



**HAL**  
open science

## **popharvest: An R package to assess the sustainability of harvesting regimes of bird populations**

Cyril Eraud, Tiphaine Devaux, Alexandre Villers, Fred A Johnson, Charlotte Francesiaz

### ► To cite this version:

Cyril Eraud, Tiphaine Devaux, Alexandre Villers, Fred A Johnson, Charlotte Francesiaz. popharvest: An R package to assess the sustainability of harvesting regimes of bird populations. *Ecology and Evolution*, 2021, 10.1002/ece3.8212 . hal-03457223

**HAL Id: hal-03457223**

**<https://hal.sorbonne-universite.fr/hal-03457223>**

Submitted on 30 Nov 2021

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

## RESEARCH ARTICLE

# popharvest: An R package to assess the sustainability of harvesting regimes of bird populations

Cyril Eraud<sup>1</sup>  | Tiphaine Devaux<sup>1,2,3</sup>  | Alexandre Villers<sup>1</sup>  | Fred A. Johnson<sup>4</sup>  | Charlotte Francesiaz<sup>5</sup> 

<sup>1</sup>Office Français de la Biodiversité, Direction de la Recherche et de l'Appui Scientifique, Unité Avifaune migratrice, Villiers-en-Bois, France

<sup>2</sup>Office de Génie Ecologique, Agence Nord-Est, Strasbourg, France

<sup>3</sup>Centre d'Ecologie et des Sciences de la Conservation – UMR 7204 MNHN-CNRS-UPMC, Muséum National d'Histoire Naturelle, Paris, France

<sup>4</sup>Aarhus University, Department of Bioscience—Kalø, Rønde, Denmark

<sup>5</sup>Office Français de la Biodiversité, Direction de la Recherche et de l'Appui Scientifique, Unité Avifaune migratrice, Ile-D'Olonne, France

## Correspondence

Cyril Eraud, Office Français de la Biodiversité, Direction de la Recherche et de l'Appui Scientifique, Unité Avifaune migratrice, 79360 Villiers-en-Bois, France.  
Email: cyril.eraud@ofb.gouv.fr

## Abstract

Bird harvest for recreational purposes or as a source for food is an important activity worldwide. Assessing or mitigating the impact of these additional sources of mortality on bird populations is therefore crucial issue. The sustainability of harvest levels is however rarely documented, because knowledge of their population dynamics remains rudimentary for many bird species. Some helpful approaches using limited demographic data can be used to provide initial assessment of the sustainable use of harvested bird populations, and help adjusting harvest levels accordingly. The *Demographic Invariant Method* (DIM) is used to detect overharvesting. In complement, the *Potential Take Level* (PTL) approach may allow setting a level of take with regard to management objectives and/or to assess whether current harvest levels meet these objectives. Here, we present the R package `popharvest` that implements these two approaches in a simple and straightforward way. The package provides users with a set of flexible functions whose arguments can be adapted to existing knowledge about population dynamics. Also, `popharvest` enables users to test scenarios or propagate uncertainty in demographic parameters to the assessment of sustainability through easily programming Monte Carlo simulations. The simplicity of the package makes it a useful toolbox for wildlife managers or policymakers. This paper provides them with backgrounds about the DIM and PTL approaches and illustrates the use of `popharvest`'s functionalities in this context.

## KEYWORDS

demographic invariant method, harvest, hunting management, potential take level, sustainability

## 1 | INTRODUCTION

Addressing the demographic consequences of additional—human-induced—sources of mortality on bird populations is a central topic in conservation biology. Both sedentary and migratory birds face a large array of additional sources of mortality to natural ones, with

types and importance varying according to economical or cultural contexts (Drewitt & Langston, 2008; Eason et al., 2016; Hirschfeld & Attard, 2017; Loss et al., 2015; Munilla et al., 2011; Yeh et al., 2013).

Hunting for recreational purposes is one of the major sources of additional mortality. A large number of species are legally exposed to various harvest regimes at continent scales, with millions

This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

© 2021 The Authors. *Ecology and Evolution* published by John Wiley & Sons Ltd.

of birds harvested each year (e.g., Hirschfeld & Attard, 2017). In this context, addressing the demographic consequences of hunting activity is an important issue to assess whether regulatory commitments and underlying conservation objectives are met (e.g., AEWA, 2018; EU Birds Directive, 2009). However, the sustainability of harvest levels is rarely documented which constitutes a source of conflicts between hunters and organizations in charge of conservation.

For numerous species, the lack of detailed knowledge of their full life cycle hinders the ability of assessing harvest sustainability. Demographic information is often limited to coarse estimates for a few parameters (e.g., Dillingham & Fletcher, 2008), and their temporal variance is poorly documented. Detailed information on hunting bags is also limited for most countries. Total harvests by country are often the only existing statistics used to characterize harvest regimes, with no further information about how those estimates were obtained (hence their robustness), and how they balance among sex and/or age-classes (e.g., Hirschfeld & Attard, 2017). This situation may preclude the use of fine-tuned matrix population models. However, alternative approaches using sparse/incomplete data can be used to provide initial assessment of the sustainable use of harvested bird populations, and help wildlife managers or policymakers to adjust harvest levels accordingly.

One of these approaches consists in comparing hunting bags with annual growth rate for a population under optimal conditions, that is, at low density and with no limiting factors (Niel & Lebreton, 2005). Under such a situation, population growth rate ( $\lambda_{\max}$ ) is maximized and the population increases by a proportion of  $\lambda_{\max}-1$  each year (i.e., *maximum recruitment rate*,  $r_{\max}$ ; Dillingham & Fletcher, 2008). Given that such a limitation-free situation is rarely observed in nature, any harvest level that exceeds this potential excess growth is a cue of overharvesting, which may ultimately lead to population depletion or extinction (Robinson, 2000). Niel and Lebreton (2005) further extended this approach, then solutioning the challenge of estimating  $\lambda_{\max}$  under optimal or "rarified" conditions (Johnson et al., 2012). Their procedure, termed the *Demographic Invariant Method* (DIM), allows estimating  $\lambda_{\max}$  on the basis of very few demographic parameters such as adult survival and age at first breeding, considering tight and permanent allometric relationships among demographic parameters. DIM is therefore a useful approach to assess whether some additional sources of mortality like hunting are unsustainable, even in poorly studied species. However, the DIM formulation does not include a priori management objectives in its formulation. Consequently, this approach cannot assess whether current harvest levels are sustainable with regard to the objective of maintaining populations at levels complying with some ecological, economical, or cultural requirements.

A complementary approach that can be used for this purpose is the *Prescribed Take Level* method (PTL, Runge et al., 2009). This approach is a generalization of the *Potential Biological Removal* (PBR) initially developed to assess the sustainability of human-caused

mortality in marine mammals (Wade, 1998). The PTL is grounded on the theory of density-regulated population growth, which can be formulated as a discrete theta-logistic model allowing for linear or nonlinear density dependence (Johnson et al., 2012). The PTL uses properties of this formulation to derive a maximum sustainable yield (MSY), which, over time, drives the population size at an equilibrium point corresponding to a fraction of the carrying capacity ( $K$ , Runge et al., 2009). Then, MSY can be used to set a harvest strategy or can be compared to current hunting bags to assess whether they are sustainable with regard to a predefined management objective. In practice, the PTL approach requires few parameters and is therefore a useful starting point for assessing allowable take within a bird population with limited demographic information. These parameters include population size,  $r_{\max}$  (or  $\lambda_{\max}-1$ ) and a shape parameter describing the shape of the density-dependent relationship. Management objective is a priori specified within the formulation of the PTL as a scaling factor representing the desired population size relative to  $K$  and the take level relative to MSY (Johnson et al., 2012; Koneff et al., 2017; Runge et al., 2009).

In data-poor environments, the DIM and PTL approaches represent two facets of how to detect overharvest and setting allowable take of birds that may satisfy some management or conservation objectives (e.g., Johnson et al., 2012; Koneff et al., 2017; Lormée et al., 2020; Turrin & Watts, 2016; Watts et al., 2015). However, computer tools that implement DIM and PTL algorithms in a flexible and accessible way to a broad audience are lacking. Here, we introduce a user-friendly R package *popharvest*, which allows the conduct of such assessments within an open-source interface and in a flexible R programming language.

## 2 | PACKAGE OVERVIEW

The R package *popharvest* implements the DIM and PTL approaches through a set of highly flexible functions whose arguments can be adapted to existing knowledge on the population dynamics, including situations where demographic data are very limited. In particular, the package makes available to users a series of functionalities derived from the work of Johnson et al. (2012), which can be easily mobilized when some of the key parameters used for the calculations (e.g., survival rate, shape of the density dependence) are virtually unknown. *popharvest* also provides users with a powerful tool for explicitly incorporating uncertainty/stochasticity in calculations, which may help craft more realistic management decisions or may help testing scenarios. This is done by substituting point estimates of demographic parameters by probability distributions (e.g., log-normal, uniform) and performing Monte Carlo simulations by the means of an easy programming syntax. Functions for visualizing, calculating various statistics, and exporting results are also implemented.

### 3 | METHODS AND IMPLEMENTATION

#### 3.1 | Detecting overharvest: the DIM approach

Under optimal conditions, a population increases by an annual proportion equals to  $r_{\max}$  (or  $\lambda_{\max}-1$ ). Accordingly, any level of mortality that exceeds this potential excess growth (i.e., Mortality/Potential excess growth  $>1$ ) is unsustainable. However, we stress that the inverse is not true: Any mortality levels below potential excess growth are not synonymous of sustainability (see Niel & Lebreton, 2005; Robinson, 2000). The function PEG provides users with an algorithm that performs the assessment on the basis of few demographic information.

The potential excess growth (PEG) is calculated following Niel and Lebreton (2005):

$$\text{PEG} = N \times (\lambda_{\max} - 1) \times F_s \quad (1)$$

where  $N$  is the population size and  $\lambda_{\max}$  is the maximal annual growth rate measured under optimal conditions (i.e., at low density and with no limiting factors).  $F_s$  is a «safety» parameter taking values within the [0-1] range, which allows users taking only a proportion of the maximal PEG when performing their assessments of sustainability. Several arguments are given in Niel and Lebreton (2005) justifying its use to reduce the risk of drawing wrong conclusions. For instance, a species or population may be exposed to other unquantified sources of additional mortality, which may accumulate to value above maximum PEG. Current harvests may also be imbalanced toward age-classes with the highest reproductive value, then increasing the potential for negative demographic consequences. In addition, populations exposed to harvest can be small and/or subject to high environmental stochasticity, then needing a conservative approach. As a rule,  $F_s$  needs to be a priori defined and should not exceeds 0.5 (Niel & Lebreton, 2005; Wade, 1998). One optional approach is to set  $F_s$  according to species' IUCN conservation status (Dillingham & Fletcher, 2008): 0.5 for «Least concern (LC)» species, 0.3 for «Near threatened (NT)» and 0.1 when they are categorized as «Vulnerable (VU)».

Having  $N$ ,  $\lambda_{\max}$ , the mortality level and setting  $F_s$ , the user supplies the corresponding values to arguments listed in the overharvest function (respectively: `pop.fixed`, `lambdaMax`, `harvest.fixed`, `Fs`). The user can specify up to 3 values for  $F_s$ , which allows estimations to be made for different management/conservation scenarios for one species/population (`NSp=1`). The function `output.summary()` returns a *sustainable harvest index* (SHI) calculated as the ratio Mortality/Potential excess growth. A value  $>1$  suggests unsustainability.

```
PEG(pop.fixed=10000,NSp=1,Fs=c(0.1,0.2,0.3),lambdaMax.fixed=1.35, harvest.fixed=1500)
```

In most cases,  $\lambda_{\max}-1$  (or  $r_{\max}$ ) is an unknown quantity as populations face limiting factors and consequently grow under nonoptimal

conditions. Niel and Lebreton (2005) hence developed a procedure to estimate  $\lambda_{\max}$ , taking advantage of the relationship linking  $\lambda_{\max}$  to (optimal) generation time. Under this procedure,  $\lambda_{\max}$  is derived from estimates of the maximal (annual) adult survival ( $S_a$ ) and the age at first reproduction ( $\alpha$ , in year). The function PEG allows users to estimate  $\lambda_{\max}$  using this approach, by simply replacing the argument `lambdaMax` by the arguments `surv.` ( $S_a$ ) and `alpha.` ( $\alpha$ ). One subtlety of the procedure is that it needs to be adapted to the species' life-history strategy (Niel & Lebreton, 2005).

For long-lived species,  $\lambda_{\max}$  is estimated by solving:

$$\lambda_{\max} \approx \frac{(s\alpha - s + \alpha + 1) + \sqrt{(s - s\alpha - \alpha - 1)^2 - 4s\alpha^2}}{2\alpha} \quad (2)$$

For short-lived species,  $\lambda_{\max}$  is estimated by numerically solving:

$$\lambda_{\max} = \exp \left[ \left( \alpha + \frac{s}{\lambda_{\max} - s} \right)^{-1} \right] \quad (3)$$

The argument `living.rate="short"` or `"long"` is then passed to the function to call one of these two formulas.

```
PEG(pop.fixed=10000, NSp=1, Fs=0.3, surv.fixed=0.80, alpha.fixed=1, living.rate="short", harvest.fixed=1500)
```

Calculations for  $n$  species or populations can simultaneously be performed in a single run. The argument `NSp=n` is passed to the function, and the values of the function arguments are then filled as a vector of length  $n$  in which their order follows the sequence of the  $n$  studied populations/species.

```
PEG(NSp=2, living.rate = c("long", "short"), surv.fixed = c(0.8, 0.65), alpha.unif = TRUE, min.alpha = c(2, 1), max.alpha = c(2, 1), pop.fixed = c(3600000, 55000000), harvest.fixed = c(100000, 8000000), Fs = c(0.1, 0.3, 0.5))
```

One of the assumptions of the DIM approach is that survival estimates correspond to the maximum values ( $S_a$ ) measured under optimal conditions. If  $S_a$  is underestimated, this increases the resulting  $\lambda_{\max}$  and consequently, the PEG value used as a comparison threshold with current harvest levels. Banding and capture–recapture designs are common tools to assess survival in free-ranging birds, but careful consideration must be given to their accuracy. For instance, emigration may negatively bias survival estimates derived from live-recaptures (Ergon & Gardner, 2014). Some approaches may overcome this problem, such as spatial capture–recapture (Ergon & Gardner, 2014) or dead recovery models fitted at large geographic scales (Brownie et al., 1985). However, insofar as the species or population is exposed to harvest, the observed survival may be lower than the true survival achieved in the absence of this additional

source of mortality. Using the upper interval bounds of estimates of adult survival may be an option. However, there are situations where no estimates of survival are available or, alternatively, those existing are not useful because they rely upon obsolete or inappropriate sampling designs/methodologies (e.g., life tables in old 1960s publications).

One solution to this problem is that, in birds,  $S_a$  can be estimated from an allometric relationship linking maximal adult survival and species body mass. This relationship was formulated by Johnson et al. (2012) on the basis of survivorship data measured on captive birds (Ricklefs, 2000):

$$S_a = \left( \frac{p}{S_j} \right)^{1/(\exp[3.22+0.24 \times \ln(M)+e]-a)} \quad (4)$$

$S_j$  is (average) annual survival for birds aged 1 year to  $\alpha$ , and  $p$  is defined from a beta distribution  $\sim\beta(3.34, 101.24)$ .  $M$  is body mass (in kg),  $e$  are normally distributed residuals  $\sim N(0, \sigma^2 = 0.087)$ , and  $\alpha$  is age at first breeding (in years). This solution is implemented in the PEG function. To do that, the user replaces the argument related to survival (`surv.fixed`) by `mass.fixed`. With the arguments `type.p` and `type.e="determinist"`,  $p$  and  $e$  are set at their means (i.e.,  $3.34/(3.34+101.24)$  and 0). When `type.p` and `type.e="random"` are used,  $p$  and  $e$  are sampled within their respective distributions (Table 1).  $\alpha$  is specified either as a fixed value (`alpha.fixed`) or drawn from a uniform (`alpha.unif`) or a log-normal (`alpha.lognorm`) distribution. By default, the  $S_j$  argument is silent (i.e.,  $S_j = 1$ ), assuming that the situation  $S_j < S_a$  does not hold for birds aged 1 to  $\alpha$ . In contrary case, the user can provide a value for  $S_j$  by using the (fixed) argument `surv.j`. Full details on estimating survival from body mass are provided in Appendix 1.

```
PEG(pop.fixed=10000, NSp= 1, Fs=0.3, mass.
fixed=0.125, type.p = "random", type.e = "random",
alpha.fixed=1, living.rate="short", harvest.fixed =
1500)
```

The high flexibility of the PEG function allows users to explicitly incorporate uncertainty in parameter values when performing assessment of the sustainable use of a species/population (Figure 1). Survival can be specified as following a beta distribution  $Beta(\alpha, \beta)$  and all others parameters (i.e., population size,  $\lambda_{max}$ , age at first reproduction, body mass, harvest level) as following a log-normal distribution  $\ln N(\mu, \sigma)$ , whose associated parameters are specified by the method of moments. Population size, age at first breeding, and harvest level can also be specified as following continuous uniform distributions  $U(\text{lower}, \text{upper}; \text{Table 1})$ , since these parameters are commonly published as ranges of values. The `Nsim` argument is then passed to the function to specify the number of Monte Carlo simulations. At each simulation, the function samples within the distributions specified for each argument and then performs calculations. Each simulation and underlying parameters are stored in a data frame (e.g., `df`).

```
df<-PEG(Nsim=100000,NSp=1,Fs=0.3,pop.unif=TRUE,min.
pop=10000, max.pop=20000,surv.beta=TRUE,mean.surv=
0.65,sd.surv=0.08,alpha.lognorm=TRUE,mean.alpha=
1.2,sd.alpha=0.02,living.rate="short",harvest.
unif=TRUE, min.harvest=1000,max.harvest=2000)
```

The `output.summary(df)` function provides summary statistics (mean, median, 95%CI) for relevant parameters including  $r_{max}$  (`Rmax`), potential excess growth (PEG), and the sustainable harvest index (SHI). The function also returns the percentage (with 1 = 100%) of simulations (`unsustain.harvest`) for which mortality level exceeds the potential excess growth (i.e.,  $SHI > 1$ ). The function `SHI.plot(df)` function allows users figuring a distribution histogram of the simulated sustainable harvest indices (SHI, Figure 2).

### 3.2 | Setting allowable take levels: the PTL approach

The growth of a population exposed to additional mortality like harvest can be formulated as a theta-logistic model, which includes density-dependent regulation mechanisms:

$$N_{t+1} = N_t + N_t \times r_{max} \times \left[ 1 - \left( \frac{N_t}{K} \right)^\theta \right] - H_t \quad (5)$$

$N$  is the population size,  $r_{max}$  is the maximal annual recruitment rate,  $K$  is the population carrying capacity,  $\theta$  is a shape parameter describing the shape of the density-dependent function, and  $H$  is the harvest level. From this formulation, a potential and sustainable harvest level can be derived following Johnson et al. (2012):

$$PTL_t = F_{obj} \times \left( \frac{r_{max}^\theta}{(\theta + 1)} \right) \times N_t \quad (6)$$

A critical assumption of the PTL approach is that population growth rates compensate for additional mortality through density-dependent mechanisms (O'Brien et al., 2017). Assuming a logistic growth and a linear form of density dependence (i.e.,  $\theta = 1$ ), a maximum sustainable yield (MSY) is achieved at  $r_{max}K/4$  or at a rate  $r_{max}/2$ . With such a rate, the population is expected to equilibrate at one-half of its carrying capacity ( $K/2$ , Runge et al., 2009). If density dependence follows a concave ( $\theta < 1$ ) or a convex ( $\theta > 1$ ) form, the sustainable rate, which maximizes harvest level (MSY) is  $(r_{max}^\theta)/(\theta + 1)$  and the population size is expected to converge to an equilibrium point (Johnson et al., 2012) equal to:

$$N_e = K \times \left[ 1 - F_{obj} \times \frac{\theta}{(\theta + 1)} \right]^{1/\theta} \quad (7)$$

PTL is a useful approach to set allowable take level but also to assess whether current harvest levels are sustainable with regard to management objectives (i.e., Mortality/Potential take level  $< 1$ ). The PTL function provides users with an algorithm that performs

these calculations on the basis of the demographic traits  $N$ ,  $r_{\max}$ ,  $\theta$ , and current harvest levels. Architecture and syntax of this function are identical in all respects to the PEG function. Again,  $r_{\max}$  can be approximated as  $r_{\max} \approx \lambda_{\max} - 1$  (Dillingham & Fletcher, 2008; Runge et al., 2009), so users can take advantage of all the optional ways to estimate  $\lambda_{\max}$  (see Section 3.1). Theta (i.e.,  $\theta$ ) is an additional argument, which describes the shape of the density dependence function. When  $\theta$  is unknown, users may set a priori values (e.g., Koneff et al., 2017) using the arguments `theta.fixed` or alternatively estimate it from the equation established by Johnson et al. (2012, from Saether & Engen, 2002), relating  $\theta$  to  $r_{\max}$  (`estim.theta`):

$$\ln(\theta) = 1.129 - 1.824 \times r_{\max} + e \quad (8)$$

where  $e$  are normally distributed residuals  $-N(0, \sigma^2 = 0.942)$ . Here, users have two options: setting a deterministic relationship ( $e = 0$ , argument `estim.theta="determinist"`) or taking into account errors by sampling within the distribution of residuals (argument `estim.theta="random"`; Table 1).

```
PTL(pop.fixed=10000,NSp=1,Fobj=c(0.1,0.3,0.5),surv.
fixed=0.80,alpha.fixed=1,living.rate="short",theta.
fixed=1,harvest.fixed=1500)
```

The `output.summary` function returns PTL values and a *sustainable harvest index* (SHI), which assesses whether current harvest levels are sustainable with regard to management objectives. SHI is calculated as the ratio Harvest/PTL, with value  $>1$  indicating unsustainability.

An important parameter of the PTL formulation is the scaling factor  $F_{\text{obj}}$ , which represents management objectives. This argument is passed to the function as `Fobj`. It is up to the users to fix  $F_{\text{obj}}$  values according to the «desired long-term population size relative to the carrying capacity» (Runge et al., 2009).  $F_{\text{obj}}$  can take all values along the gradient  $0 < F_{\text{obj}} < (\theta + 1)/\theta$ . With  $F_{\text{obj}} = 1$ , the harvest strategy seeks achieving maximum sustained yield (MSY) and holding population size at  $N_e$  (Equation 7). With  $F_{\text{obj}} < 1$ , the harvest strategy is more conservative, taking only a fraction of MSY. When  $F_{\text{obj}}$  values approach zero, very limited harvest is allowed and population size is expected to reach its carrying capacity. At the opposite end of the gradient, when  $F_{\text{obj}}$  approaches  $(\theta + 1)/\theta$ , the harvest strategy is to hold the population at a small fraction of its carrying capacity (Johnson et al., 2012; Runge et al., 2009). Any value of  $F_{\text{obj}}$  produces a “sustainable” harvest strategy. Here, users should be aware that this concept has a different meaning than for the DIM approach (cf. 3.1). Under the PTL approach, a sustainable strategy corresponds to harvest levels that allow management objectives to be met, which does not necessarily include a reduction in the risk of population decline. Even  $F_{\text{obj}}$  close to  $(\theta + 1)/\theta$  may reflect a suitable strategy such as eradicating exotic pest species. Users should be aware that when harvest levels exceeds MSY, population equilibria are unstable. Uncontrolled variation in population size can lead either to

collapse or to a growing population in the face of a constant harvest. If avoiding resource depletion is a prominent objective,  $F_{\text{obj}}$  should be set while taking into account important aspects, including the frequency at which population size is updated, the population carrying capacity, the strength of density dependence, the magnitude of environmental stochasticity, or the existence of additional sources of mortality (see Johnson et al., 2012; O'Brien et al., 2017).

The PTL function provides users with similar flexibility and simulation capabilities than the PEG function (Table 1, Figure 1). The `output.summary()` function also provides with similar summary statistics (mean, median, 95%IC) for relevant parameters including  $r_{\max}$  (`Rmax`), potential take level (PTL), the sustainable harvest index (SHI), and the percentage of simulations (`unsustain.harvest`) for which a harvest level exceeds the potential take level. The function `SHI.plot()` allows users to plot the simulated sustainable harvest indices calculated under the PTL approach. Detailed examples are provided as Appendix S1.

## 4 | CONCLUSION

`popharvest` provides wildlife managers and policymakers with a useful toolbox to assess or mitigate the impact of harvest regimes on bird populations. The package uses an intuitive programming language, allowing R beginners to easily perform analyses, including testing management scenarios for a set of parameters (e.g., best vs. worst scenarios) and incorporating uncertainty/stochasticity in calculations through Monte Carlo simulations. Accounting for uncertainty is a key component to limit the risk of taking wrong management decisions (Johnson et al., 2012). The propagating of uncertainty in the assessment process is one strength of the package.

In this note, both the DIM and PTL approaches are briefly described, providing users with an overview of their potential and how they are implemented in the `popharvest` R package. We refer users to the original articles in gauging the limits, advantages, and assumptions inherent to the methods covered by `popharvest` (e.g., Johnson et al., 2012; Niel & Lebreton, 2005; O'Brien et al., 2017; Runge et al., 2009; Wade, 1998). For instance, several important aspects and limitations are not further discussed, including the importance of having reliable and frequent estimates for both population sizes and harvest levels, the need for a relevant delineation of management units (especially for migrants), or the importance of taking into account all additional mortality sources in calculations. In the PTL approach, the continuum of possible population responses to harvest levels (compensatory-additivity/depensatory, Péron, 2013) is also overlooked, as well as some components of the underlying population dynamics (e.g., demographic stochasticity). Most often, these aspects are either imprecise or unknown in a context of limited demographic data. Accordingly, users should be aware that making some assumptions and using some key parameters/hypotheses will influence the output of the functions, with potentially important consequences on management decision. For example, using underestimated survival values that do not correspond to those observed

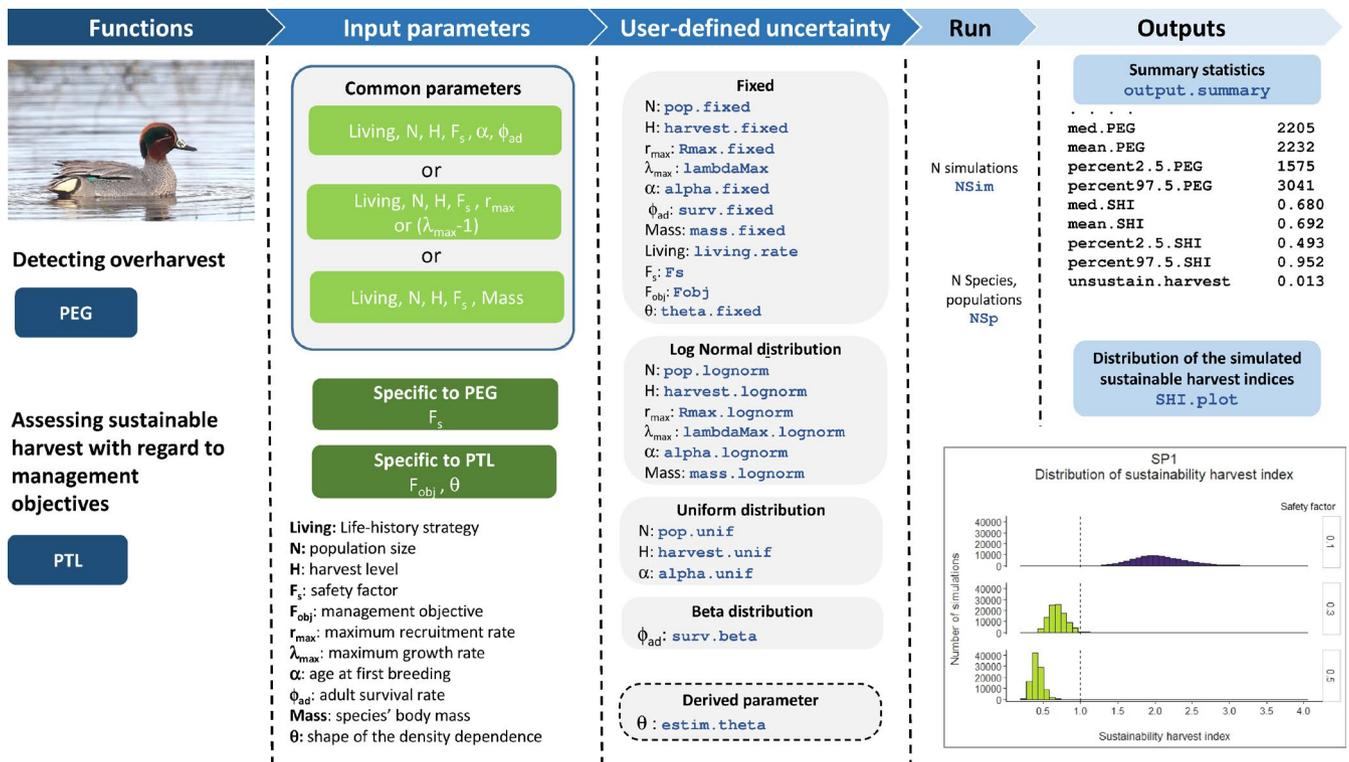
TABLE 1 List and description of arguments for the PEG<sup>(1)</sup> and PTL<sup>(2)</sup> functions

Arguments	Argument types	Description	Examples
Pop.fixed <sup>(1,2)</sup>	Integer	Point estimate of population size	Pop.fixed=10000
pop.unif <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to draw population size from a uniform distribution bounded with minimum (min.pop) and maximum (max.pop) values	pop.unif=TRUE, min.pop=6000, max.pop=12500
pop.lognorm <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to draw population size from a log-normal distribution $\ln N(\text{mean.pop}, \text{sd.pop})$	pop.lognorm=TRUE, mean.pop=6000, sd.pop=130
Rmax.fixed <sup>(1,2)</sup>	Decimal	Point estimate of maximal annual recruitment rate	Rmax.fixed=0.79
Rmax.lognorm <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to draw $R_{\max}$ from a log-normal distribution $\ln N(\text{mean. } R_{\max}, \text{sd. } R_{\max})$	Rmax.lognorm=TRUE, mean. Rmax=XX, sd. Rmax=XX
lambdaMax <sup>(1,2)</sup>	Decimal	Point estimate of maximal annual growth rate	lambdaMax=1.35
lambdaMax.lognorm <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to draw lambda max from a log-normal distribution $\ln N(\text{mean. lambdaMax}, \text{sd. lambdaMax})$	lambdaMax.lognorm=TRUE, mean.lambdaMax=6000, sd.lambdaMax=130
surv.fixed <sup>(1,2)</sup>	Decimal (range 0–1)	Point estimate of adult annual survival	surv.fixed=0.80
surv.beta <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to draw survival from a beta distribution. The function uses the method of moments to specify associated parameters from mean and sd.	surv.beta=TRUE, mean.surv=0.80, sd.surv=0.12
surv.j <sup>(1,2)</sup>	Decimal (range 0–1)	Point estimate of the (average) annual survival ( $S_j$ ) for birds aged 1 to $\alpha$ (age at first breeding). Argument used when estimating adult survival from body mass (i.e., to calculate $p/S_j$ ) in situation when $S_j < S_a$ (adult survival) for birds aged 1 to $\alpha$ . Default is silent, assuming $S_j < S_a$ solely for birds aged <1 year.	surv.j = 0.66
alpha.fixed <sup>(1,2)</sup>	Decimal	Point estimate for the age at first breeding	alpha.fixed=1
alpha.unif <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to draw age at first breeding from a uniform distribution bounded with minimum (min.alpha) and maximum (max.alpha) values	alpha.unif=TRUE, min.alpha=1, max.alpha=2
alpha.lognorm <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to draw age at first breeding from a log-normal distribution $\ln N(\text{mean.alpha}, \text{sd.alpha})$	alpha.lognorm=TRUE, mean.alpha=1.2, sd.alpha=0.3
mass.fixed <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to estimate adult survival from point estimate of body mass (in kg)	mass.fixed=0.174
mass.lognorm <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to estimate adult survival from estimates of body mass (in kg) drawn from a log-normal distribution $\ln N(\text{mean. weight}, \text{sd.weight})$	mass.lognorm=TRUE, mean.mass=0.174, sd.mass=0.018
type.p <sup>(1,2)</sup>	Character string	Can either be "determinist" or "random." When it is "determinist," calls the function to estimate adult survival from body mass while keeping parameter $p$ equal to $3.34/(3.34 + 101.24)$ . When it is "random," calls the function to estimate adult survival from body mass while sampling the parameter $p$ of the underlying relationship within the beta distribution $\beta(3.34, 101.24)$	type.p = "determinist"
type.e <sup>(1,2)</sup>	Character string	Can either be "determinist" or "random." When it is "determinist," calls the function to estimate adult survival from body mass while keeping residuals of the underlying relationship at their means (i.e., 0). When it is "random," calls the function to estimate adult survival from body mass while sampling residuals of the underlying relationship within the normal distribution $N(0, 0.087)$	type.e = "determinist"
harvest.fixed <sup>(1,2)</sup>	Integer	Point estimate of take (mortality) level	harvest.fixed=1500
harvest.unif <sup>(1,2)</sup>	Boolean (Default =FALSE)	Calls the function to draw take level from a uniform distribution bounded with minimum (min.harvest) and maximum (max.harvest) values	harvest.unif=TRUE, min.harvest=1000, max.harvest=2000
harvest.lognorm <sup>(1,2)</sup>	Boolean (Default = FALSE)	Calls the function to draw estimates of take level from a log-normal distribution $\ln N(\text{mean.harvest}, \text{sd.harvest})$	harvest.lognorm=TRUE, mean.harvest=1500, sd.harvest=250

(continues)

TABLE 1 (continues)

Arguments	Argument types	Description	Examples
theta.fixed <sup>(2)</sup>	Decimal	Point estimate of the shape parameter describing the shape of the density-dependent function	theta.fixed=1
estim.theta <sup>(2)</sup>	Character string	Can either be “determinist” or “random.” When it is “determinist,” calls the function to estimate theta from $r_{max}$ while keeping residuals of the underlying relationship at their means (i.e., 0). When it is “random,” calls the function to estimate theta from $r_{max}$ while sampling residuals of the underlying relationship within the normal distribution $N(0, 0.942)$	estim.theta = “determinist”
full.option	Boolean (Default = FALSE)	If TRUE, provides all parameter values used in calculations.	full.option=TRUE

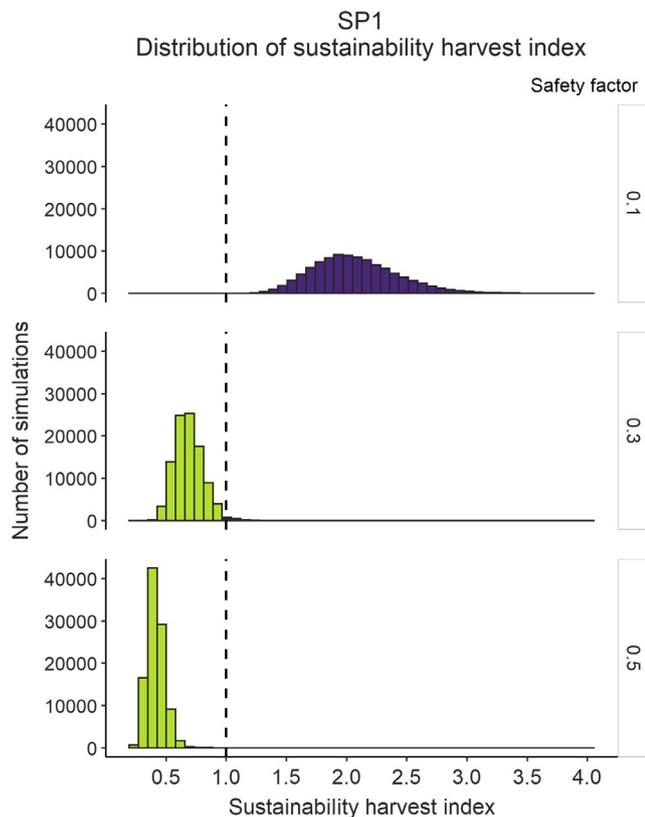


**FIGURE 1** Workflow, functionalities, and main arguments of the package `popharvest` for assessing the sustainability of harvesting regimes of bird populations. (a) Input parameters are filled by users. The functions accommodate for a priori knowledges of population dynamics and allow for some missing parameters ( $r_{max}$ ,  $\lambda_{max}$ ,  $\phi$ ,  $\theta$ ) to be estimated. Some input parameters are common to PEG and PTL functions, and others are functions specific. (b) Uncertainty for some input parameters is specified by users either by setting fixed values or simulating values drawn within a priori distributions. (c) Results are provided as data frames or histograms of sustainable harvest indices distribution. Arguments listed in this figure are detailed in Table 1

under optimal conditions will result in overestimating growth rates and consequently the values of sustainable harvest levels. One option may be to consider a conservative approach, whether in the choice of the initial values of safety/objective factors (i.e.,  $F_s$ ,  $F_{obj}$ ), for some parameters, or in assumptions about the demographic functioning of the populations of the studied species (e.g., shape of density dependence). For instance, Wade (1998) recommends using a minimal value for population size (i.e.,  $N_{min}$ ) equal to the lower bound of a 60% confidence interval (see also Dillingham & Fletcher, 2008; Runge et al., 2009). The `popharvest` package enables a quick

and easy assessment of a multitude of scenarios including different initial values for population sizes, different shapes and strengths of density dependence, or different maximum population growth rates. Such scenarios can help inform debates about the sustainability of the exploitation of species for which little knowledge about their demography exists.

Estimates of  $\lambda_{max}$  derived from the approach of Niel and Lebreton (2005) are based on two equations, which apply differentially depending on whether the studied species is considered as long-lived (Equation 2) or short-lived (Equation 3). Choosing between these two



**FIGURE 2** Density distribution of the simulated sustainable harvest indices under three scenarios for the safety factor. Results are from 100,000 Monte Carlo simulations performed using PEG function for a short-lived species, including uncertainty for various initial parameters (see example “*dim.5.un*,” Appendix S1)

categories can be tricky, especially in the context of limited demographic data, but users should be aware that this can strongly affect  $\lambda_{\max}$  estimates. For the same value of alpha ( $\alpha$ : age at first reproduction),  $\lambda_{\max}$  calculated for a given species will be lower if the species is categorized as “long-lived.” Simulations show that the magnitude of the difference may turn out to be very large, in particular for low values of  $\alpha$ . For example, with survival = 0.7 and  $\alpha = 3$  years,  $\lambda_{\max}$  will be, respectively, estimated at 1.27 and 1.23, depending on whether the species is considered as short- or long-lived. More significant deviations are observed when  $\alpha = 1$  (with  $s = 0.7$ ,  $\lambda_{\max} = 1.87$  vs. 1.55). However, with  $\alpha = 1$ , a species can be categorized unambiguously as short-lived, but uncertainty may arise for species with intermediate life-history strategies. Niel and Lebreton (2005) do not provide explicit arguments on the threshold used to assign a species to either categories. In their original works, “short-lived” species that warranted the use of Equation (3) were only those reproducing at 1 year (i.e., Great Tit *Parus major*, Rock Sparrow *Petronia petronia*). Thus, the use of Equation (3) only for species that breed at 1 year can be viewed as a sensible and conservative option.

We anticipate that `popharvest` will encourage a wider use of the DIM and PTL approaches to perform initial assessments of the consequences of harvest regimes on certain species and/or populations, and to prioritize research works or management actions. The

package allows wildlife conservationists and managers to easily share scenarios (e.g., best vs. worst) and analyses using a common language that does not require an in-depth mastery of the R syntax. In addition, the ease of use of the `popharvest` package ensures both the reproducibility of the analyses and the transparency about assumptions and initial values assigned to the various key parameters.

It is noteworthy that although `popharvest` focuses primarily on hunted birds, some functionalities can be used in other contexts (e.g., poaching, Brochet et al., 2016, incidental fatalities), or even for nonavian models if  $\lambda_{\max} - 1$  (or  $r_{\max}$ ) is a known quantity.

## ACKNOWLEDGMENTS

We are grateful to Matthieu Guillemain, Olivier Duriez, Alexis Carron, and Maxime Lahournat for their help, fruitful discussions, and/or careful readings of first drafts. Revised version of this article greatly benefited from the comments and suggestions of the Associate Editor and two reviewers. Photo courtesy of Ferry Duyvelaar.

## CONFLICT OF INTEREST

All authors declare no conflict of interest.

## AUTHOR CONTRIBUTIONS

**Cyril Eraud:** Conceptualization (equal); Methodology (equal); Software (equal); Writing-original draft (lead). **Tiphaine Devaux:** Conceptualization (equal); Methodology (equal); Software (lead); Writing-original draft (equal). **Alexandre Villers:** Methodology (supporting); Software (supporting); Writing-original draft (equal). **Fred A. Johnson:** Conceptualization (equal); Methodology (equal); Software (equal); Writing-original draft (equal). **Charlotte Francesiaz:** Conceptualization (equal); Methodology (equal); Software (supporting); Writing-original draft (equal).

## DATA AVAILABILITY STATEMENT

R scripts and examples are provided as Appendix S1. The package (version 1.0.0; <https://doi.org/10.5281/zenodo.5522675>) and help files are available on Github (<https://github.com/popharvest/popharvest>).

## ORCID

Cyril Eraud  <https://orcid.org/0000-0003-1517-9835>

Tiphaine Devaux  <https://orcid.org/0000-0001-7285-8805>

Alexandre Villers  <https://orcid.org/0000-0003-3050-4106>

Fred A. Johnson  <https://orcid.org/0000-0002-5854-3695>

Charlotte Francesiaz  <https://orcid.org/0000-0002-2944-695X>

## REFERENCES

- AEWA (2018). *Agreement on the Conservation of African-Eurasian Migratory Waterbirds* (AEWA). *Agreement Text and Annexes As amended by MOP 7*. UNEP/AEWA Secretariat.
- Brochet, A.-L., Van den bossche, W., Jbour, S., Ndang'ang'a, P. K., Jones, V. R., Abdou, W. A. L. I., Al-hmoud, A. R., Asswad, N. G., Atienza, J. C., Atrash, I., Barbara, N., Bensusan, K., Bino, T., Celada, C., Cherkaoui, S. I., Costa, J., Deceuninck, B., Etayeb, K. S., Feltrup-azafaf, C., ... Butchart, S. H. (2016). Preliminary assessment of the scope and scale of illegal killing and taking of birds in the Mediterranean. *Bird*

- Conservation International*, 26, 1–28. <https://doi.org/10.1017/S0959270915000416>
- Brownie, C., Anderson, D. R., Burnham, K. P., & Robson, D. S. (1985). *Statistical inference from band recovery data – a handbook*. United States Department of the Interior Fish and Wildlife Service Resource Publication No. 156, Washington, DC.
- Dillingham, P. W., & Fletcher, D. (2008). Estimating the ability of birds to sustain additional human-caused mortalities using a simple decision rule and allometric relationships. *Biological Conservation*, 141, 1783–1792. <https://doi.org/10.1016/j.biocon.2008.04.022>
- Directive 2009/147/CE du Parlement européen et du Conseil du 30 novembre 2009 concernant la conservation des oiseaux sauvages (JO L 20 du 26.1.2010, pp. 7–25).
- Drewitt, A. L., & Langston, R. H. W. (2008). Collision effects of wind-power generators and other obstacles on birds. *Year in Ecology and Conservation Biology*, 1134, 233–266. <https://doi.org/10.1196/annals.1439.015>
- Eason, P., Rabia, B., & Attum, O. (2016). Hunting of migratory birds in North Sinai, Egypt. *Bird Conservation International*, 26, 39–51. <https://doi.org/10.1017/S0959270915000180>
- Ergon, T., & Gardner, B. (2014). Separating mortality and emigration: Modelling space use, dispersal and survival with robust-design spatial capture–recapture data. *Methods in Ecology and Evolution*, 5, 1327–1336. <https://doi.org/10.1111/2041-210X.12133>
- Hirschfeld, A., & Attard, G. (2017). Bird hunting in Europe. An analysis of bag figures and their effect on the conservation of threatened species. *Berichte Zum Vogelschutz*, 53(54), 15–42.
- Johnson, F. A., Walters, M. A. H., & Boomer, G. S. (2012). Allowable levels of take for the trade in Nearctic songbirds. *Ecological Applications*, 22, 1114–1130. <https://doi.org/10.1890/11-1164.1>
- Koneff, M. D., Zimmerman, G. S., Dwyer, C. P., Fleming, K. K., Padding, P. I., Devers, P. K., Johnson, F. A., Runge, M. C., & Roberts, A. J. (2017). Evaluation of harvest and information needs for North American sea ducks. *PLoS One*, 12, e0175411. <https://doi.org/10.1371/journal.pone.0175411>
- Lormée, H., Barbraud, C., Peach, W., Carboneras, C., Lebreton, J.-D., Moreno-Zarates, L., Bacon, L., & Eraud, C. (2020). Assessing the sustainability of harvest of the European Turtle-dove along the European western flyway. *Bird Conservation International*, 30, 506–521. <https://doi.org/10.1017/S0959270919000479>
- Loss, S. R., Will, T., & Marra, P. P. (2015). Direct mortality of birds from anthropogenic causes. *Annual Review of Ecology, Evolution, and Systematics*, 46, 99–120. <https://doi.org/10.1146/annurev-ecolsys-112414-054133>
- Munilla, I., Arcos, J. M., Oro, D., Álvarez, D., Leyenda, P. M., & Velando, A. (2011). Mass mortality of seabirds in the aftermath of the Prestige oil spill. *Ecosphere*, 2, 1–14. <https://doi.org/10.1890/ES11-00020.1>
- Niel, C., & Lebreton, J.-D. (2005). Using demographic invariants to detect overharvested bird populations from incomplete data. *Conservation Biology*, 19, 826–835. <https://doi.org/10.1111/j.1523-1739.2005.00310.x>
- O'Brien, S. H., Cook, A. S. C. P., & Robinson, R. A. (2017). Implicit assumptions underlying simple harvest models of marine bird populations can mislead environmental management decisions. *Journal of Environmental Management*, 201, 163–171. <https://doi.org/10.1016/j.jenvman.2017.06.037>
- Péron, G. (2013). Compensation and additivity of anthropogenic mortality: life-history effects and review of methods. *Journal of Animal Ecology*, 82, 408–417. <https://doi.org/10.1111/1365-2656.12014>
- Ricklefs, R. E. (2000). Intrinsic aging-related mortality in birds. *Journal of Avian Biology*, 31, 103–111. <https://doi.org/10.1034/j.1600-048X.2000.210201.x>
- Robinson, J. G. (2000). Calculating maximum sustainable harvests and sustainable offtakes. In J. G. Robinson & E. L. Bennett (Eds.), *Hunting for sustainability in tropical forests* (pp. 521–524). Columbia University Press.
- Runge, M. C., Sauer, J. R., Avery, M. L., Blackwell, B. F., & Koneff, M. D. (2009). Assessing allowable take of migratory birds. *The Journal of Wildlife Management*, 73, 556–565. <https://doi.org/10.2193/2008-090>
- Saether, B., & Engen, S. (2002). Pattern of variation in avian population growth rates. *Philosophical Transactions of the Royal Society of London. Series B: Biological Sciences*, 357, 1185–1195. <https://doi.org/10.1098/rstb.2002.1119>
- Slade, N. A., Gomulkiewicz, R., & Alexander, H. M. (1998). Alternatives to Robinson and Redford's method of assessing overharvest from incomplete demographic data. *Conservation Biology*, 12, 148–155. <https://doi.org/10.1111/j.1523-1739.1998.96273.x>
- Turrin, C., & Watts, B. D. (2016). Sustainable mortality limits for migratory shorebird populations within the East Asian-Australian Flyway. *Stilt*, 68, 2–17.
- Wade, P. R. (1998). Calculating limits to the allowable human-caused mortality of cetaceans and pinnipeds. *Marine Mammal Science*, 14, 1–37. <https://doi.org/10.1111/j.1748-7692.1998.tb00688.x>
- Watts, B. D., Reed, E. T., & Turrin, C. (2015). Estimating sustainable mortality limits for shorebirds using the Western Atlantic Flyway. *Wader Study*, 122, 37–53. <https://doi.org/10.18194/ws.00005>
- Yeh, Y. M., Huang, H. W., Dietrich, K. S., & Melvin, E. (2013). Estimates of seabird incidental catch by pelagic longline fisheries in the South Atlantic Ocean. *Animal Conservation*, 16, 141–152. <https://doi.org/10.1111/j.1469-1795.2012.00588.x>

## SUPPORTING INFORMATION

Additional supporting information may be found in the online version of the article at the publisher's website.

**How to cite this article:** Eraud, C., Devaux, T., Villers, A., Johnson, F. A., & Francesiaz, C. (2021). `popharvest`: An R package to assess the sustainability of harvesting regimes of bird populations. *Ecology and Evolution*, 00, 1–10. <https://doi.org/10.1002/ece3.8212>

## APPENDIX 1

## ESTIMATING SURVIVAL FROM BODY MASS

The mathematical formulation linking adult survival under ideal conditions ( $S_a$ ) and body mass is detailed in Johnson et al. (2012). This approach is grounded on the Slade et al. (1998)'s assumption that 1% of a new cohort survives to senescence. With  $\alpha$  = age at first breeding,  $\omega$  = age at last breeding (or maximum longevity) and  $S_j$  = annual survival of all birds of age  $< \alpha$  (i.e. from fledging to  $\alpha$ ), the original Slade et al.'s formula takes the form:

$$s_a = \left( \frac{0.01}{S_j} \right)^{1/(\omega-\alpha)} \quad (\text{A1})$$

Johnson et al. (2012) have shown that the fraction of a cohort surviving to senescence ( $p$ ) may achieve upper values, ranging from 0.01 to 0.08. Accordingly, they reformulated the original equation by substituting the 1% rule (i.e.  $p = .01$ ) by the mean of their estimates ( $p = .03$ ; SD: 0.017):

$$s_a = \left( \frac{0.03}{S_j} \right)^{1/(\omega-\alpha)} \quad (\text{A2})$$

Because  $p$  estimates were calculated using only data for birds aged  $\geq 1$  year (see Johnson et al., 2012),  $S_j$  can be ignored when  $\alpha = 1$ . Hence:

$$s_a = p^{1/(\omega-\alpha)} \quad (\text{A3})$$

According to Johnson et al. (2012),  $p$  is assumed following a beta distribution whose shape parameters ( $\alpha = 3.34$ ,  $\beta = 101.24$ ) are specified from mean and variance of  $p$  ( $p = .03$ ; SD: 0.017) and the method of moments. This formulation is neither conditional on mass nor on age at sexual maturity. The link between  $s_a$  and body mass is provided by the relationship between body mass ( $M$ , in kg) and maximum longevity ( $\omega$ ). In birds, Johnson et al. (2012) have shown that this relationship follows:

$$\omega = \exp(3.22 + 0.24 \times \ln(M)) \quad (\text{A4})$$

Having  $p \approx \beta(3.34, 101.24)$ , adult survival under ideal conditions ( $s_a$ ) is estimated as:

$$s_a = p^{1/(\exp[3.22+0.24 \times \ln(M)]-\alpha)} \quad (\text{A5})$$

This formula is correct for any  $\alpha$  if, and only if,  $s_j < s_a$  solely for birds aged  $< 1$  year. This is a reasonable approximation especially in passerines and some other medium-sized bird species.

If, however,  $s_j < s_a$  for birds aged 1 to  $\alpha$ , then the appropriate formula should revert to:

$$s_a = \left( \frac{p}{S_j} \right)^{1/(\exp[3.22+0.24 \times \ln(M)]-\alpha)} \quad (\text{A6})$$

where  $s_j$  is (average) annual survival for birds aged 1 to  $\alpha$  (note that this contrasts to the original Slade et al.'s formula in which  $s_j$  is annual survival from fledging to  $\alpha$ ). Unfortunately, this approach requires an estimate of  $s_j$ , which is typically very difficult to attain, even for first year birds.

By default, Equation A5 is implemented in the current version of the `popharvest` package (i.e. functions `mass.fixed`, `mass.lognorm`). Practically,  $s_j$  is ignored in calculations by setting  $s_j = 1$  in Equation A6. This means that for species with delayed sexual maturity and when  $s_j < s_a$  for birds aged 1 to  $\alpha$ ,  $s_a$  approximated from Equation A5 is negatively biased. However, if estimates of  $s_j$  are available, users can call corresponding functions to use Equation A6 instead. This is done by passing the (fixed) argument `surv.j = average annual survival for birds aged 1 to  $\alpha$` , which is used to estimate  $p = p/s_j$ .

Using the default parametrization [Equation A5], the following example estimates survival to 0.89 and  $\lambda_{\max}$  to  $\approx 1.21$ .

```
PEG(full.option = TRUE, pop.fixed = 10000, NSp = 1,
Fs = 0.3, mass.fixed = 2.5, type.p = "determinist",
type.e = "determinist", alpha.fixed = 2, living.
rate = "long", harvest.fixed = 1500).
```

By passing the argument `surv.j = 0.66`, the function returns 0.90 and 1.199, respectively:

```
PEG(full.option = TRUE, pop.fixed = 10000, NSp = 1,
Fs = 0.3, mass.fixed = 2.5, surv.j = 0.66, type.p = "de-
terminist", type.e = "determinist", alpha.fixed = 2,
living.rate="long", harvest.fixed = 1500).
```