified as the most  
342 abundant group in POM sinking out of the euphotic zone, in addition to phytoplankton

343 and metazoan zooplankton, usually considered as having a major role in vertical export  
344 in the CCE (Figures 5 and 6). The observed dominance of radiolarian sequences in fixed  
345 traps (Figures 5 and 6) is also consistent with clone library analysis of trap samples  
346 from the eastern subtropical North Atlantic, where radiolarian clones were most abundant  
347 (Amacher *et al.*, 2009), and water column samples from the Sargasso sea, where high  
348 proportion of radiolarian clones were obtained below the euphotic zone (Not *et al.*, 2007).  
349 In contrast to these observations, the relative abundance of radiolarians determined by 18S  
350 rRNA metabarcoding in the western Antarctic Peninsula was negatively linked to  
351 community export potential (Lin *et al.*, 2017), suggesting that the role of this group in  
352 vertical export may differ across systems.

353 Metabarcoding of 18S rRNA genes is largely used for assessing the composition of  
354 aquatic microbial communities (De Vargas *et al.*, 2015; Pearman *et al.*, 2016; Piredda *et al.*,  
355 2017; Lopes dos Santos *et al.*, 2017b) and the spatio-temporal patterns of specific  
356 taxonomic groups (Egge *et al.*, 2015; Ichinomiya *et al.*, 2016; Zouari *et al.*, 2018). Yet, a  
357 number of well identified limitations and potential biases are acknowledged with respect to  
358 absolute quantification of plankton groups. Along with PCR biases, the presence of  
359 multiple copies of 18S rRNA genes and its variation across taxa (Zhu *et al.*, 2005) affects  
360 the quantitative interpretation of community compositional changes from read abundance  
361 data. In this regard, the prevalence and high relative contributions of Dinophyceae and  
362 Syndiniales in this study (Figures. 6 and 7) and other meta-genetic surveys (Guillou *et al.*,  
363 2008; Amacher *et al.*, 2009; Lie *et al.*, 2013; De Vargas *et al.*, 2015; Piredda *et al.*, 2017)  
364 would be partially due to the high DNA content and number of gene copies in these  
365 groups. Similarly, the multi-nuclear nature of Radiolaria (Anderson, 1983; Suzuki *et al.*,  
366 2009; Suzuki and Aita, 2011) and the high gene copy number observed in Collodaria  
367 (Biard *et al.*, 2017) could explain their high relative abundances in meta-genetic surveys  
368 (Decelle *et al.*, 2014; De Vargas *et al.*, 2015, Biard *et al.*, 2017).

369 Nonetheless, a positive relationship between 18S rRNA copy number and cell length  
370 has been reported across protists spanning orders of magnitude in cell size (Zhu *et al.*,  
371 2005; Godhe *et al.*, 2008; De Vargas *et al.*, 2015; Biard *et al.*, 2017), which encourages the  
372 cautious use of read abundances to infer community composition general patterns and  
373 dynamics on a quasi-biomass basis (i.e., the larger organisms have proportionately more  
374 DNA reads). Further support for an ecological, rather than bias, interpretation of  
375 compositional changes between water-column and trap samples (Figures. 4 and 5) and  
376 group-specific relative-abundance spatial patterns (Figures 6 and 7) comes from a recent

377 study that reported good agreement between relative picoeukaryotic cell and environmental  
378 sequence abundances (Giner *et al.*, 2016). The community composition obtained by  
379 parallel sequencing of replicated PCR products was virtually identical (Figure S1),  
380 suggesting that potential errors linked to sequencing were not responsible for the  
381 differences observed among samples. Moreover, the comparative approach adopted here  
382 provides a robust framework for interpreting diversity and compositional changes between  
383 fixed and live traps from an ecological and biogeochemical perspective.

384 The prevalence of radiolarian sequences in sinking POM observed in this study  
385 (Figure 6) is remarkable, considering the major differences in productivity and trophic  
386 structure of the epipelagic communities along the CCE environmental gradient (Table 1  
387 and Supplementary Figure S2). Analysis of our preserved samples under epifluorescence  
388 microscopy (data not shown) confirmed the presence of Radiolaria, although the  
389 dissolution of hard structures observed for this, and other taxonomic groups (e.g. diatoms),  
390 precluded reliable estimates of biomass and finer taxonomic assignment. Analysis of  
391 DNA sequences, however, shows that different radiolarian groups (Figure 8) and OTUs  
392 (Supplementary Table S3 and Figure S6) contribute broadly to this result, highlighting the  
393 functional diversity of the group. Previous microscopical and geochemical analyses of  
394 sediment trap material have stressed the importance of major radiolarian groups to export  
395 (Takahashi, 1983; Gowing and Coale, 1989; Michaels *et al.*, 1995; Decelle *et al.*,  
396 2013; Boltovskoy, 2017). The enrichment of Radiolaria in sediment traps may reflect both  
397 the high contribution of this group to particle export and the inadequacy of CTD and net  
398 tow sampling to capture these fragile and patchy amoeboid organisms (Michaels *et al.*,  
399 1995; Dennett *et al.*, 2002; Suzuki and Not, 2015).

400 Several characteristics of radiolarian cell structure and ecology are consistent with  
401 their high export potential. The silica or strontium sulfate skeletons of most Polycystines  
402 (Collodaria, Nasellaria, and Spumellaria) and Acantharia, respectively, provide substantial  
403 mineral ballast (Takahashi, 1983; Suzuki and Not, 2015). In addition, their amoeboid  
404 nature and sticky pseudopodia can catalyze the formation of aggregates with high sinking  
405 velocities (Takahashi, 1987). In contrast to recent genomic studies pointing to Acantharia  
406 and Collodaria as the key 'export' taxa in tropical and subtropical oceans (Fontanez *et al.*,  
407 2015; Guidi *et al.*, 2016), we found that Spumellaria are the most important export  
408 contributors in the CCE (Figure 8). Gowing (1986) also showed Spumellaria to be the  
409 dominant radiolarian group in microscopical analyses of sediment trap material from the  
410 oligotrophic NPSG VERTEX station. In addition to Acantharia, which represented > 50%

411 of radiolarian sequences in C1 and C4, we also detected a substantial contribution by  
412 Collodaria in C2 (Figures 6 and 8). Considering the patchy distributions and large size  
413 ( $\mu\text{m}$ -to- $\text{cm}$ ) of single-celled and colonial Collodaria, it is difficult to assess their  
414 contributions to export accurately, based only on discrete sediment trap analysis (Michaels  
415 *et al.*, 1995). However, the high relative abundance of Collodaria in the CCE region  
416 inferred from molecular (Figures 6 and 8) and *in situ* image analysis (Underwater  
417 Vision Profiler 5, UVP5) during this and previous cruises (Ohman *et al.*, 2012; Biard *et al.*,  
418 2016) supports their important role in export flux.

419 Dinoflagellates were the most abundant non-radiolarian protists in both water-  
420 column and fixed-trap samples from cycles C1 and C2 in this region (Figures 6 and 7).  
421 The warm anomaly that developed in the NE Pacific during the 2013-2014 winter (Bond  
422 *et al.*, 2015), colloquially referred to as ‘the blob’, had already hit the California coast at  
423 the time of our cruise (Gentemann *et al.*, 2017) and was responsible for the weak summer  
424 upwelling that we encountered. Off the Oregon coast, changes in plankton community  
425 composition associated with ‘the blob’ included higher dinoflagellates abundances and  
426 penetration of open-ocean copepod species onto the continental shelf region (Peterson *et al.*,  
427 2017). Increased abundance of autotrophic dinoflagellates has also been observed in  
428 the Point Conception region during years of delayed upwelling (Taylor *et al.*, 2015).  
429 Consistent with this, our molecular survey revealed high sequence abundance of *P.*  
430 *noctiluca* (Supplementary Table S3 and Figure S6), a dinoflagellate species with distinctive  
431 cell covering characteristics and widespread distribution (Gómez *et al.*, 2016) that is  
432 generally less common in the CCE. Among green algae, the dominance of prasinophytes  
433 clade VII-B1 (Chloropicophyceae), a group typical of offshore waters (Lopes dos Santos  
434 *et al.*, 2017a), over Mamiellophyceae, characteristics of coastal waters, is another  
435 indication of the oceanic characteristics of the epipelagic zone. Collodaria is typically  
436 associated to warm oligotrophic waters (Dennett *et al.*, 2002; Biard *et al.*, 2016), and the  
437 observed abundance in mesotrophic coastal waters of C2 (Figures 6 and 8) could have been  
438 favored by the warm anomaly. It therefore seems likely that the relatively low contribution  
439 of diatoms to the water-column assemblage and export fluxes was a consequence of  
440 anomalous conditions. Whether the high contribution of Radiolaria indicated by our  
441 molecular analyses was enhanced by these 2014 conditions cannot be directly addressed  
442 without comparable data from years with ‘normal’ conditions, although several lines of  
443 evidence argue against this idea.

444 On one hand, Radiolaria have been shown to dominate clone libraries recovered  
445 from sediment traps despite the dominance of diatoms in the upper water column in the  
446 eastern subtropical North Atlantic (Amacher *et al.*, 2009). On the other hand, although  
447 higher fluxes of polycystine radiolarians (living + empty skeletons) have been found in  
448 coastal upwelling compared to oligotrophic offshore waters of the CCE, a clear  
449 relationship between primary production and these fluxes could not be established  
450 (Gowing and Coale, 1989). In the subarctic Pacific, however, a positive relationship  
451 between polycystine radiolarian fluxes and primary production was reported at station  
452 PAPA (Takahashi, 1987), suggesting differences in the export role of these group among  
453 systems. In our 2014 CCE study, microscopical counts of large phaeodarians from fixed  
454 traps and in UVP5 profiles, yielded lower abundances compared to previous and  
455 following year estimates for this group of rhizarians (Stukel *et al.*, 2018; Biard *et al.*, 2018).  
456 Nonetheless, our study showed that phaeodarians represented a significant fraction of the  
457 protistan sequences in live traps (Figures 5 and 7). Altogether, these findings argue against  
458 the idea that Radiolaria dominance was due to anomalous conditions and support the key  
459 role of rhizarians in export fluxes as a general feature of the CCE, and potentially other  
460 eastern boundary upwelling systems.

461 One striking result was the markedly lower contribution of Radiolaria taxa in live  
462 traps (Figures 5 and 6), indicating rapid remineralization of organic matter associated  
463 with this group. Sinking particles and aggregates serve as natural ‘hot spots’ for both  
464 microbial and metazoan activity (Karl and Knauer, 1984; Taylor *et al.*, 1986). Live traps  
465 were highly enriched with sequences from larger copepods like *Eucalanus* and *Metridia*  
466 spp. (Supplementary Table S3 and Figures S6), consistent with the potential of copepods to  
467 consume and transform sinking organic particles (Lampitt *et al.*, 1990; Noji, 1991; Iversen  
468 and Poulsen, 2007).

469 Rather than ingesting sinking particles directly, detritivorous zooplankton may  
470 benefit from the microbial growth enhanced by particle fragmentation (Mayor *et al.*,  
471 2014). Such “microbial gardening” would be consistent with the dramatic increase of  
472 opportunistic HNFs such as *Caecitellus* spp. and *P. imperforata* (Supplementary Table  
473 S3 and Supplementary Figure S6), which dramatically increase in enrichment cultures  
474 (Lim *et al.*, 1999). Heterotrophic protists are known to thrive around marine snow particles  
475 (Caron *et al.*, 1982, 1986; Silver *et al.*, 1984), but whether this microbial boost is directly  
476 exploited by larger metazooplankton or consumed first by small microzooplankton is  
477 unclear. The increase of ciliates in both fixed (1.8 % of protistan sequences) and live traps



478 (2.2 %) relative to the water column (0.66 %) in this and previous studies (Amacher *et al.*,  
479 2009), indicates that they likely prey on HNF, which function as a trophic link in these  
480 rich microenvironments. Further evidence of heterotrophic protistan activity associated  
481 with the degradation of sinking organic material is indicated by the enrichment of  
482 Phaeodaria in live compared to fixed traps (Figures 5 and 7). This amoeboid group has  
483 been shown to feed on HNF and eukaryotic algae and proposed to consume marine snow  
484 particles (Gowing, 1986; Gowing and Bentham, 1994). Data from the UVP5 have shown  
485 large (> 600 µm) species of phaeodarian forming a high-density layer below the euphotic  
486 zone (Ohman *et al.*, 2012), although their abundances were notably lower during this cruise  
487 compared to previous ones (Stukel *et al.* 2018; Biard *et al.* 2018). The high relative  
488 abundance of Phaeodaria in live traps (Figures 5), together with their trophic biology and  
489 vertical distribution in the CCE, further supports the important role that this group may have  
490 in the cycling of particulate organic matter sinking below the euphotic zone.

## 491 CONCLUSION

492 While phytoplankton and crustacean zooplankton are generally viewed as the  
493 major biological sources of vertical export in productive systems, our results, showing  
494 the abundance and diversity of rhizarian sequences in particulate material sinking below  
495 the euphotic zone, indicate an important role for this group in the export and cycling of  
496 particulate organic matter in the CCE. Recent studies using advanced molecular and  
497 imaging technologies have revealed unprecedented global abundances of Rhizaria  
498 (Stemmann *et al.*, 2008; Not *et al.*, 2009; Biard *et al.*, 2016), but the functions and impacts  
499 of this group in pelagic ecosystems have yet to be accurately quantified. The high  
500 abundance and diversity of Radiolaria reported here, and their prevalence from coastal  
501 upwelling to oceanic oligotrophic conditions, stress the need to better characterize the  
502 group's functional diversity to improve understanding of biological controls on vertical  
503 fluxes.

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## CONFLICT OF INTEREST

The authors declare no conflict of interests.

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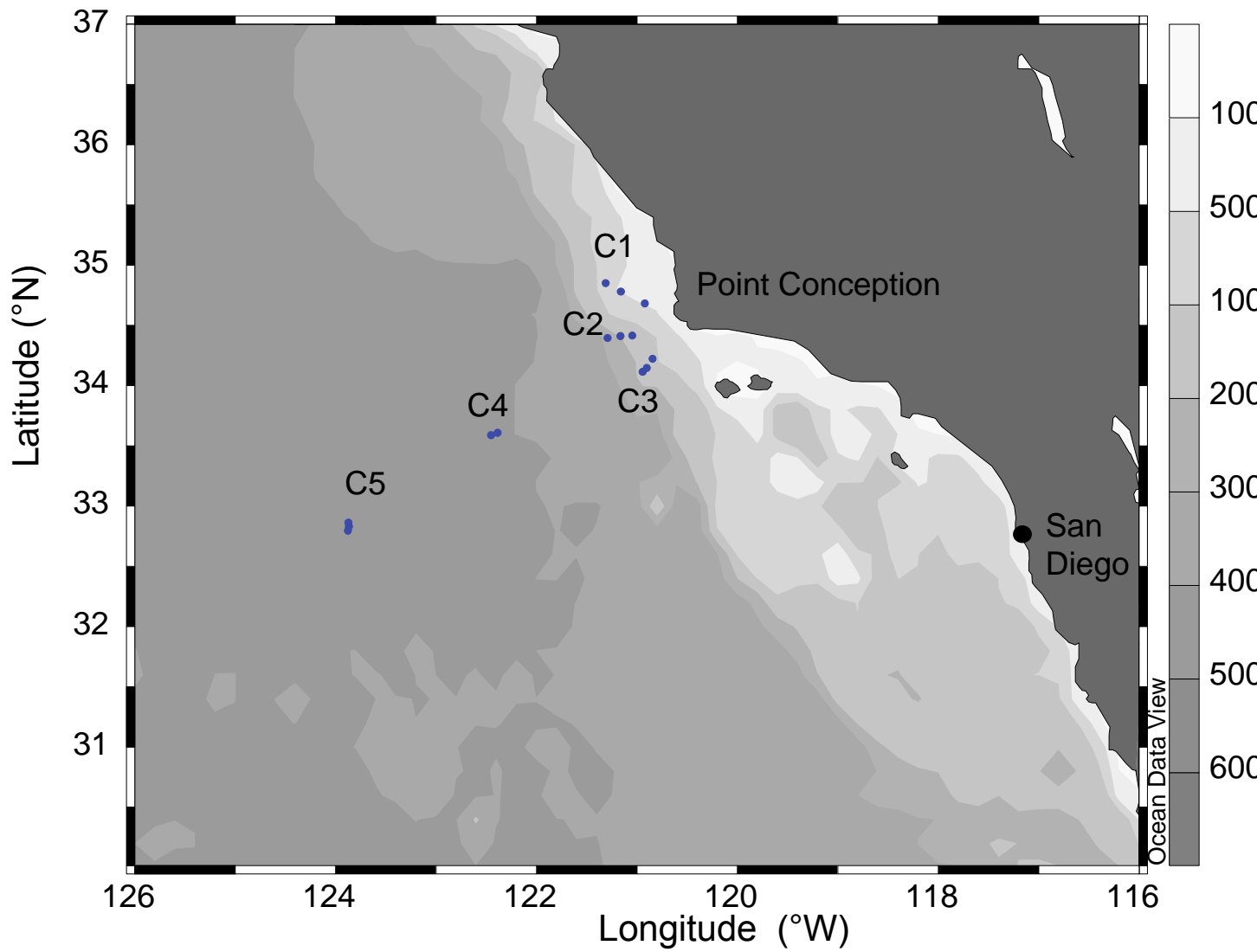
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## 747 TABLES

748 **Table 1.** Water column physical, chemical and biological properties averaged from 7 CTD casts  
749 conducted throughout the 3-days duration of each Cycle.

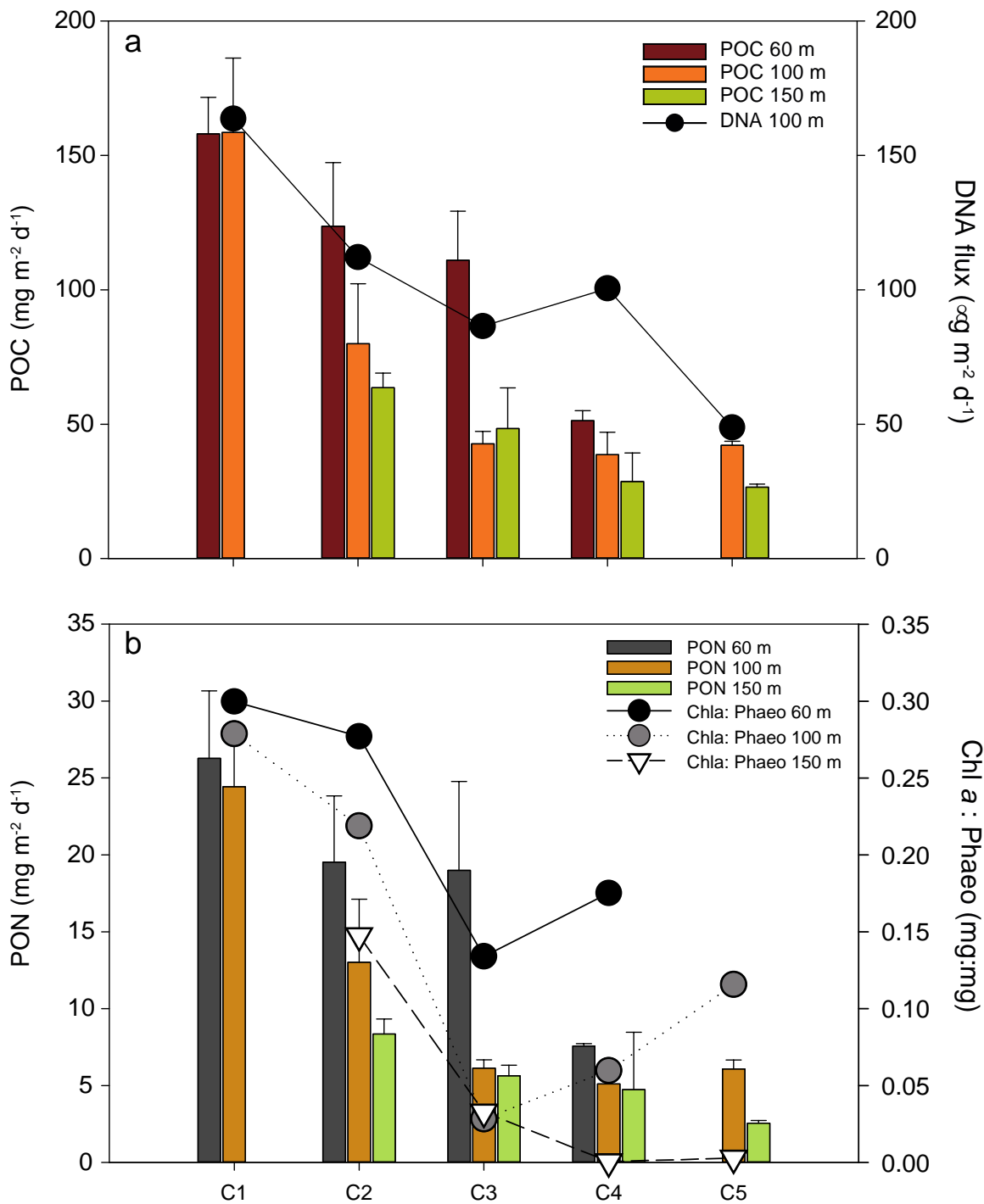
|                                  | <b>C1</b>   | <b>C2</b>   | <b>C3</b>   | <b>C4</b>   | <b>C5</b>   |
|----------------------------------|-------------|-------------|-------------|-------------|-------------|
| Salinity                         | 33.47 ± 0.0 | 33.42 ± 0.0 | 33.33 ± 0.0 | 33.02 ± 0.0 | 33.09 ± 0.0 |
| Temperature (°C)                 | 16.5 ± 0.0  | 16.9 ± 0.0  | 18.6 ± 0.1  | 19.2 ± 0.0  | 19.8 ± 0.0  |
| Mixed layer depth (m)            | 17.6 ± 4.3  | 29.3 ± 5.7  | 17.3 ± 5.9  | 24.3 ± 4.8  | 28.6 ± 4.4  |
| Nitracline depth (m)             | 30.9 ± 5.8  | 40.4 ± 14.0 | 33.1 ± 6.0  | 70.6 ± 2.1  | 91.6 ± 5.7  |
| Surface Chl <i>a</i> (µg/L)      | 0.59 ± 0.15 | 0.69 ± 0.08 | 0.21 ± 0.03 | 0.10 ± 0.02 | 0.08 ± 0.01 |
| Surface Chl <i>a</i> > 20 µm (%) | 42.6 ± 5.0  | 39.0 ± 5.7  | 3.4 ± 1.2   | 1.9 ± 0.3   | 1.8 ± 0.1   |

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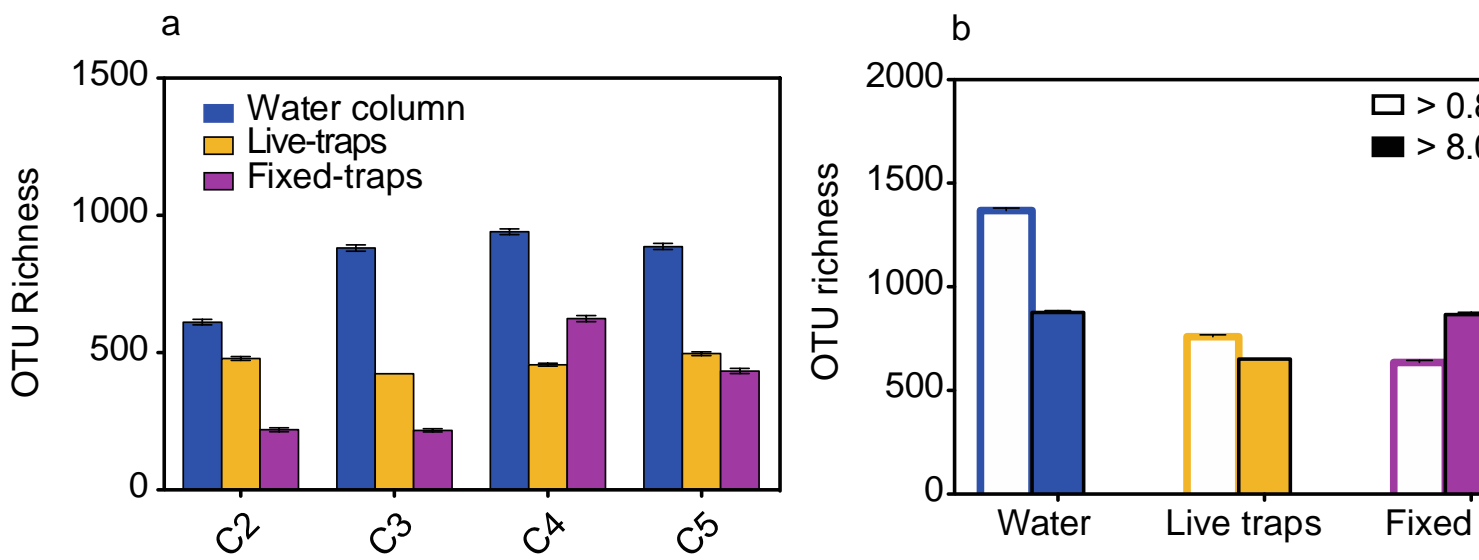


**Figure 1.** Map showing the location of the different water masses surveyed. Each waterparcel was tracked for three consecutive days called Cycles (C1-C5). Blue dots represent the position of the water parcel during the predawn CTD carried out daily during the cycles.

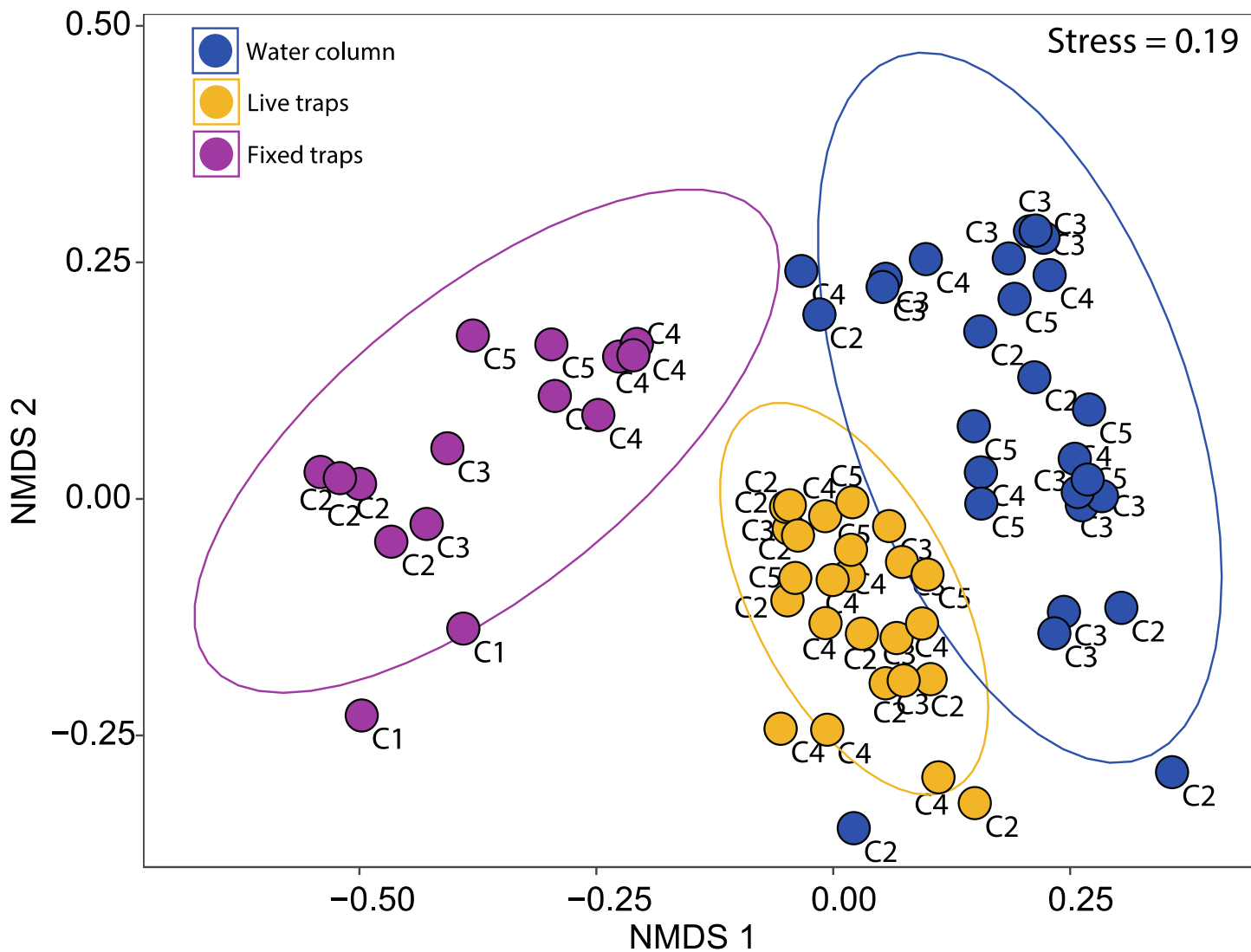




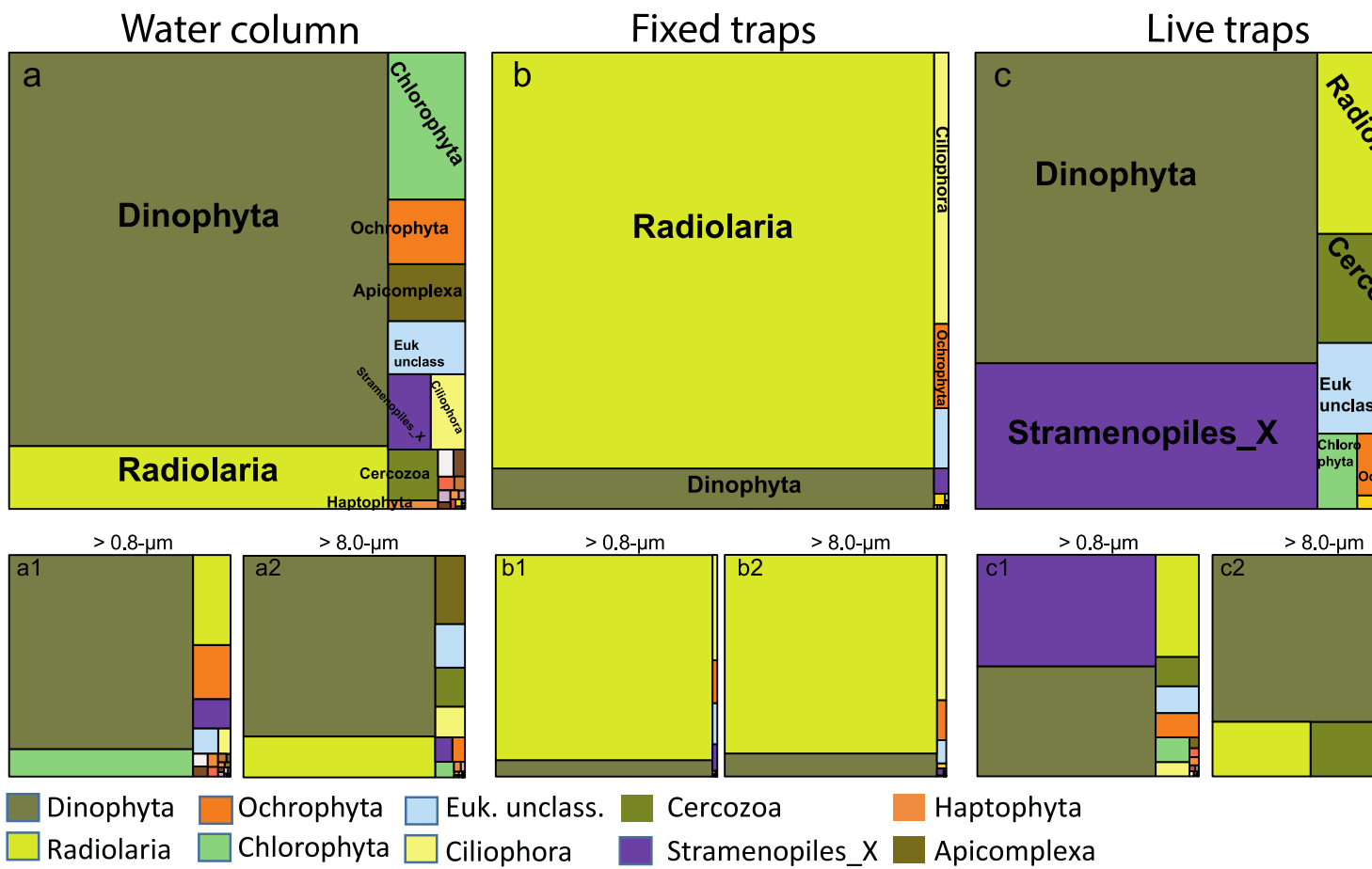
**Figure 2.** (a) POC and DNA fluxes in standard fixed sediment traps. (b) PON fluxes and Chlorophyll to Phaeopigments ratio in the traps. Error bars refer represent the standard deviation of POC and PON fluxes ( $n = 3$ ) in traps deployed at 60, 100, and 150 m.



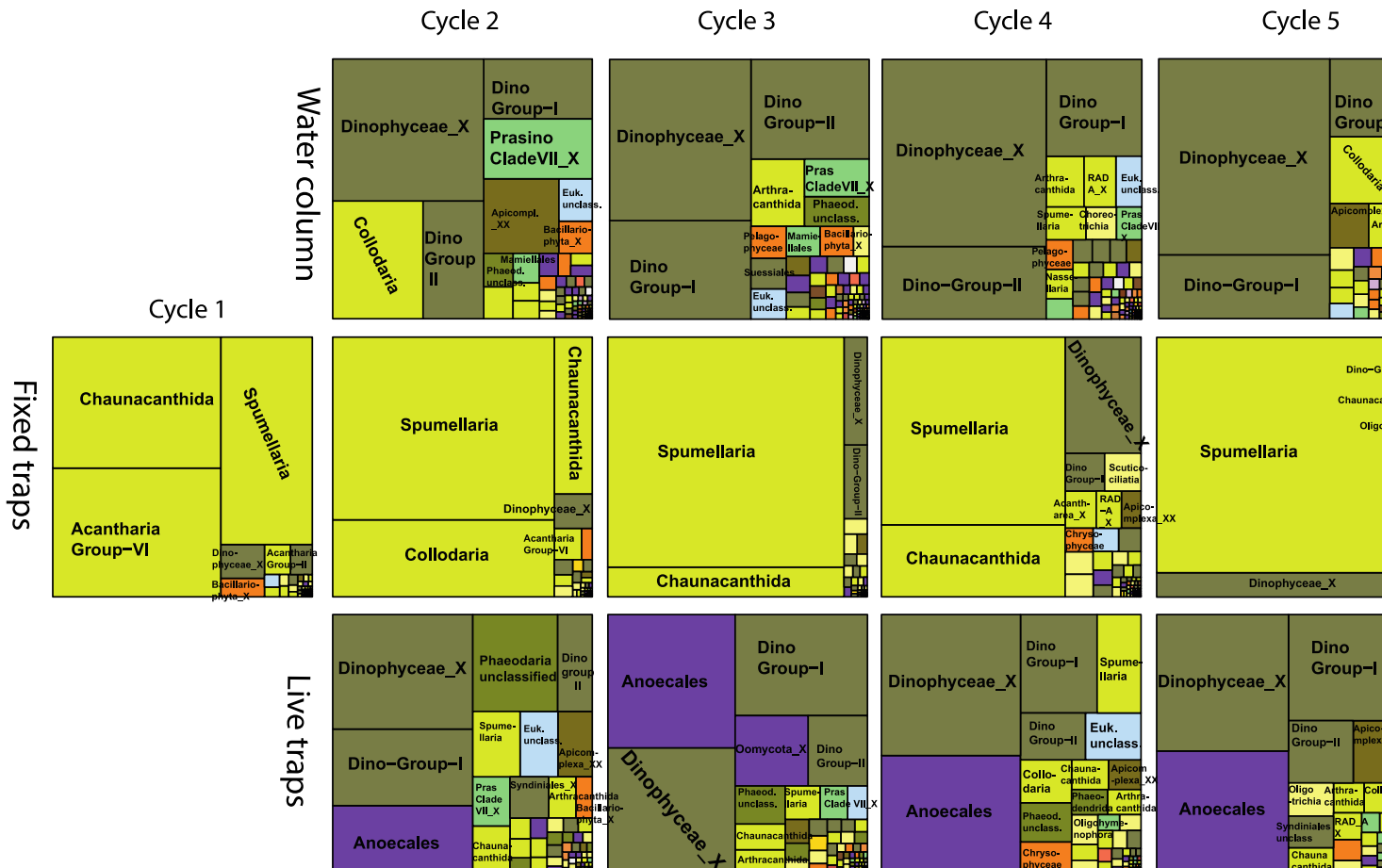
**Figure 3.** Protistan OTU richness calculated for the minimum sampling effort common to all samples compared. (a) Water column, live- and fixed-sediment trap samples comparison across different cycles. (b) Small and large size fraction comparison in different sample types. Colours refer to samples collected from the water column (blue), live (yellow), and fixed traps (purple).



**Figure 4.** Non-parametric multidimensional scaling ordination (nMDS) plot in 2-dimension configuration (K=2) based on Bray-curtis dissimilarity between protist OTU community composition of all samples. 95% confidence ellipses for each method type are represented.



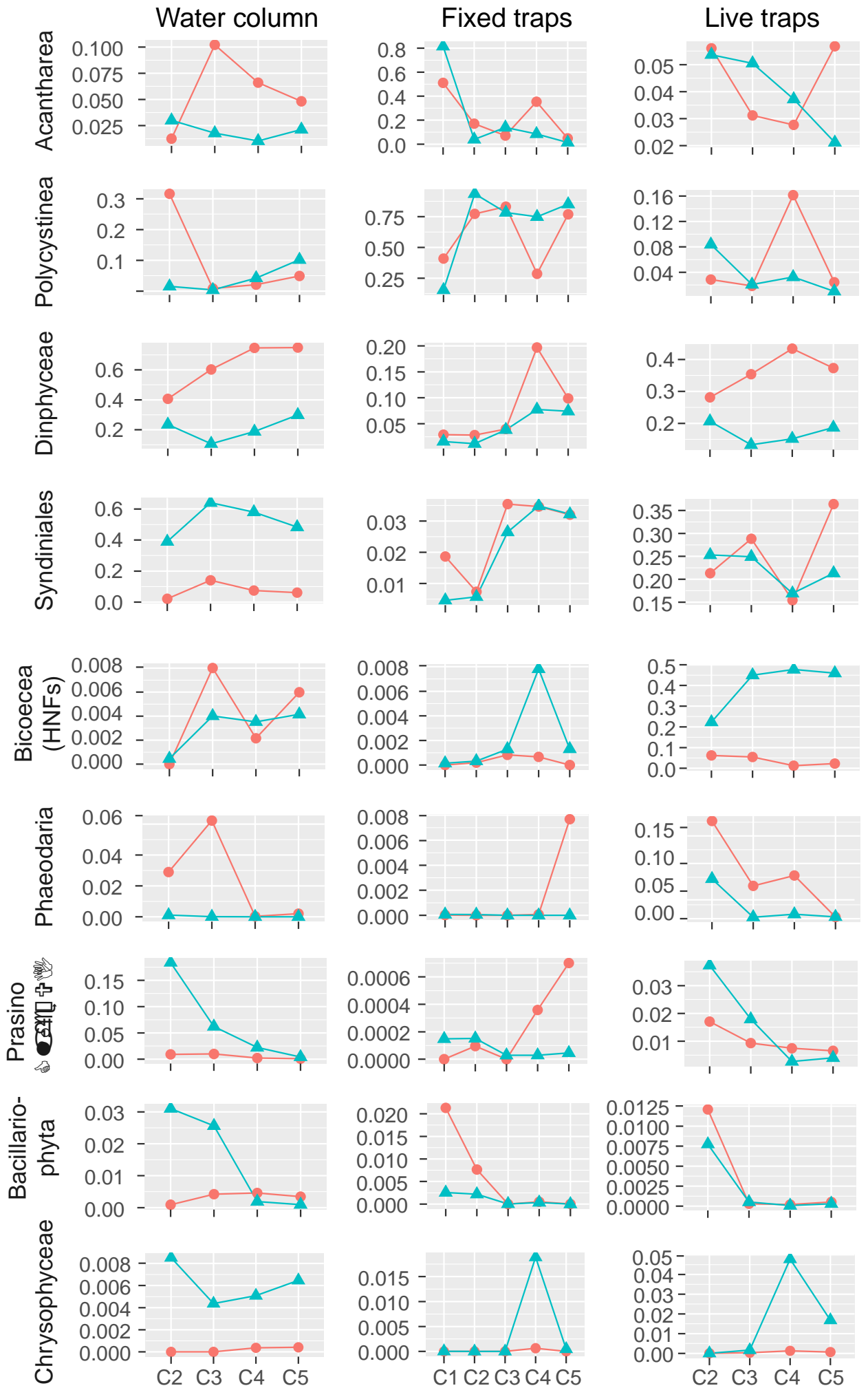
**Figure 5.** Mean percentage of 18S rDNA reads affiliated to protistan divisions and major taxonomic groups in samples from the water column (a), fixed (b) and live traps (c). Total (> 0.8  $\mu\text{m}$ ) and large (> 8  $\mu\text{m}$ ) size-fractions are presented at the bottom left (1) and right (2) of each sample type, respectively.



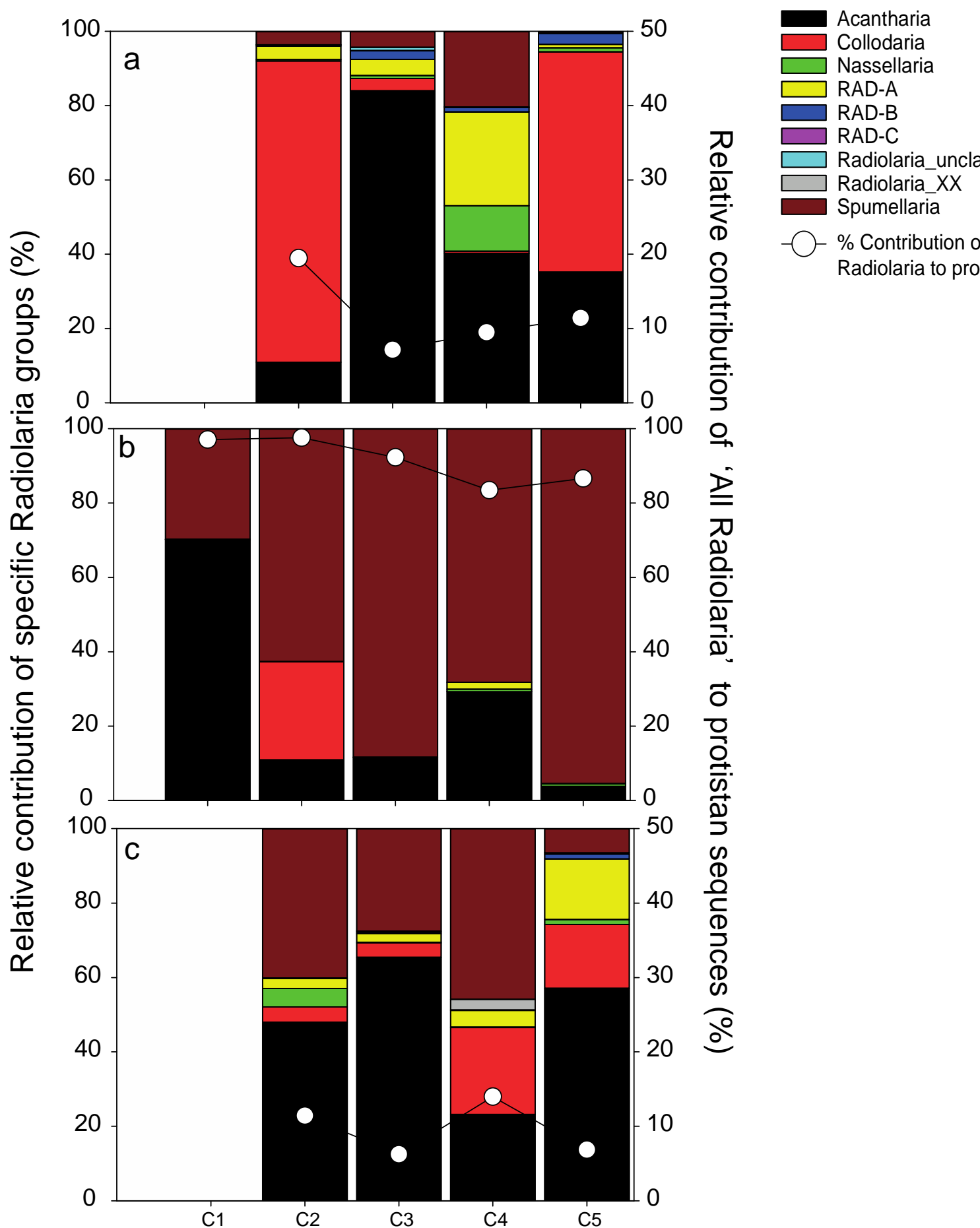
**Figure 6.** Mean percentage of 18S rDNA reads affiliated to protistan taxonomic groups (classes and orders mainly) in water-column, fixed- and live-trap samples across different sampling cycles. The area represents the mean percentage of reads affiliated to each protistan group (only groups with > 1% mean contribution are labelled). Color codes, as in Figure 5, represent the taxonomic affiliation, mainly at the division level.

● > 8.0- $\mu$ m      ▲ > 0.8- $\mu$ m

Fraction of protistan sequences



**Figure 7.** Mean relative abundance of 18S rRNA sequences of most abundant protistan groups in the total (> 0.8  $\mu\text{m}$ ) and large (> 8  $\mu\text{m}$ ) size-fraction samples from the upper water column (0-100 m), fixed- and live-traps, across different cycles.



**Figure 8.** Relative contribution of Radiolaria to total protistan sequences (solid points) and percentage contribution of specific radiolarian groups in the water column (a), fixed (b) and live (c) trap samples across different cycles.